

**Contaminant Assessment and Reduction Project
Water**

CARP

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New York State Department of Environmental Conservation

August 2003

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The Hudson River flows from the eastern Adirondacks through power dams, and into the long lower estuary before reaching New York Harbor and the sea. As it moves, it passes through different social and economic settings where it picks up and transports residuals of those settings. By the time it reaches New York Harbor the river is carrying traces from all the places it has passed. This is an attempt to quantify those traces and to gain an understanding of how the deleterious ones might be reduced.

DEVELOPMENT OF CARP

In the spring of 1996, a few New York State Department of Environmental Conservation (NYSDEC) staffers took the train down from Albany to the Hudson River Foundation on W 20th St. in Manhattan to sit in on a meeting that Dennis Suszkowski, HRF's science director, had organized. The Hudson River Foundation was created to accept money levied from a court settlement against paper mills and power generators using Hudson River water. HRF distributes the money as grants for research on the Hudson. It also hosts meetings, arranges seminars, and generally serves as a meeting site for academics, consultants, people from the Environmental Protection Agency, the Army Corps of Engineers, the Port Authority of New York/New Jersey, the New York City Department of Environmental Protection, the New Jersey Harbor Discharges Group, citizens groups like the Bay Keeper and Friends of Jamaica Bay, and people from the New Jersey and New York state conservation departments.

HRF had awarded a grant to Robert Thomann and Kevin Farley, both professors at Manhattan College in the Bronx. In the 1970s, Thomann had developed a mathematical model for PCBs in the Hudson. HRF's grant was to update the model and to include more chemicals, particularly dioxins, pesticides, and a class of substances produced by combustion called polynuclear aromatic hydrocarbons, or PAHs for short.

Mathematical modeling is a scientific rather than simply descriptive approach to environmental studies. It begins with a general framework of boxes with lines between them. Each box represents a compartment such as "fish" or "water" or "sediment" and the lines are rates of movement of a substance between the boxes. The model looks at the study area as a grid and calculates the rates of movement of chemicals, water, suspended solids, and so forth, both within and between the grid segments over some length of time. Models organize information and help understand what is important and what isn't. If the modelers really understand the relevant rates and have information about the concentrations of the chemicals in the various compartments and at the systems boundaries, they can predict the consequences of changing inputs or chemical loading. A loading is a rate at which some chemical enters the system. Regulatory programs or changes in the uses of chemical should be capable of reducing these loadings and the model could predict how long it would take before reduced concentrations are seen in fish or the water.

After two meetings at HRF a sampling strategy was sketched out calling for sediment and biota (birds, fish, shellfish, benthos, zooplankton), and water sampling. The water component called for samples from sewage treatment plant discharges, sewage treatment

plant sludges, combined sewer overflows, storm sewer overflows, landfill leachates, industrial effluents, and tributaries. Loads (mass per time such as kilograms of a chemical/year) would then be calculated. In order to tie the loads together, the strategy also requested ambient seasonal sampling at 19 sites throughout the area including the Hudson, Passaic, and Hackensack Rivers, the Arthur Kill, Raritan Bay, Upper and Lower Bays, Jamaica Bay, East River, Long Island Sound, and New York Bight.

The logic of this modeling approach is sound but the execution is extremely difficult. Tidal systems like New York Harbor are physically complicated. There was almost no information on chemical concentrations in the water and the extensive data from landfills, sewage treatment plants, and tributaries were usually inadequate or incomplete.

Harbor Estuary Program (HEP)

Much of the focus for the attention on water quality in New York Harbor comes from the 1996 Comprehensive Conservation and Management Plan (CCMP). This document is a product of the Harbor Estuary Program, itself stemming from 1987 amendments to the Clean Water Act. The CCMP is, as implied, comprehensive, and deals with floating debris, pathogenic bacteria, nutrients, habitat, storm discharges, dredging, and toxic chemicals. Toxic impacts are noted in sediments of the harbor and some areas in the Bight and in ambient harbor waters to sensitive organisms in laboratory tests. Reproductive impairments to fish-eating birds have been attributed to DDT. Some birds nesting in the Kills may have suffered from decreased reproductive success and some fish have exhibited fin rot (winter flounder) and liver tumors (tomcod), developmental abnormalities, behavioral impairments, and altered life histories (mummichogs) attributable to chemical pollution.

Body burdens of some chemicals exceed levels believed safe for human consumption. The CCMP identifies 15 chemicals (“chemicals of concern”) as either exceeding enforceable standards (mercury, PCBs, dioxin, PAHs, chlordane), exceeding unenforceable criteria (arsenic, cadmium, DDT and metabolites, dieldrin, heptachlor, heptachlor epoxide, hexachlorobenzene, and gamma-hexachlorocyclohexane (?-HCH), or predicted by modeling to exceed enforceable criteria (copper).

HEP calls for 13 specific actions to reduce continuing inputs of toxic chemicals to the harbor. These are:

- 1) Reduce municipal discharges of chemicals of concern.
- 2) Reduce industrial discharges of chemicals of concern.
- 3) Minimize the discharge of toxic chemicals from CSOs, storm water, and non-point sources.
- 4) Reduce air emissions of chemicals of concern.
- 5) Remediate identified solid and hazardous waste sites.
- 6) Track-down and clean-up of other sources of chemicals of concern.
- 7) Improve chemical/oil spill response and prevention.
- 8) Focus pollution prevention activities on chemicals of concern.

- 9) Identify and remediate selected contaminated sediments.
- 10) Establish consistent methodology to assess risks and improve communication of fish advisories.
- 11) Review and develop criteria for copper and other priority chemicals.
- 12) Assess ambient levels, loadings, and effects of chemicals.
- 13) Develop mass balances for metals and organic chemicals.

Mud Dump Site/HARS/Dredging

Besides HEP's CCMP, other actions directly related to navigational dredging deal with toxic chemicals. The Port of NY/NJ is the largest on the eastern seaboard. Parts of it are naturally very shallow necessitating navigational dredging. Historically, dredge spoils were deposited in the bay and later just beyond the Rockaway/Sandy Hook line. Ocean dumping of dredge spoil is regulated by the 1972 Marine Protection, Research and Sanctuaries Act (MPRSA). In 1982, USEPA Region 2 designated a Mud Dump Site six miles east of Sandy Hook, NJ and eleven miles south of Rockaway, New York. By 1984, the New York District of the USACOE and USEPA Region 2 published a regional guidance manual to implement the national manual (revised also in 1984) in New York Harbor. The local guidance established three categories of dredge spoil:

Category I which is suitable for unrestricted ocean disposal,
Category II sediments may be ocean disposed if capped with Category I material, and

Category II materials have total DDT (sum of DDT, DDE, and DDD) greater than 40 ppb, cadmium greater than 0.3 ppm, or mercury greater than 0.2 ppm in either clams or worms. PCBs are greater than 100 ppb in clams and, as of September 2000, greater than 113 ppb in worms (formerly 400 ppb in worms). Also, if 2,3,7,8-TCDD is greater than 1 ppt and less than 10 ppt or if total **TEQ** (minus 2,3,7,8-TCDD) exceeds 4.5. And finally, Category II is not toxic to clams or worms.

Category III sediments are not suitable for ocean disposal.

Category III material is toxic to laboratory organisms or has dioxin TEQ exceeding 10 ppt.

Under the original 1984 protocols, 95% of the dredged material was Category I and a little less than 5% was Category II. Thus, more than 99% of the harbor dredge spoil could be ocean dumped at a cost of \$5-\$10 per cubic yard. However, growing public pressure for a clean environment forced the federal agencies in 1992 to reevaluate the criteria. The revised criteria resulted in 66% of the dredge spoil being classified Category III (not suitable for ocean disposal) and 9% became Category II (suitable only if promptly covered by Category I). This change in categorization greatly increases dredging costs perhaps to the point of threatening the continued economic viability of the port.

Furthermore, continued ocean disposal of Category II material was halted by executive order in 1996 and in 2000 the criteria for categorization were yet again revisited and made more stringent.

New York/New Jersey Harbor is estimated to have 124,000 directly related jobs with a combined payroll of \$16.5 billion.

With this economic background, the need to get a better understanding of toxics in the harbor became apparent to the governors of both New York and New Jersey and to the Port Authority of New York/New Jersey. The Port Authority articulated a coherent vision of an alliance between the states, the relevant federal agencies, major dischargers, citizen environmental groups, and the Port Authority. The Port Authority also brought \$130 million to the table. The Army Corps of Engineers offered to fund a data management contractor. Thus was born CARP.

Contaminant Assessment and Reduction Project (CARP)

CARP is a cooperative effort of the States of New York and New Jersey, with assistance from EPA and the Army Corps of Engineers, as well as academic and private scientists and engineers, to understand and to reduce contaminants in the harbor (www.carpweb.org).

The principal issues requiring address are:

- 1) To what extent is chemical contamination of harbor sediments and biota historical versus ongoing?
- 2) If a significant portion of harbor chemical contamination is ongoing, what can be done to reduce that load?
- 3) How long will it take before harbor sediments and biota attain certain qualities following cessation or diminution of new inputs?

The target chemicals, to be discussed in greater detail below, are PCBs, dioxins, DDTs and chlordane, mercury, and cadmium. A major impediment to open ocean disposal of dredge material is toxicity. While the above listed chemicals are toxic, they are of interest for their bioaccumulation and carcinogenicity. They are not expected to be at concentrations responsible for the toxicity seen in short exposure laboratory tests. Some existing data points to polynuclear aromatic hydrocarbons (PAHs) as potential short-term toxicants to harbor test organisms.

The 1996 strategy set the sampling sites and chemical list. The other parts of the project were settling the chemical analytical methods, the logistics of getting people and supplies to where they need to go, deciding what to do in the field, and setting up a data management process.

Some of the People Who Made CARP Possible

The NYSDEC workplans were extensively discussed during many meetings held at the Hudson River Foundation in 1997 and 1998. Active participants in the process were Dennis Suszkowski (HRF), Tom Wakeman (Port Authority, NY/NJ), Seth Ausable (the EPA Habor Estuary Plan coordinator), Bruce Brownawell (SUNY Stony Brook), Carleton Hunt (Battelle Ocean Sciences), Richard Bopp (RPI), Dominic DiToro (Hydroqual), John St. John (Hydroqual), Mick DeGraeve (Great Lakes Environmental Center), George Korfiatias (Stevens Institute), Mike Bruno (Stevens Institute), Phil Heckler (NYCDEP), Alan Stubin (NYCDEP), Fred Grassle (Rutgars University), Steve Eisenreich (Rutgars University), Greg Durell (Battelle Ocean Sciences), Eric Evenson (USGS, New Jersey), and Pat Phillips (USGS, New York).

The project leaders from the NYSDEC were Paul Gallay (Special Assistant to the Commissioner) and Jeff Sama (Director, Division of Regulatory Affairs). The supervisor of the NYSDEC sampling operation was Italo Carcich, Chief of the Bureau of Watershed Assessment and Research. Sharon Hotaling helped with the proof-reading.

The labor involved in just accomplishing the water part of the program was substantial. Most of the ambient samples were taken with boats large enough to have internal labs and AC power. Steve Cluett of SUNY Stony Brook helped with the 50 foot *Onrust* based in Port Jefferson on Long Island. Through the assistance of Dore LaPosta and Doug Pabst we were able to get sea time on EPA's *Anderson*. We also spent a considerable amount of time on EPA's smaller 55 foot *Cleanwaters* with the assistance of Randy Braun and Steve Hale. The City of New York generously provided the Marine Science's harbor vessel *Osprey* and field assistance from Jordan Adelson and Mike Cacioppo under the direction of Alan Stubin and Beau Ranheim.

Tributary sampling was performed by Pat Phillips and Gary Wall of the USGS out of Rensselaer, New York.

Wastewater treatment plants in New York City were sampled with the help of Max Obra and his crew at Wards Island. Additional help in ambient and WPCF sampling came from the NYSDEC Region 2 staff, particularly Annetta Vitale, Selvin Southwell, and particularly the Dredge Team members George Hyde and Dare Adelugba. Jimmy Pyn graciously put up with a great many unannounced visits to the Newtown Creek WPCF.

Landfill sampling in New York was done with help from Susan Pepitone of NYCDEP (Pelham Bay), and Ted Nabavi of NYC DOS (Fresh Kills). NYSDEC's Dan Walsh provided much of the basic concepts of landfill sampling. In New Jersey, we were helped by Tom Maturano of the New Jersey Meadowlands Commission.

The design of CSO sampling was worked out by Tom Newman, then of Hydroqual, and accomplished by Iris Martin and Al Torres of Staunton-Chow, a contractor to Hydroqual. SWO sampling was done by a NYCDEP team under the direction of Jerry Volgende and Carol Neptune.

Lily Lee and Ronald Lochan, in addition to Gerry Volgende's staff designed and assisted in the trackdown work in New York City.

The chemists at the labs, particularly Brian Fowler, Coreen Hamilton, Georgina Brooks, Dale Hoover, and Laurie Phillips at Axys Analytical, provided an enormous amount of advice, insight, quantitative data, and encouragement.

Much of the logistics, that is keeping track of supplies and equipment, was handled by Mike Dauphinais here at NYSDEC. NYSDEC rented space at the USACOE facility at Caven Point in Jersey City, NJ to keep sampling equipment for the ambient sampling cruises. In this endeavor we were assisted by Alan Dorfman at the USACOE and by our attorney, Jennifer Hairie.

Larry Bailey and his staff, particularly Gail Dieter and Sue Barbuto, provided a great deal of assistance with lab contracts and in assisting with the interpretation of lab QC procedures.

Particular thanks go to John Donlon at NYSDEC. John built and maintained the TOPS, participated in a great many sampling surveys, assembled equipment and supplies, shipped samples to the labs, entered data, created all the maps, and organized the massive amount of paper and other reporting media returned by the labs.

Over the life of the project people have changed jobs and I'm sure there are many others whose important contributions I've left out. I thank them all.

NEW YORK HARBOR

New York Harbor lies at the bottom of Hudson, Passaic, and Hackensack Rivers. The surrounding counties have a combined population of 13.5 million. New York City and the adjacent communities in New Jersey and New York operate 31 sewage treatment plants having a total design capacity of 2,400 million gallons per day (mgd). These plants generate about 15,000 tons of dry sewage sludge a month. One hundred and 85 square kilometers (20% of the surface area of NYC) is composed of landfill. Almost all of the landfills were built on wetlands adjacent to the estuary. Rainwater percolating through these sites results in 41 mgd of Coca Cola® colored leachate. New York City's 394 combined sewer outfalls are estimated to discharge 135 mgd of untreated sewage.

Major Tributaries

Yearly average discharges of five major rivers (at head of tide) are:

cubic feet per second, CFS

Tributary Rivers:

| | |
|---|-------|
| Hudson at Waterford above confluence with Mohawk R. | 5,300 |
| Mohawk River at Cohoes | 3,700 |
| Wallkill at Gardiner, New York | 690 |
| Passaic River at Little Falls, NJ | 740 |
| Raritan River and Bound Brook, NJ | 770 |

There is an enormous degree of daily and even hourly variability in these river discharges. The estuary responds to freshwater flows, to tides, and to other forces such as wind velocity and barometric pressure.

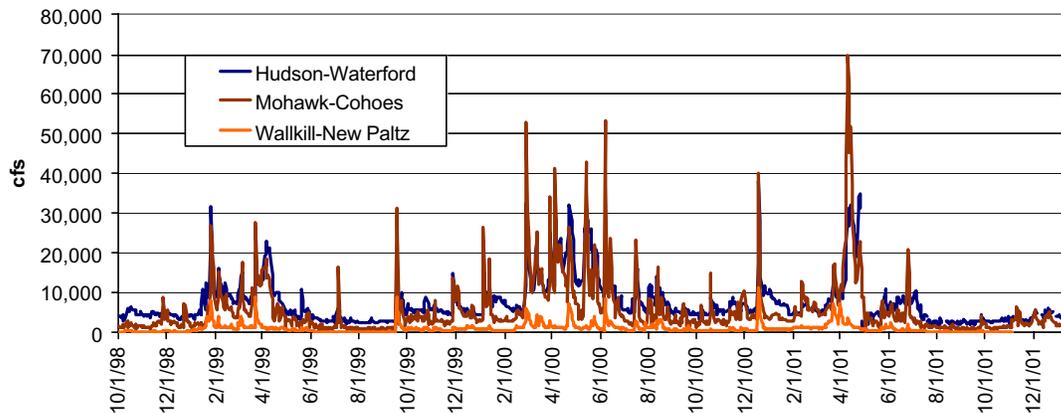


Figure 1. Comparison of discharges of the three largest freshwater sources to the Hudson Estuary over the life of CARP.

One can safely assume that a parcel of water entering the estuary will eventually reach the sea; it's much harder to know when. It is very much more difficult to predict when a

particle entering the estuary will reach the sea or the harbor. It may spend a significant amount of time buried in a deposition area and in regions nearer the ocean large amounts of sediments are brought into the harbor from the continental shelf.

USGS researchers working within the CARP studied the loading of particles particularly during high flow events. Generally, about half a river's discharge occurs in about 10% of the year in a small number of events. The loading of solids and more particularly, organic carbon may occur over shorter spans. A separately funded investigation by Gary Wall of the USGS and Rocky Geyer of Woods Hole, instigated by CARP observations, is investigating some of the fine-grain detail of particle movement in the mid-Hudson near Poughkeepsie, New York. This work will yield insights that will assist in predicting the movement of particles through the tidal portion of the Hudson.

Minor Tributaries

The CARP design called for investigating the contributions of some of the minor tributaries. The extent of urbanization in New York City is such that there are few surface streams. CARP selected three; the Bronx River in the Bronx, Saw Mill River in Westchester, and the Gowanus Canal in Brooklyn. The Gowanus Canal and Saw Mill River are obviously affected by street run-off, CSOs, and probably contaminated groundwater. The Bronx River is, particularly as it makes its way through the New York Botanical Garden, surprisingly lovely but still chemically affected by its passage through Westchester County and the Bronx.

While neither the Saw Mill River nor the Bronx River are gauged, discharges may be estimated by assuming proportionality to that of nearby streams that are monitored.

The Gowanus Canal is an odd choice as a tributary. It was at one time a real tidal creek but is no longer. It had been canalized and served as part of the terminus of the Erie Canal. Raw sewage had rendered the Gowanus obnoxious and, in 1911, the City of New York installed a pumping system to flush it out with East River or Upper Bay water. The flushing system was rebuilt in 1999 and uses East River water. The present system has an average pump rate of 200 mgd and a maximum rate of 300 mgd. The pump rate varies with the tidal elevation so as not to mobilize contaminated bottom sediments. In this case, discharge is East River water that may have some entrained bottom sediments.

There is a concern that the flushing activities would mobilize contaminated sediments and we decided to sample the Gowanus as a potential contaminant source via transport of resuspended solids.

Sewage Treatment Plants

There are 14 sewage treatment plants in New York City alone. They are on Staten Island (Port Richmond in the north and Oakwood Beach in the south), two in Manhattan (North River in the northwest and Wards Island in the northeast – actually in the East River), one in the Bronx (Hunts Point), three in Queens (Bowery Bay in Steinway near Rikers Island,

Tallman Island in College Point near the Whitestone Bridge, and Jamaica near Kennedy Airport), and six in Brooklyn (Newtown Creek in Greenpoint, Red Hook next to the Brooklyn Navy Yard, Owls Head in Bay Ridge, Coney Island in Sheepshead Bay, 26th Ward in Spring Creek, and Rockaway).

Besides the 14 NYC plants, the CARP also sampled plants in Rensselaer (near Albany), Poughkeepsie, Rockland, and Yonkers. NYSDEC sampled two plants in New Jersey, Passaic Valley Sewerage Commissioners (PVSC) and Edgewater.

Table 1. Harbor area WPCFs.

| WPCFs | MGD |
|--------------------------------|-----|
| Newtown Creek (NC) | 286 |
| <i>Passaic Valley (PVSC)</i> | 283 |
| Wards Island (WI) | 258 |
| North River (NR) | 161 |
| Hunts Point (HP) | 148 |
| Bowery Bay (BB) | 126 |
| Owls Head (OH) | 124 |
| Coney Island (CI) | 115 |
| Yonkers (Westchester Co.) (YO) | 92 |
| Jamaica (JA) | 81 |
| 26th Ward (26) | 68 |
| Tallman Island (TI) | 55 |
| Red Hook (RH) | 41 |
| Port Richmond (PR) | 35 |
| Oakwood Beach (OB) | 27 |
| Rockaway (RO) | 27 |
| Rockland County (RK) | 26 |
| Rensselaer (RE) | 24 |
| Poughkeepsie (City) (PO) | 14 |
| Orangetown SD2, | 13 |
| Tri-City | 12 |
| Newburgh | 9 |
| Haverstraw | 8 |
| Kingston | 7 |
| Beacon | 6 |
| Poughkeepsie (Town) | 4 |
| Wallkill (Town) | 4 |
| <i>Edgewater</i> | 3 |
| Ulster (Town) | 1.6 |
| Yorktown Heights | 1.5 |

Each of the targeted WPCFs was sampled at least three times with the exception of Red Hook where there were field problems and one sample set was rejected. Some plants

(Newtown Creek, Port Richmond, Hunts Point, and 26th Ward) were sampled more intensively. Two facilities in New Jersey (PVSC and Edgewater) were visited once each as part of an investigation in sampling technique.

Sewage treatment plants generally do four things. They remove stuff that sinks (“grit”) and stuff that floats (“scum”). They grow and harvest bacteria that degrade and metabolize dissolved organic material. And they disinfect the final water with chlorine. Solids, called sludge or biosolids, are dewatered, palletized, and used as fertilizer. Not all the WPCFs in New York City have dewatering facilities necessitating the transport of watery sludges by ship to plants with drying capabilities from those without. The eight plants that have dewatering capabilities and their estimated monthly output of sludge are listed below:

Table 2. Biosolids production at NYSDEP WPCFs.

| WPCF | Tons/month, dry weight |
|-----------------------|------------------------|
| Wards Island | 3000 |
| Hunts Point | 2500 |
| 26 th Ward | 1200 |
| Bowery Bay | 1000 |
| Oakwood Beach | 820 |
| Jamaica | 690 |
| Tallman Island | 450 |
| Red Hook | 230 |

Municipal treatment plants are not designed to remove toxic chemicals and on occasion, toxic chemicals discharged to sewers may disrupt treatment plants by harming the bacteria and protozoa essential to the process. Therefore, sewage treatment plant operators run programs, called “pretreatment”, to regulate what is discharged into sewers by manufacturers or certain commercial establishments. The following table shows the kinds and numbers of facilities in the New York City pretreatment program by wastewater treatment catchment area¹.

¹ Data kindly supplied by Leslie Lipton, NYCDEP, 7/17/2003.

Table 3. Type and number of NYC industries discharging to city WPCFs. The key to the abbreviations is in Table 1.

| | Totals | 177 | 110 | 68 | 28 | 26 | 23 | 19 | 13 | 10 | 9 | 8 | 6 | 1 | 1 |
|-----------------------------------|--------|-----|-----|----|----|----|----|----|----|----|----|----|----|----|----|
| | | NR | NC | BB | OH | WI | HP | JA | 26 | TI | RH | CI | PR | OB | RK |
| new source metal finishing | 248 | 141 | 53 | 32 | 6 | 1 | 3 | 4 | 5 | 2 | 1 | | | | |
| radiator shop | 31 | | 2 | 3 | 4 | 6 | 6 | 4 | 2 | 1 | | 2 | 1 | | |
| industrial launderer | 26 | | 9 | 3 | 2 | 3 | 1 | 3 | 1 | | 1 | 1 | 1 | 1 | |
| metal finishing/non-cat | 24 | 6 | 8 | 4 | 2 | 1 | | 2 | 1 | | | | | | |
| miscellaneous | 23 | 3 | 5 | 4 | | | 3 | 3 | 1 | 1 | 2 | 1 | | | |
| paint/ink formulator | 21 | 2 | 4 | 2 | 3 | 3 | 3 | | 1 | | 1 | 2 | | | |
| metal finishing | 17 | 10 | 1 | 1 | 1 | 1 | | | | 1 | 1 | | 1 | | |
| textile dyer | 14 | 2 | 6 | 1 | | 1 | 1 | | | 2 | | 1 | | | |
| soap & other detergents | 9 | | 2 | 1 | 3 | 1 | | 1 | 1 | | | | | | |
| steam electric generation | 9 | | 1 | 2 | 1 | 2 | | | | | 1 | | 1 | | 1 |
| centralized waste treatment | 8 | 1 | 2 | 1 | 1 | 1 | | | | 1 | 1 | | | | |
| electroplating-> 10k gpd | 7 | | 1 | 5 | | | | | | 1 | | | | | |
| electroplating- < 10k gpd | 7 | | 3 | 1 | 1 | | 1 | | | | | 1 | | | |
| organic chemicals/non-cat | 7 | | | 3 | 1 | 1 | 1 | | | 1 | | | | | |
| pharmaceutical manfg. | 6 | | 1 | 1 | | 3 | | 1 | | | | | | | |
| metals molding & casting | 4 | | 1 | 1 | 1 | | | | | | 1 | | | | |
| nonferrous metals form & powders. | 4 | 1 | | | 1 | 1 | | 1 | | | | | | | |
| organic chemical/categorical | 4 | | 1 | | | | 2 | | | | | | 1 | | |
| photoengraver | 4 | 3 | 1 | | | | | | | | | | | | |
| steel drum reconditioner | 4 | | 4 | | | | | | | | | | | | |
| fur dresser & dyer | 3 | 1 | 2 | | | | | | | | | | | | |
| pesticide chemicals | 3 | | | 1 | | 1 | 1 | | | | | | | | |
| heat treater | 2 | | 1 | | | | 1 | | | | | | | | |
| metals molding & casting/non-cat | 2 | 1 | | | | | | | 1 | | | | | | |
| nonferrous metals manfg. | 2 | 2 | | | | | | | | | | | | | |
| NS metal molding & casting | 2 | 1 | | 1 | | | | | | | | | | | |
| photofinishing | 2 | 2 | | | | | | | | | | | | | |
| copper forming | 1 | | | | 1 | | | | | | | | | | |
| inorganic chemicals/non-cat | 1 | | 1 | | | | | | | | | | | | |
| instruments & related products | 1 | | 1 | | | | | | | | | | | | |
| new source metal finishing | 1 | 1 | | | | | | | | | | | | | |
| NS metal molding & casting/MF | 1 | | | 1 | | | | | | | | | | | |
| pulp & paper products | 1 | | | | | | | | | | | | 1 | | |

CSOs and SWOs

Sewage gets to sewage treatment plants by means of sewers. Sewers and sewage treatment plants have designed capacities. Too much water cannot be properly treated and may, if unchecked, harm the process by washing out the bacteria being farmed at the sewage treatment plants. The collection system can divert excess water and the sewage treatment plant operators also watch their intake and can divert excess water. In some cases, at 26th Ward in Brooklyn for example, diverted water is held in vast underground tanks and processed at the treatment plant during dry weather but in most instances,

diverted water goes directly into surface water. When the diverted sewage is a mixture of rain water and what is politely called “sanitary” waste, its called combined sewer overflow (CSO); when it’s only run off from the streets or roofs, it’s called storm water overflow (SWO). In newer areas, cities build separate collection systems for storm water and sanitary water. In a separated system, the discharge of excess water during a storm is less contaminated and the treatment plants are less exposed to excessive flows. Most of New York City has combined sewers but there are some separate sewers in parts of Queens and in southern Staten Island. We thought that the CSOs might turn out to be important sources for loading of chemicals into the harbor.

The capacities, number of CSOs, and estimated average CSO discharges are shown below for 13 New York City WPCFs (Oakwood Beach omitted).

Table 4. NYC CSOs. See Table 1 for the key to the abbreviations.

| WPCF | Max Cap. (mgd) | # CSO | Est. Avg. CSO Disch. (mgd) |
|------|----------------|-------|----------------------------|
| NC | 602 | 66 | 14 |
| WI | 457 | 76 | 11 |
| BB | 300 | 47 | 13 |
| NR | 298 | 46 | 5 |
| HP | 272 | 30 | 15 |
| OH | 234 | 8 | 9 |
| CI | 181 | 3 | 10 |
| JA | 160 | 14 | 31 |
| 26 | 125 | 4 | 12 |
| TI | 114 | 16 | 10 |
| PR | 109 | 33 | 1.0 |
| RH | 92 | 29 | 3.7 |
| RO | 37 | 22 | 0.47 |

CSOs (combined sewer overflows) should occur only when the amount of water entering a treatment plant exceeds twice its design capacity. When influent flows are greater than design capacity but below twice design capacity, some primary treatment (removal of solids) is possible but higher flow rates may damage the facility. Treatment plant operators throttle down the intakes and the water exits the system through diversions. The system’s design permits some overflow to escape under minimal rains so that CSOs actually occur during times when twice design capacity has not been reached. Modeling studies estimate that the city’s 394 CSOs release about 140 mgd of untreated wastewater.

During CARP, CSOs were assessed indirectly by taking samples at the influent to wastewater treatment plants during wet weather. Sampling persisted over the time that influent flows exceeded the plant’s design capacity. Water at inlets is a mixture of the entire system and it avoids the parochialism of a particular sampling point. Furthermore, access to the wastewater treatment plant is simpler than to actual CSOs. Simpler is not necessarily simple. Wastewater treatment plants recycle water from various operations back into the influent often at a place upstream of the most convenient sampling point.

We do not want to sample these mixed streams. Also, some plants receive water in more than a single trunk. These were sampled separately. The amount of water required (about 100 L) and the length of time over which samples were collected (4 hours) limited the field crew to a single sample per storm event.

Landfills

About 20% of the surface area of New York City is landfill. Much of the 46,000 acres of landfill were created to hold ash generated in heating and cooking. Only a small proportion, some 2000 acres, are modern landfill. These acres are in the Bronx (Pelham Bay, 100 acres), Brooklyn (Pennsylvania Ave., 100 acres; Fountain Ave., 300 acres; Edgemere, 120 acres), and Staten Island (Fresh Kill, 1200 acres, and Brookfield, 180 acres). Assuming a yearly rainfall of 1.1 meters and infiltration rate (proportion of rainfall that becomes part of the groundwater) ², the estimated leachate production is 2.6 mgd. Furthermore, an appreciable amount, perhaps 1 mgd of this total, is treated, at Fresh Kills Leachate Treatment Plant for the Fresh Kills and at Hunts Point WPCF for Pelham Bay Landfill. Experience at the Fresh Kill Landfill Treatment Plant suggests that this estimate may be high but the discharge may suffice for modeling.³

Some of the leachate from the New Jersey Meadowlands Commission sites is treated at the Passaic Valley Sewerage Commissioners (PVSC) WPCF in Newark, NJ but some leachate flows directly into the Passaic River. Assuming a similar size of landfills in New Jersey, a rough estimate of 4.2 mgd of untreated leachate may be entering the harbor area surface waters from both states.

Leachate is colored, high pH, and strong smelling. It is the product of the breakdown of mounded garbage flushed out by ground water or rainwater. Toxic chemicals, particularly metals, occur in leachates. Illegally dumped waste oils and other substances also have gotten into either the wastestream or have been directly placed in landfills.

Leachate was taken from two New York City facilities, Pelham Bay in the Bronx and Fresh Kills in Staten Island. Pelham Bay has a leachate collection system that delivers leachate to the Hunts Point WPCF. Our samples came from holding tanks at Pelham Bay. The Fresh Kills site consists of numbered mounds. Leachate from the mounds is gathered by a system of trenches and pumps. Most of the leachate production comes from mounds 1 and 9. Mounds 6 and 7 are also important producers. As they enter the Fresh Kills treatment plant they are combined into 1/9 and 6/7. There are several sampling points around the mounds, a few of which have been sampled in CARP. These include 1/9 B and 1/9 F. Three mounds, 1A, 1D, and 1E, were sampled in New Jersey at the New Jersey Meadowlands Commission (formerly called the Hackensack Meadowlands Development Commission).

² Walsh, D.C., 1996. Geochemistry of solid waste landfills. Ph.D. Thesis submitted to Rensselaer Polytechnic Institute, Troy, NY.

³ Personal conversation, Philip Gleason, NYS Department of Sanitation, June 1, 2001.

SAMPLING HISTORY AND METHODS

Despite an enormous effort to monitor trace contaminants, Bob Thomann and Kevin Farley had virtually no data for their model. Why? In most places the concentrations of these chemicals are very low relative to the capabilities of conventional analytical methods. They are not, however, low relative to the risk-based water quality standards required for protection of human health. For example, the canonical technique for measuring PCBs (US EPA Method 608 – based on gas chromatography, electron capture detection, and pattern recognition of PCB congeners or domains characteristic of Monsanto's Aroclor mixtures) in wastewater has a method detection limit of 65 parts per trillion (ppt) but a practical detection level, taking variability and interferences into account, of often more than 300 ppt. The old New York ambient water quality standard for PCB (to protect humans eating wildlife) was 1 ppt. A three hundred-fold difference is a little larger than the difference between the speed of the space shuttle (17,500 mph) and New York's highway speed limit (65 mph). This situation is somewhat analogous to equipping speed cops with radar guns incapable of telling whether the space shuttle is moving or staying still. The current New York State (NYS) water quality standard for PCBs is three orders of magnitude lower (0.001 ppt or 1 part per quadrillion).

Persistent bioaccumulative chemicals like PCBs occur in all surface waters and in all wastewaters but the methods most often used are incapable of measuring toxicologically significant concentrations. Even when they are detected, sampling and laboratory errors introduce so much variability that the value of the data to modeling becomes suspect. This problem was compounded by the magnitude of the project. Sampling would be performed by many teams in two states using six different labs for organic chemical analysis and two for metals. The field methods themselves were novel and under development. The laboratory methods were also far from routine.

TOPS

One of the fundamental goals of the CARP is consistent detection of all target chemicals from all media. This has been a new idea. As noted before, the field and lab methods commonly used in regulatory programs are often incapable of detecting PCBs in surface or waste waters and always incapable of measuring dioxins and furans in water. This is caused by insufficient mass of analyte, interferences, and sometimes in the case of PCBs, unexpected patterns of congeners.

Taking these issues in reverse order, the occurrence of non-Aroclor congeners will not be seen by pattern-recognition where Aroclors are expected. The question of interferences becomes large in places where a lot is going on, like New York Harbor. Consider the task of weighing yourself. Step on a bathroom scale in the privacy of your bathroom and the reading you get from the instrument will only reflect your own weight. But if you set the scale up on a crowded sidewalk other people might be stepping on the scale at the same time you are trying to weigh yourself. Much of what the analytical chemist does with the sample is to reduce these interferences and techniques have gotten quite good – but still not perfect. And, investigation into interferences has played an important part in

environmental science. It was Søren Jensen's studies of interferences in DDT analysis that initially revealed the wide-spread existence of PCB in 1963. Later researchers saw unexplained chromatographic peaks while looking for PCBs and discovered the flame retardant and insecticide Mirex. Interfering chemicals may be mistaken for the target chemical or they may be recognized and eliminated but in the clean-up process, target analyte is also lost. More commonly, interferences increase the signal noise and consequently result in degradation of the signal.

The bathroom scale analogy is relevant to the question of accuracy. Consistent data produced by a single system are useful for making decisions even if not highly accurate. However, data collected by multiple systems are useful only if they be related back to a standard. The difficulty in doing this increases with the complexity of the data collection system. There are differences in extraction efficiencies between different environmental matrices (sewage versus landfill leachate versus surface water), between sampling systems (TOPS, PISCES, grab samples), between labs, between sets of field personnel, and so on.

The analyte mass issue can also be analogized by trying to weigh something that's very small, say a feather on our bathroom scale. The scale is too insensitive to register such a light object. However, if we collected enough feathers, we would be able to use the scale to measure them. But then, we'd need to be able to put the resulting mass into the correct units. Perhaps we might count the feathers and divide the total mass by the count to get a mean mass per feather. Similarly, in doing trace organics sampling, processing large water volumes can result in significantly lower detection limits – and we have to be able to measure accurately the volumes of water processed.

When I started working for DEC in 1979 we sampled for PCBs by collecting water into a quart Mason jar. Since the early 1980s researchers in Canada and the United States have been experimenting with ways to field concentrate larger and larger volumes. Early work on the Niagara River, by Peter Goulden and others at Environment Canada's Canada Centre for Inland Waters in Burlington, Ontario, resulted in a series of devices to mix sample water with the solvent dichloromethane chloride (DCM) and then to remove the DCM for analysis. The result of their work was a complicated piece of glassware called the Large Sample Extractor and, after Peter's unexpected death, renamed the Goulden Large Sample Extractor or GLSE. The GLSE is used by Environment Canada on the Niagara River and elsewhere, and the USEPA has operated it on board their Great Lakes Research Vessel *Lake Guardian*. To use the GLSE water is first clarified by filtration or by centrifugation. Particles collected on the filter or from the centrifuge can be sent to a lab for extraction and analysis. Some of the clarified water is fed into the GLSE where its heated and then stirred up in a mixing vessel with DCM. The rest of the GLSE separates the DCM from the water. The water is wasted and the DCM is recirculated back to the mixing vessel. Since DCM dissolves rather well into warm water, a separate pump makes up for the DCM losses. Dissolved PCBs are captured by the bulk phase DCM so the DCM that is lost through dissolving into the water does not carry with it much of the PCB. The retained DCM then contains the dissolved phase of the PCB.

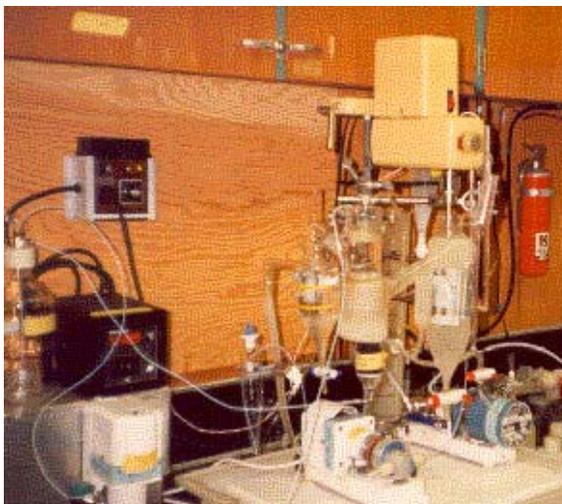


Figure 2. Goulden Large Sample Extractor set up in Environment Canada's monitoring station in Fort, Erie, Ontario.

In 1992 over a 32 hour period, we tested duplicate GLSEs, one set up, with the assistance of Dave DeVault then at EPA, on a bridge over the Oswego River in upstate New York at Minetto and another set up on the *Lake Guardian* moored a few miles away in the City of Oswego. Both processed Oswego River water collected at the same time through the same plumbing.

At the same time we were performing that experiment, Tony Ethier of Seastar, a Canadian oceanographic equipment manufacturer in Sidney, British Columbia, had set up a commercial sampling device called Infiltrax also from the Minetto Bridge. Tony sat in his car reading a book while I was rushing around tending to the GLSE.

The results of the experiment on the Oswego River Bridge showed that minute differences in glassware or operation produced order of magnitude differences between samples processed on the bridge and on the ship. This finding dashed the hope that we could have sufficiently sensitive and comparable field technique deployed in open lake waters from the *Guardian* and also from truck mounted sampling on the tributaries. A few years later, in 1995, the EPA gave NYSDEC a grant to determine if dioxins, PCBs, and chlorinated pesticides occurred in final effluents from sewage treatment plants discharging to Lake Ontario. EPA suggested that we consider using a device that captures a "cubic meter" of water. This device turned out to be the same one that Tony Ethier had deployed on the Minetto Bridge.

Infiltrax is a self-contained submersible unit powered by a stack of lantern batteries. It pumps water through a glass fiber filter and then through a Teflon[®] column holding a synthetic resin called XAD-2. After a set volume, or time, or if the filter is plugged, Infiltrax shuts off. The filter and the XAD are extracted and the extract may be analyzed. While the GLSE processed 50 to 100 L, Infiltrax dealt with 200 to 300 L of water. EPA was suggesting that we process 1000 L. After speaking with a number of Infiltrax users we decided to build our own version using AC power and many of the pumps left over from the ill-fated GLSE experiment. Infiltrax could be used with either a flat Whatman GFF grade glass fiber filter (0.7 micron nominal porosity) or a 4-inch wound glass cartridge filter (1.0 micron nominal porosity). We selected the cartridge filter as it had a

much greater capacity than the flat filter and would make the unit less likely to suffer plugging and then to pre-maturely shut-down. Everything needs a name and one of our engineers, Cynthia Leece, suggested calling it “TOPS” for Trace Organics Platform Sampler.

TOPS was designed with our GLSE experience in mind. It was, unlike GLSE, entirely self-contained so no part of the sample was exposed to the air. There had been problems with GLSE picking up PCBs from the air. TOPS was also intended to be physically rugged in contrast to all the breakable glassware in GLSE. Unlike Infiltrax, TOPS was capable of filtering more water than it passed through the XAD. Most synthetic chemicals of interest are poorly soluble in water and tend to adhere to suspended particles. We had seen in the Niagara many cases where we were detecting substances only from the particulate phase. TOPS was intended to filter very large volumes of water, particularly in low turbidity situations. And finally, TOPS was to be very easy to run in the field. The last wish was achieved initially but has been steadily receding.

Using TOPS in 1996, we were able to determine the presence of the target analytes in all the treatment plant effluents. In 1997, we used TOPS to investigate dioxin and PCB levels in tributaries to Lake Ontario. In the later summer I brought it to Burlington, Ontario where Melanie Nielson and her staff and colleagues at the Canada Centre for Inland Waters criticized the design. We followed up on the many excellent suggestions and were able to re-engineer a better instrument. An improved TOPS permitted measuring open lake PCBs from the *Lake Guardian* in October of 1997 during a four day cruise.

For that work we had to figure out a way to get water from the lake while the ship was in motion. Ships are dirty and clean sampling requires avoiding smoke and ship-generated effluents. The ship’s crew and captain, David Moser, helped rig a 45 pound bomb-shaped device, called “DL-76” (made to be lowered from a bridge into a flowing river to collect a sample of water for measuring suspended sediments) from an A-frame on the starboard rear. On contact with the flowing stream the DL-76 swings around to point into the current. NYSDEC Engineering Technician John Donlon attached a TOPS intake to the top surface of the DL-76. A test run out of the Port of Rochester showed the set up to be stable up to about 5 knots but at higher speeds the tow-fish had a tendency to dolphin, particularly in choppy water.



Figure 3. TOPS on the EPA's New York Harbor Survey Vessel *Clean Waters*. The "U" shaped device near the top holds a cartridge filter and the two white double-ended cartridges hold XAD. The two small boxes in the upper right side are flow meters.



Figure 4. The DL-76 tow fish on board the R/V *Lake Guardian* moored at the mouth of the Genesee River, Rochester, New York.

As CARP was getting going in 1997 and 1998 we were gaining experience in obtaining very large volume samples from ships and at wastewater treatment plants. We knew we could detect all the target analytes. XAD was still new to us and somewhat mysterious. We knew that it was being used in large lake studies in Green Bay, Wisconsin and in the Lake Michigan Mass Balance Study. It had had some 20 years of use in the environmental field and we had seen that it appeared to behave similarly to the GLSE set up on the Niagara River. While initially we didn't have a way to evaluate it, we believed that if we followed the literature and used the same procedure everywhere, we would have a consistent data set.

Over the next few years we followed suggestions from Brian Fowler, at the Axys Group in British Columbia, to add more and more quality control tests. These were:

- 1) to set XAD columns in series and to analyze each individually as a way to measure break-through,
- 2) to spot the columns with chemical surrogates to test for wash-out and recovery,
- 3) to pump water through the XAD at different rates (pump speeds), and finally
- 4) to meter into the water stream chemical surrogates that would mimic trapping dissolved chemicals by the XAD.

The last, at least for the analytes most similar to the surrogates, is a more realistic test of XAD trapping efficiency, but one not without flaws. Naturally occurring macro-

molecules like humic and fulvic acids appear to bind synthetic hydrophobic chemicals and reduce their availability for extraction by hexane. Research by Mark Driscoll at the SUNY Environmental and Forestry College in Syracuse, New York, suggests that the length of time needed for PCBs to reach a binding equilibrium with these dissolved materials is days or even weeks. This effect is not being captured in the metered surrogate delivery used with TOPS. Dr. Driscoll had some success with simultaneous hot chromic acid digestion and extraction for PCBs but this harsh treatment would have destroyed some target pesticides.

The specific details of operating the TOPS during CARP are set out in an attached document, TOPS- Standard Operating Procedure. Dr. Gary Wall of the USGS devised some critical modifications for long duration TOPS operation in streams with very high suspended sediment concentrations. These are discussed in an attached paper, Use of a Large Volume Sampler in River Settings.

Many samples were taken during the course of CARP to help evaluate the efficiency of both XAD and the filter. At the outset of CARP a large sediment sample was taken, thoroughly mixed, subdivided, frozen, and periodically sent in for analysis to help understand inter and intralab variability. Results of these studies are reported and discussed in an appended paper (The Performance of An XAD/ One-Micron Cartridge Filter Trace Organic Platform Sampling System (TOPS)).

PISCES

In the mid-1980's we were looking for a cheap way to capture PCBs. The only really useable tools then available were to collect natural concentrators like sediments or biota. Those media were spotty and inconsistent. In many instances, a sample collection could be very time consuming. John Hassett at the State University of New York College of Environmental Science and Forestry in Syracuse, New York, had been thinking about passive samplers as surrogates for fish. Passive samplers use a membrane separating water from a solvent. PCBs, as well as pesticides and some PAHs (in theory dioxins and furans too) will migrate through a solvent saturated membrane and become trapped in a non-polar medium. The rate of movement is a function of the analyte/solvent interaction, the porosity of the membrane, and temperature. Dr. Hassett was working on a passive sampler that used hexane as the solvent and one mil polyethylene as the membrane. He called it PISCES for **P**assive **I**n-Situ **C**hemical **E**xtraction **S**ampler.

In 1988 we had success in identifying a PCB source to the Black River at Carthage, New York using PISCES and, in 1992, we found a PCB source to the Arthur Kill in New York Harbor with the device. The NYSDEC Fish and Wildlife program uses them as well as the USGS in Massachusetts, the New Jersey Harbor Dischargers Group at Linden Roselle, and a private contractor working on Alaska's North Slope uses PISCES to locate PCBs on the Colville River.

Different teams use slightly different versions of the same idea. Basically, PISCES consists of a container holding about 200 mL of hexane. We use a schedule 80 4-inch

brass nipple sealed at the top with a Teflon® disk and at the bottom with the polyethylene. I place them in rivers, streams, or sewers with the membrane side down and leave them in the water for about two weeks. Derivation of a sampling rate requires knowing the water temperature. By knowing the rate, membrane area, and time of exposure, we can roughly estimate the volume of water sampled through the use of an empirically derived equation that John determined. PISCES are cheap (our model costs about \$20 each) enough to permit placement in risky locations where losses are possible and rugged enough to withstand severe buffeting in storm charged sewers or streams.



Figure 5. Preparing PISCES in the field. PISCES are usually deployed in pairs in case one fails. The bottles in the truck hold hexane.

Data Management

The data produced by CARP are voluminous. Battelle Ocean Sciences of Duxbury, MA, maintains an official data base.

SAMPLES

Sampling for organic chemicals was accomplished using TOPS. The sampling routine was similar everywhere but not identical. For sampling the three minor tributaries and the 20 ambient stations, a 100 mesh Nytex plankton net was used to strain out larger zooplankton. In some cases, the plankton net removed a considerable amount of material. This required deploying a submersible pump and use of a stainless steel can as a receptacle for the strained water.

Landfill samples were taken from leachate collection systems. The purpose of the sampling was to obtain a quality estimate for leachate escaping the various treatment systems. Since the escaped leachate was entering the surface waters through diffusion, there was no way to estimate the particle bound phase of the transport. Therefore, only glass fiber filtered water was processed in CARP. There is no reason to collect POC or TSS data from leachate samples.

Two of the three major tributary monitoring stations (Hudson and Mohawk) were essentially identical. Water from a submerged intake ran into a shed holding a TOPS and ISCO samplers for the metals, whole water PAHs, and DOC/POC, and suspended sediment samples. The set up on the Wallkill was a little different. Very high event related suspended sediment concentrations resulted in the rapid plugging of the glass fiber filters. To overcome this problem, Dr. Gary Wall of the US Geological Survey (USGS) devised a settling tank made from a modified stainless steel milk can. The can was on an electronic scale. As the can filled and as it's weight increased, a datalogger triggered the TOPS to pull partially clarified water from the can. At the end of the event, the heavy material that settled out in the can was scraped out and incorporated with the filter as the particulate sample.

The success of sampling the tributaries can be seen in graphs displaying the hydrograph of the study period and the hydrograph of the period sampled.

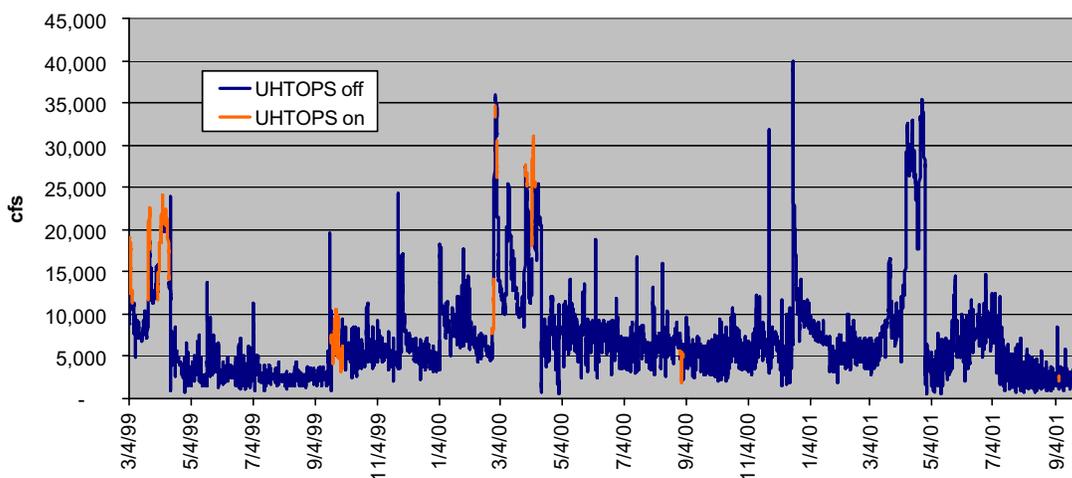


Figure 6. Over the period of study, TOPS was pumping while 7% of the Hudson's flow was passing.

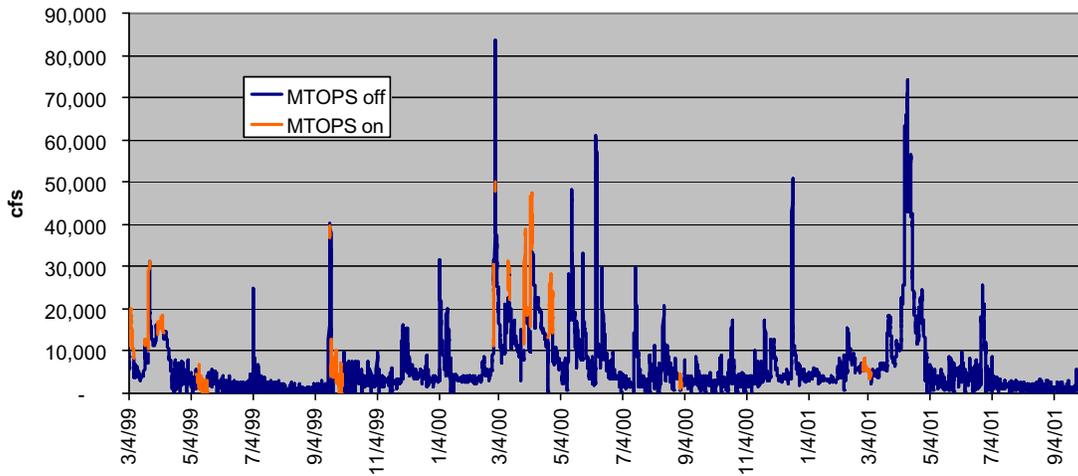


Figure 7. Over the study period, TOPS was pumping while 11% of the Mohawk's flow passed.

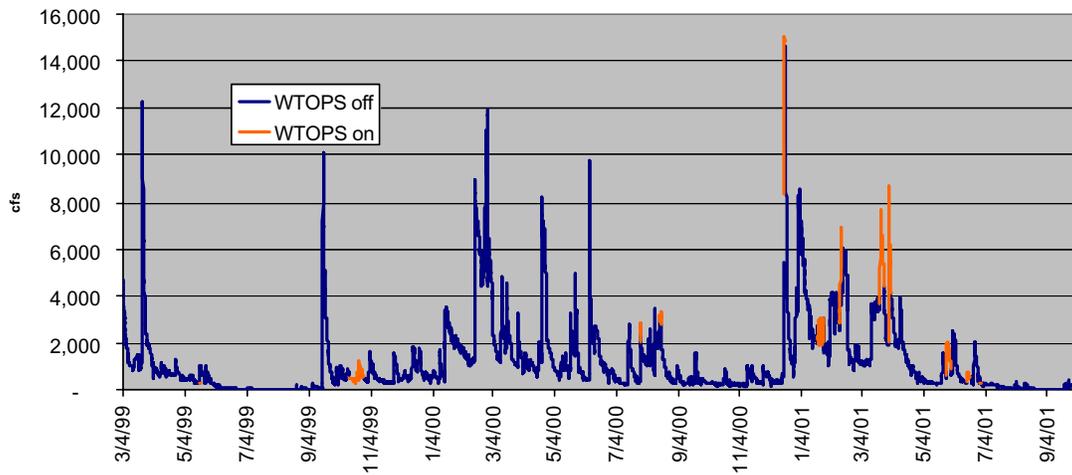


Figure 8. During the period of record, TOPS was pumping while 10% of the Wallkill flow was passing.

Most of the samples taken under CARP were done so through the use of conventionally set up TOPS and allowed the sample volume, the number of liters passed through the filter or the XAD columns, to be adjusted in the field. The average volumes of water (in L) processed using conventional TOPS are shown below:

Table 5. Average volume of water passed through TOPS media, by sample type.

| Sample Type | XAD | GLASS FIBER CARTRIDGE |
|----------------------|-----|-----------------------|
| AMB-clean | 680 | 3100 |
| AMB-Hudson | 210 | 890 |
| AMB-Kills | 170 | 760 |
| AMB-Non Kills | 190 | 720 |
| Industrial effluents | 190 | 1000 |
| Landfill leachate | 90 | |
| Major Tribs | 260 | 850 |
| Minor Tribs | 190 | 740 |
| WPCF | 140 | 380 |



Figure 9. TOPS running on the Passaic River. Notice the stainless steel can on the left where large zooplankton are removed by filtration. A syringe pump delivering metered surrogates is seen on the table to the right of the TOPS. The plastic carboys on the deck collected water during timed intervals as an independent check of pumping rates. Note the lack of simplicity.

We had wanted to use metered surrogates at an early date in the project but experienced a variety of problems executing it. The “gas-tight” glass syringes turned out to be temperature sensitive and the fittings leaked. These problems were eventually solved by the decision to move the process into the lab. The lab-based set-up was called “TOPS-Next Generation”. Beginning in February of 2001, 35 samples were processed using TOPS-Next Generation. This modification was made to permit much slower XAD processing rates (from about 600 mL/min to about 16 mL/min) and it allowed for much better control of the metered surrogates. The disadvantage of the process was that it limited the total sample size to a little less than 100 L and there were more opportunities for sample contamination.

Cosine Tide

The average duration of sampling in tidal ambient sites (in areas other than Long Island Sound and the New York Bight) was 5.6 hours. This is a portion of a tidal cycle and results may be affected by the direction of the tide. To obtain a quantitative value expressing the tides over the duration of sampling we considered the tide to be a sine wave where each point of the tidal cycle can be mapped as an angle. If we take the cosine of the wave, high tide has a cosine of 1 ($\cos 0 = 1$), low tide has an angle of 180 ($\cos 180 = -1$), and points half way between (90 and 270 degrees) have cosines of 0. The

difference between the end of the run and the beginning can have a maximum value of 2 if sampling starts at high tide and ends at low tide; a value of -2 if it starts at low tide and ends at high tide, and 0 if it starts at the mid-ebb and ends at the mid-flood.

High and low tides were obtained from NOAA gauges at the Battery, the Narrows, Kill van Kull, and Kings Point on the eastern end of the East River.

Ambient Sampling Stations

Twenty ambient sampling stations are shown in Figure 10. There were two Long Island Sound sites (LISE and LISJ), and two New York Bight stations.

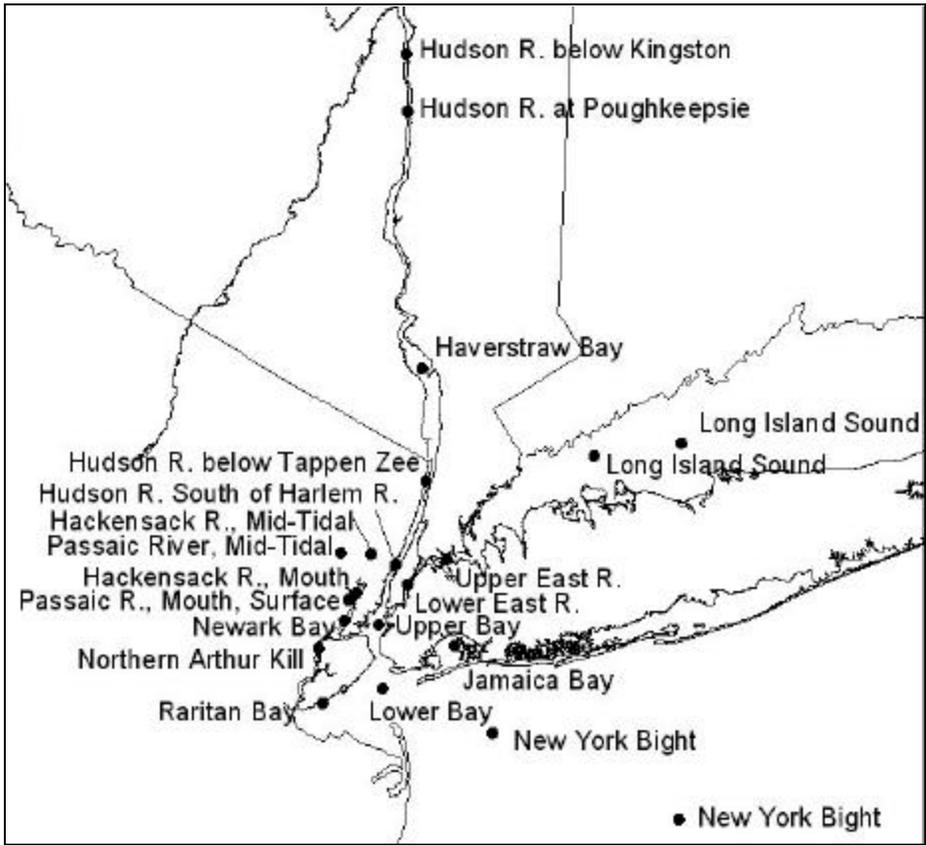


Figure 10.
Centroids of ambient
sampling stations.

Table 6 shows the ambient samples, the date sampling began, cosine tide, and mean sample values for DOC, POC, and SS. Some of the samples were performed over more than one tidal cycle and hence, no cosine tide value was calculated.

Table. 6. Ambient samples.

| Site type and name | Date | Cosine tide | DOC, mg/L | POC, mg/L | SS, mg/L |
|--|----------|----------------|--------------|--------------|-------------|
| Ambient-Non_Kills: Lower East R. | 9/18/98 | -0.41 | | 1.23 | 5.33 |
| Ambient-Non_Kills: Lower East R. | 3/11/99 | -1.52 | 3.79 | 1.18 | 19.00 |
| Ambient-Non_Kills: Lower East R. | 7/27/99 | 0.86 | 2.74 | 0.67 | 57.20 |
| Ambient-Non_Kills: Lower East R. | 6/2/00 | 1.78 | 3.35 | 0.37 | 18.60 |
| Ambient-Hudson: Haverstraw Bay | 11/24/98 | -1.50 | 4.38 | 0.20 | |
| Ambient-Hudson: Haverstraw Bay | 2/10/99 | -1.04 | 4.86 | 0.18 | 19.60 |
| Ambient-Hudson: Haverstraw Bay | 7/11/99 | 1.93 | 3.63 | 1.04 | 316.00 |
| Ambient-Hudson: Haverstraw Bay | 4/4/00 | | | 0.55 | 30.60 |
| Ambient-Hudson: Hudson R. South of Harlem R. | 12/17/98 | 0.47 | | 0.33 | 10.60 |
| Ambient-Hudson: Hudson R. South of Harlem R. | 3/16/99 | -1.11 | 4.29 | 0.97 | 47.50 |
| Ambient-Hudson: Hudson R. South of Harlem R. | 8/12/99 | 1.92 | 2.65 | 0.44 | 19.70 |
| Ambient-Hudson: Hudson R. South of Harlem R. | 12/14/99 | 1.88 | 3.88 | 0.42 | 11.88 |
| Ambient-Hudson: Hudson R. South of Harlem R. | 6/14/00 | 1.11 | 6.59 | 0.33 | 9.74 |
| Ambient-Hudson: Hudson R. below Kingston | 5/25/99 | -1.64 | 3.53 | 0.27 | |
| Ambient-Hudson: Hudson R. below Kingston | 10/8/99 | | 6.05 | 0.72 | 25.70 |
| Ambient-Hudson: Hudson R. below Kingston | 6/28/00 | | 4.86 | 0.49 | 16.50 |
| Ambient-Kills: Hackensack R., Mouth | 11/12/98 | | 6.97 | 0.62 | 5.90 |
| Ambient-Kills: Hackensack R., Mouth | 2/8/99 | -1.67 | 5.81 | 0.39 | 3.20 |
| Ambient-Kills: Hackensack R., Mouth | 7/7/99 | -1.88 | 5.48 | 0.46 | 24.40 |
| Ambient-Kills: Hackensack R., Mouth | 4/11/00 | -1.98 | 9.28 | 2.09 | 12.70 |
| Ambient-Kills: Hackensack R., Mid-Tidal | 3/17/99 | 0.31 | 13.94 | 2.77 | 43.50 |
| Ambient-Kills: Hackensack R., Mid-Tidal | 9/2/99 | -1.06 | 8.36 | 0.30 | 35.30 |
| Ambient-Kills: Hackensack R., Mid-Tidal | 10/12/99 | -0.71 | 7.46 | 1.25 | 37.00 |
| Ambient-Kills: Hackensack R., Mid-Tidal | 11/2/99 | | | | |
| Ambient-Kills: Hackensack R., Mid-Tidal | 5/10/00 | -0.72 | 9.15 | 1.56 | 46.10 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 3/1/99 | | 1.49 | 3.31 | 53.5 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 3/28/99 | | 2.21 | 3.34 | 61.54 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 4/16/99 | | 2.61 | 3.1 | 93.16 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 4/17/99 | | 2.15 | 3.06 | 96.32 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 10/23/99 | | 2.23 | 4.58 | 90.82 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 3/18/00 | | 4.07 | 3.59 | 146.1 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 5/17/00 | | 2.79 | 4.17 | 85.71 |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 6/15/00 | | 2.28 | 4.45 | 94.49 |
| Ambient-Hudson: Hudson R. below Tappen Zee | 12/1/98 | | 10.52 | 0.64 | |
| Ambient-Hudson: Hudson R. below Tappen Zee | 2/19/99 | | 4.73 | 0.61 | 38.20 |
| Ambient-Hudson: Hudson R. below Tappen Zee | 7/10/99 | 1.03 | 3.76 | 0.92 | 14.40 |
| Ambient-Hudson: Hudson R. below Tappen Zee | 4/4/00 | | 10.70 | 0.58 | 41.10 |
| Ambient-Non_Kills: Jamaica Bay | 10/14/98 | -1.98 | 2.45 | 0.65 | 7.50 |
| Ambient-Non_Kills: Jamaica Bay | 2/23/99 | -1.80 | 3.71 | 1.07 | 19.90 |
| Ambient-Non_Kills: Jamaica Bay | 7/9/99 | 0.21 | 4.55 | 1.40 | 33.50 |
| Ambient-Non_Kills: Jamaica Bay | 5/4/00 | 1.23 | 5.03 | 0.96 | 6.09 |
| Ambient-Non_Kills: Lower Bay | 12/3/98 | 0.04 | 2.60 | 0.19 | 2.80 |
| Ambient-Non_Kills: Lower Bay | 3/2/99 | 1.45 | 4.26 | 2.00 | 22.40 |
| Ambient-Non_Kills: Lower Bay | 7/28/99 | 1.13 | 2.58 | 1.04 | 13.80 |

Table 6 continued.

| Site type and name | Date | Cosine tide | DOC, mg/L | POC, mg/L | SS, mg/L |
|---|----------|----------------|--------------|--------------|-------------|
| Ambient-Non_Kills: Lower Bay | 6/1/00 | 0.43 | 3.54 | 1.09 | |
| Ambient-clean: Long Island Sound | 11/19/98 | | 3.17 | 0.17 | |
| Ambient-clean: Long Island Sound | 3/2/99 | | 2.74 | 0.13 | 5.27 |
| Ambient-clean: Long Island Sound | 5/27/99 | | 2.46 | 0.23 | |
| Ambient-clean: Long Island Sound | 10/19/99 | | 4.58 | | 4.69 |
| Ambient-Kills: Northern Arthur Kill | 11/17/98 | 1.67 | 3.75 | 0.61 | |
| Ambient-Kills: Northern Arthur Kill | 2/17/99 | 1.56 | 18.18 | 0.67 | 4.9 |
| Ambient-Kills: Northern Arthur Kill | 7/8/99 | -1.29 | | 0.71 | 21.9 |
| Ambient-Kills: Northern Arthur Kill | 4/18/00 | 0 | 8.43 | 0.99 | 9.95 |
| Ambient-Kills: Newark Bay | 11/25/98 | -1.07 | 3.52 | 0.28 | |
| Ambient-Kills: Newark Bay | 1/27/99 | -1.94 | 4.89 | 0.5 | 3.8 |
| Ambient-Kills: Newark Bay | 8/11/99 | 1.52 | 3.36 | 0.46 | 44.2 |
| Ambient-Kills: Newark Bay | 12/15/99 | -1.51 | 3.82 | 0.38 | 6.14 |
| Ambient-Kills: Newark Bay | 4/12/00 | -1.22 | 6.04 | 0.75 | 11.3 |
| Ambient-clean: New York Bight | 12/9/98 | | 1.5 | 0.11 | |
| Ambient-clean: New York Bight | 1/29/99 | | 1.73 | 0.07 | 7.8 |
| Ambient-clean: New York Bight | 1/30/99 | | 1.76 | | 7.85 |
| Ambient-clean: New York Bight | 1/31/99 | | 3.82 | 0.1 | 2.87 |
| Ambient-clean: New York Bight | 4/27/99 | | | 0.09 | 1.78 |
| Ambient-clean: New York Bight | 3/14/00 | | 19.03 | 0.13 | |
| Ambient-Kills: Passaic R., Mouth, Bottom | 2/5/99 | -0.24 | 7.07 | 0.26 | |
| Ambient-Kills: Passaic R., Mouth, Bottom | 7/21/99 | -1.25 | 4.4 | 1.64 | 30.1 |
| Ambient-Kills: Passaic R., Mouth, Bottom | 5/2/00 | 1.45 | 8.91 | 0.89 | 60.7 |
| Ambient-Kills: Passaic R., Mouth, Bottom | 6/26/00 | -1.61 | 7.22 | 1.24 | 16.2 |
| Ambient-Kills: Passaic R., Mouth, Surface | 11/13/98 | -1.81 | 6.07 | | 4.8 |
| Ambient-Kills: Passaic R., Mouth, Surface | 2/3/99 | 1.7 | 7.61 | 0.15 | 5.92 |
| Ambient-Kills: Passaic R., Mouth, Surface | 6/17/99 | 0.07 | 4.54 | 0.08 | 23.9 |
| Ambient-Kills: Passaic R., Mouth, Surface | 6/27/00 | -1.95 | 6.8 | 1.51 | 15.8 |
| Ambient-Kills: Passaic River, Mid-Tidal | 3/16/99 | 0.07 | 4.62 | 0.21 | 11.4 |
| Ambient-Kills: Passaic River, Mid-Tidal | 8/25/99 | 0.75 | 6.27 | 2.93 | 56.5 |
| Ambient-Kills: Passaic River, Mid-Tidal | 5/9/00 | -1.78 | | 1.49 | 10.3 |
| Ambient-Kills: Passaic River, Mid-Tidal | 10/18/00 | -1.01 | 6.78 | 1.49 | 44.5 |
| Ambient-Non_Kills: Raritan Bay | 11/16/98 | 0.25 | 2.36 | 0.22 | 1.2 |
| Ambient-Non_Kills: Raritan Bay | 2/24/99 | -1.83 | 3.15 | 1.35 | 10.9 |
| Ambient-Non_Kills: Raritan Bay | 7/12/99 | 1.85 | 4.04 | 0.64 | 10.3 |
| Ambient-Non_Kills: Raritan Bay | 5/3/00 | 1.78 | 3.56 | 1.56 | 10.3 |
| Ambient-Non_Kills: Upper Bay | 12/15/98 | -1.19 | 8.87 | 0.29 | 3.7 |
| Ambient-Non_Kills: Upper Bay | 3/18/99 | 1.99 | 3.31 | 0.73 | 15.6 |
| Ambient-Non_Kills: Upper Bay | 8/11/99 | 1.1 | 2.97 | | 13.3 |
| Ambient-Non_Kills: Upper Bay | 6/15/00 | | 5.69 | 0.21 | 6.62 |

Table 6 continued.

| Site type and name | Date | Cosine tide | DOC, mg/L | POC, mg/L | SS, mg/L |
|----------------------------------|-------|-------------|-----------|-----------|----------|
| Ambient-Non_Kills: Upper East R. | 36139 | -0.97 | 3.31 | 0.21 | |
| Ambient-Non_Kills: Upper East R. | 36222 | -1.05 | 6.1 | 0.94 | 24.9 |
| Ambient-Non_Kills: Upper East R. | 36382 | 0.63 | 2.58 | 0.18 | 2.58 |
| Ambient-Non_Kills: Upper East R. | 36592 | -0.65 | 17.31 | 0.63 | 9.35 |

Tributaries

The locations of sampling points on the three major and three minor tributaries are shown on Figure 11. The site on the Mohawk at Cohoes was always in the same position. Changes were made in the TOPS intake locations on the Hudson at Pleasantdale and on the Wallkill at New Paltz. On the minor tributaries, the sampling locations on the Saw Mill River and the Gowanus Canal remained constant but two stations were used on the Bronx River.

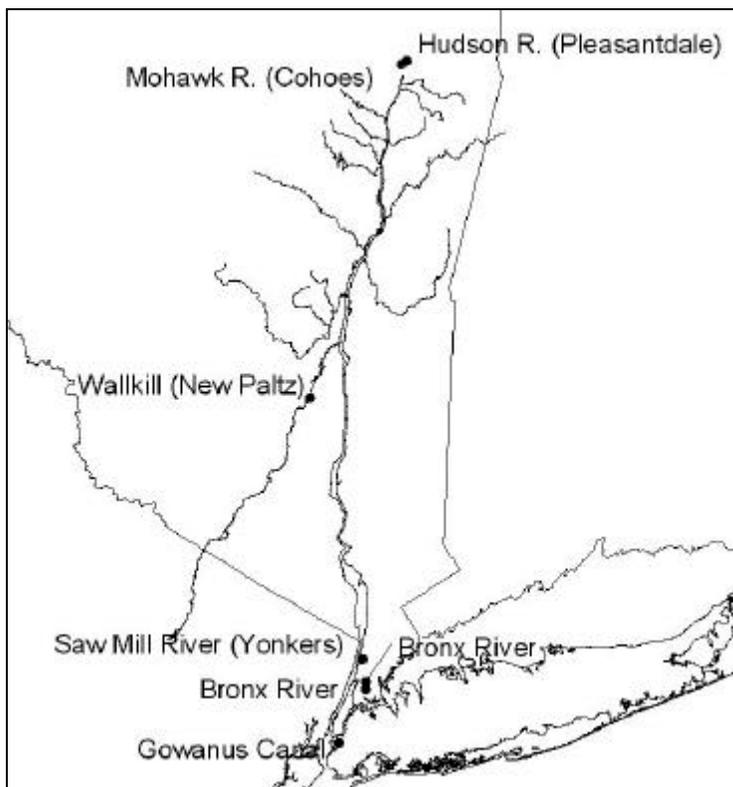


Figure 11. Locations of sampling points on major and minor tributaries.

Table 7 shows the samples taken from tributaries, their dates, discharges (in cubic feet per second), and average concentrations of POC, DOC, and suspended sediment.

Table 7. Major and minor tributary samples.

| Sample | CFS | Date_Start | Date_End | POC mg/L | DOC mg/L | SS mg/L |
|---|-------|------------|----------|-------------|-------------|------------|
| Major tributary: Hudson R. (Pleasantdale) | 18650 | 3/4/99 | 3/6/99 | | | |
| Major tributary: Hudson R. (Pleasantdale) | 18603 | 3/22/99 | 3/23/99 | 2.07 | 3.65 | 95.82 |
| Major tributary: Hudson R. (Pleasantdale) | 18887 | 4/1/99 | 4/7/99 | 1.23 | 3.22 | 18.04 |
| Major tributary: Hudson R. (Pleasantdale) | 18953 | 4/8/99 | 4/12/99 | 0.79 | 3.45 | 9.16 |
| Major tributary: Hudson R. (Pleasantdale) | 6467 | 9/20/99 | 9/30/99 | 0.47 | 4.59 | 10.63 |
| Major tributary: Hudson R. (Pleasantdale) | 13136 | 2/25/00 | 2/27/00 | 2.09 | 4.58 | 75.33 |
| Major tributary: Hudson R. (Pleasantdale) | 33634 | 2/28/00 | 2/28/00 | 5.77 | 4.05 | 293.46 |
| Major tributary: Hudson R. (Pleasantdale) | 28555 | 2/29/00 | 3/1/00 | 2.97 | 3.91 | 85.95 |
| Major tributary: Hudson R. (Pleasantdale) | 26562 | 3/29/00 | 3/30/00 | 3.30 | 3.39 | 53.61 |
| Major tributary: Hudson R. (Pleasantdale) | 27725 | 4/4/00 | 4/7/00 | 2.86 | 3.94 | 92.71 |
| Major tributary: Hudson R. (Pleasantdale) | 4967 | 8/29/00 | 8/31/00 | 0.29 | 4.89 | 2.36 |
| Major tributary: Hudson R. (Pleasantdale) | 2040 | 9/7/01 | 9/7/01 | | | |
| Major tributary: Mohawk R. (Cohoes) | 17040 | 3/4/99 | 3/23/99 | 1.94 | 3.64 | 1.94 |
| Major tributary: Mohawk R. (Cohoes) | 16515 | 4/1/99 | 4/7/99 | 1.05 | 2.97 | 23.15 |
| Major tributary: Mohawk R. (Cohoes) | 2339 | 5/10/99 | 5/20/99 | 0.43 | 3.67 | 8.56 |
| Major tributary: Mohawk R. (Cohoes) | 37790 | 9/17/99 | 9/17/99 | 5.05 | 3.92 | 158.49 |
| Major tributary: Mohawk R. (Cohoes) | 18396 | 2/26/00 | 2/27/00 | 2.86 | 3.52 | 92.66 |
| Major tributary: Mohawk R. (Cohoes) | 48240 | 2/28/00 | 2/28/00 | 10.76 | 3.84 | 478.05 |
| Major tributary: Mohawk R. (Cohoes) | 23650 | 3/12/00 | 3/13/00 | 2.81 | 4.27 | 130.54 |
| Major tributary: Mohawk R. (Cohoes) | 25950 | 3/28/00 | 3/31/00 | 3.75 | 3.80 | 143.93 |
| Major tributary: Mohawk R. (Cohoes) | 38240 | 4/4/00 | 4/5/00 | 4.78 | 3.92 | 194.46 |
| Major tributary: Mohawk R. (Cohoes) | 19790 | 4/21/00 | 4/26/00 | 1.12 | 4.13 | 33.05 |
| Major tributary: Mohawk R. (Cohoes) | 2600 | 8/29/00 | 8/31/00 | 0.46 | 4.54 | 5.09 |
| Major tributary: Mohawk R. (Cohoes) | 6060 | 2/26/01 | 3/5/01 | | | |
| Major tributary: Wallkill (New Paltz) | 279 | 5/17/99 | 5/18/99 | 0.70 | 5.90 | 10.00 |
| Major tributary: Wallkill (New Paltz) | 1350 | 9/17/99 | 9/19/99 | | | |
| Major tributary: Wallkill (New Paltz) | 608 | 10/13/99 | 10/27/99 | 0.70 | 7.40 | 9.00 |
| Major tributary: Wallkill (New Paltz) | 6346 | 2/15/00 | 2/16/00 | | | |
| Major tributary: Wallkill (New Paltz) | 2150 | 2/25/00 | 2/26/00 | | | |
| Major tributary: Wallkill (New Paltz) | 2551 | 7/27/00 | 7/28/00 | 2.20 | 6.80 | 108.00 |
| Major tributary: Wallkill (New Paltz) | 3202 | 8/15/00 | 8/17/00 | 7.40 | 9.30 | 114.00 |
| Major tributary: Wallkill (New Paltz) | 12134 | 12/17/00 | 12/18/00 | 19.70 | 4.50 | 580.00 |
| Major tributary: Wallkill (New Paltz) | 589 | 1/19/01 | 1/23/01 | 1.33 | 3.87 | 3.20 |
| Major tributary: Wallkill (New Paltz) | 1463 | 2/10/01 | 2/12/01 | 2.56 | 4.52 | 34.42 |
| Major tributary: Wallkill (New Paltz) | 6270 | 3/21/01 | 3/25/01 | 5.90 | 4.90 | 87.00 |
| Major tributary: Wallkill (New Paltz) | 6137 | 3/30/01 | 4/2/01 | 6.40 | 4.90 | 101.00 |
| Major tributary: Wallkill (New Paltz) | 1474 | 5/26/01 | 6/1/01 | 4.80 | 8.20 | 72.00 |
| Major tributary: Wallkill (New Paltz) | 589 | 6/17/01 | 6/19/01 | 1.30 | 5.30 | 39.00 |
| Major tributary: Wallkill (New Paltz) | 344 | 6/29/01 | 6/30/01 | 2.20 | 9.00 | 61.00 |
| Minor tributary: Bronx River | 16 | 10/29/98 | 10/29/98 | 0.23 | 7.09 | 3.60 |
| Minor tributary: Bronx River | 221 | 3/8/99 | 3/8/99 | 0.05 | 7.63 | 5.62 |
| Minor tributary: Bronx River | 6 | 7/27/99 | 7/27/99 | 0.76 | | 4.83 |
| Minor tributary: Bronx River | 8 | 10/26/99 | 10/26/99 | 0.34 | 4.95 | 3.22 |
| Minor tributary: Gowanus Canal | | 3/17/99 | 3/17/99 | 1.26 | 3.98 | 25.60 |
| Minor tributary: Gowanus Canal | | 8/24/99 | 8/24/99 | 0.41 | | 17.60 |

Table 7 continued.

| Sample | CFS | Date_Start | Date_End | POC mg/L | DOC mg/L | SS mg/L |
|---|-----|------------|----------|-------------|-------------|------------|
| Minor tributary: Gowanus Canal | | 3/21/00 | 3/21/00 | 1.32 | 6.47 | 6.07 |
| Minor tributary: Gowanus Canal | | 9/28/00 | 9/28/00 | 0.29 | 2.97 | 4.83 |
| Minor tributary: Saw Mill River (Yonkers) | 125 | 11/10/98 | 11/10/98 | 0.61 | 11.28 | 21.7 |
| Minor tributary: Saw Mill River (Yonkers) | 76 | 3/10/99 | 3/10/99 | 0.27 | 13.25 | 1.76 |
| Minor tributary: Saw Mill River (Yonkers) | 24 | 5/5/99 | 5/5/99 | 1.09 | 8.68 | 5 |
| Minor tributary: Saw Mill River (Yonkers) | 2 | 8/20/99 | 8/20/99 | 0.53 | | 2.13 |

Water Pollution Control Facilities (WPCFs).

The locations of the upstate WPCFs sampled by CARP are shown in Figure 12.

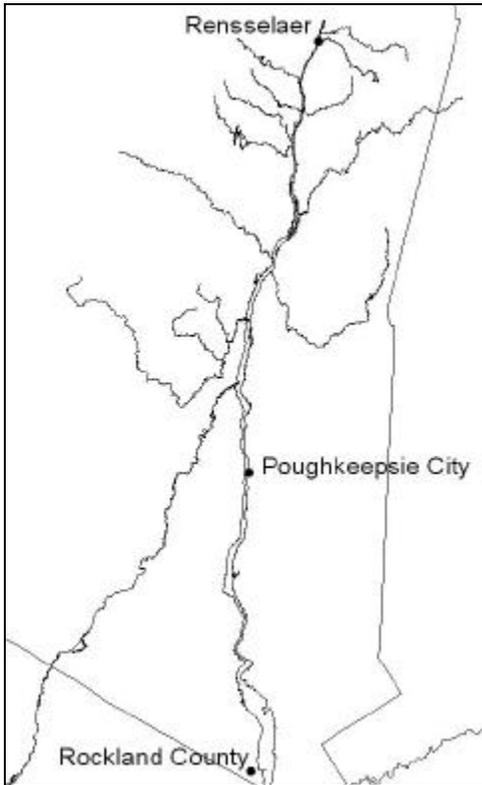


Figure 12. Upstate WPCFs samples by CARP.

Figure 13 shows the locations of New York City area WPCFs. All NYCDEP plants were sampled by CARP.

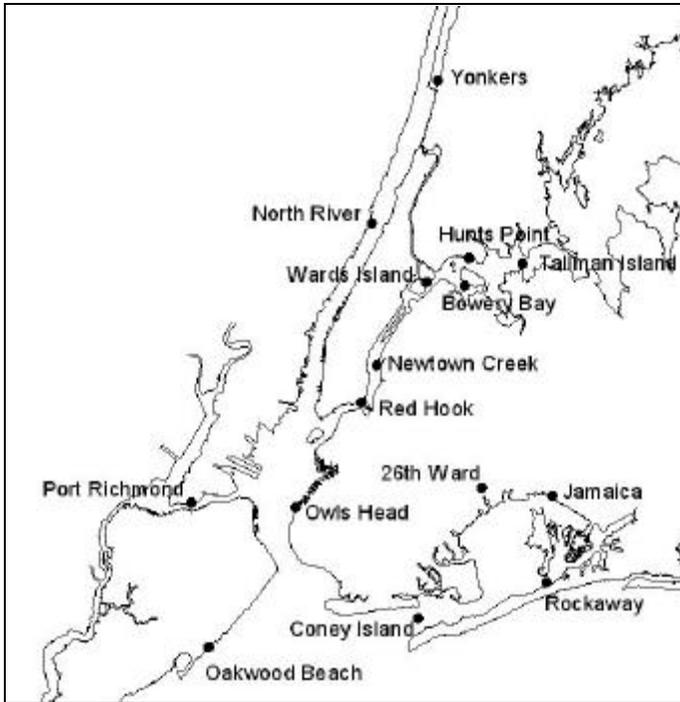


Figure 13. Locations of Yonkers and NYCDEP WPCF discharge points.

Table 8 shows summary sample data from the WPCF (sewage treatment plant) samples. WPCF discharges are conventionally shown in mgd.

Table 8. WPCF samples.

| WPCF | Date Start | MGD | POC mg/L | DOC mg/L | SS mg/L |
|--------------|------------|-----|-------------|-------------|------------|
| 26th Ward | 1/27/99 | 53 | 2.39 | 10.41 | 12.8 |
| 26th Ward | 5/5/99 | 60 | 2.4 | 8.97 | 8.38 |
| 26th Ward | 9/20/00 | 83 | 0.61 | 7.06 | 4.95 |
| 26th Ward | 6/11/01 | 64 | | | |
| 26th Ward | 6/18/01 | 68 | | | |
| Bowery Bay | 11/5/98 | 101 | 0.59 | 9.88 | 4.35 |
| Bowery Bay | 4/21/99 | 138 | 5.14 | 10.8 | 17 |
| Bowery Bay | 9/22/99 | 103 | 0 | 7.44 | 2.98 |
| Coney Island | 3/17/99 | 105 | 2.99 | 8.81 | 10.3 |
| Coney Island | 7/28/99 | 103 | 0.76 | 7.89 | 2.7 |
| Coney Island | 10/4/00 | 87 | 0.98 | 7.7 | 3.84 |
| Edgewater | 5/21/01 | 3 | | | |
| Hunts Point | 4/18/01 | 125 | | | |
| Hunts Point | 2/19/99 | 149 | 3.41 | 9.05 | 5.85 |
| Hunts Point | 4/30/99 | 133 | 0.35 | 9.66 | 48.1 |
| Hunts Point | 2/1/01 | 142 | 0.68 | | 10.2 |
| Hunts Point | 3/19/01 | 120 | | | |
| Hunts Point | 3/28/01 | 181 | | | |
| Hunts Point | 4/11/01 | 146 | | | |

Table 8 continued.

| WPCF | Date Start | MGD | POC mg/L | DOC mg/L | SS mg/L |
|-------------------|------------|-----|-------------|-------------|------------|
| Jamaica | 2/5/99 | 84 | 6.68 | 10.58 | 14.5 |
| Jamaica | 6/30/99 | 90 | 1.62 | | 6.7 |
| Jamaica | 2/15/01 | 88 | 0.7 | | 11.6 |
| Newtown Creek | 4/8/01 | 248 | | | |
| Newtown Creek | 4/30/01 | 240 | | | |
| Newtown Creek | 5/21/01 | 416 | | | |
| Newtown Creek | 3/11/99 | 257 | 2.69 | 20.54 | 44.4 |
| Newtown Creek | 6/22/99 | 260 | | 28.27 | 33.5 |
| Newtown Creek | 9/28/99 | 275 | 10.37 | 24.57 | 25.8 |
| Newtown Creek | 1/5/00 | 249 | 13.1 | | |
| Newtown Creek | 1/5/00 | 249 | 13.1 | | |
| Newtown Creek | 3/28/01 | 335 | | | |
| North River | 3/24/99 | 153 | 1.56 | 11.62 | 4.08 |
| North River | 9/1/99 | 167 | 1.28 | 9.91 | 4.28 |
| North River | 1/25/01 | 152 | 0.32 | | 7.67 |
| Oakwood Beach | 2/11/99 | 25 | 0.97 | 10.99 | 6.33 |
| Oakwood Beach | 8/18/99 | 25 | 1.2 | 9.37 | 2.28 |
| Oakwood Beach | 10/13/99 | 36 | | 9.27 | 3.98 |
| Owls Head | 9/15/98 | 113 | 4.88 | | 26.7 |
| Owls Head | 7/7/99 | 119 | 1.44 | | 7.01 |
| Owls Head | 8/23/00 | 115 | 1.88 | 8.98 | 6.41 |
| Port Richmond | 2/24/99 | 31 | 6.52 | 19.1 | 15.6 |
| Port Richmond | 8/25/99 | 35 | 1.38 | 13.04 | 2.58 |
| Port Richmond | 10/20/99 | 78 | 4.13 | 17.67 | 10.4 |
| Port Richmond | 4/11/01 | 49 | | | |
| Port Richmond | 4/30/01 | 29 | | | |
| Poughkeepsie City | 4/1/99 | 7 | 5.51 | 11.17 | 15.1 |
| Poughkeepsie City | 8/19/99 | 5 | 38.63 | 28.97 | 85.7 |
| Poughkeepsie City | 12/5/00 | 4 | 3.54 | | 9.44 |
| PVSC | 5/22/01 | 318 | | | |
| Red Hook | 2/3/99 | 40 | 2.43 | 837.12 | 7.04 |
| Red Hook | 4/14/99 | 30 | 1.92 | 12.64 | 7.71 |
| Rensselaer | 1/12/99 | 16 | 4.43 | 25.48 | 15.6 |
| Rensselaer | 3/30/99 | 23 | 0.91 | 19.44 | 6.63 |
| Rensselaer | 8/11/99 | 14 | 1.71 | | 6.04 |
| Rockaway | 4/1/99 | 21 | 0.33 | 8.17 | 3.59 |
| Rockaway | 8/11/99 | 22 | | 9.03 | 18.3 |
| Rockaway | 11/3/99 | 19 | 0.45 | 7.28 | 6.38 |
| Rockland County | 4/20/99 | 20 | 3.02 | 16.18 | 16.6 |
| Rockland County | 8/19/99 | 17 | 1.37 | 18.61 | 2.94 |
| Rockland County | 3/8/00 | 22 | 8.94 | 29.21 | |

Table 8 continued.

| WPCF | Date Start | MGD | POC mg/L | DOC mg/L | SS mg/L |
|----------------|------------|-----|-------------|-------------|------------|
| Tallman Island | 2/12/99 | 56 | 2.32 | | |
| Tallman Island | 7/20/99 | 59 | 0.89 | 8.66 | 4.1 |
| Tallman Island | 9/6/00 | 41 | 1.2 | 8.3 | 3.73 |
| Wards Island | 1/20/99 | 221 | 1.57 | 7.81 | 4.83 |
| Wards Island | 4/28/99 | 179 | 0.09 | 7.33 | 2.48 |
| Wards Island | 8/10/00 | 220 | 0.88 | 5.73 | 2.51 |
| Yonkers | 4/22/99 | 89 | 1.73 | 10.3 | 6.8 |
| Yonkers | 8/18/99 | 85 | 1.16 | 10.62 | 13.5 |
| Yonkers | 3/22/00 | 95 | 10.12 | 19.15 | 26.4 |

Industrial Effluents and Landfill Leachates

Figure 14 shows the locations of industrial effluents and leachates sampled by CARP.

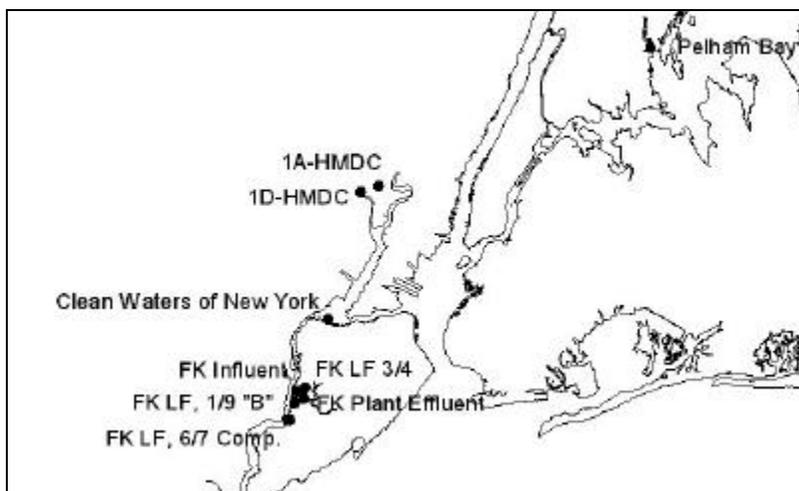


Figure 14. Locations of industrial effluents and landfill leachates sampled by CARP.

Table 9 lists the samples taken for Industrial effluents and Landfill leachates. Names, dates, DOC, POC, and TSS are also given.

Table 9. Industrial effluent and landfill leachate samples.

| Sample | Date | DOC mg/L | POC mg/L | SS mg/L |
|---|----------|-------------|-------------|------------|
| Industrial effluent: Clean Waters of New York | 4/29/99 | 6.02 | 0.19 | |
| Industrial effluent: Clean Waters of New York | 9/20/99 | 6.79 | 0.25 | 24.9 |
| Industrial effluent: FK Plant Effluent | 10/25/00 | 176 | | 4.63 |
| Industrial effluent: FK Plant Effluent | 3/20/01 | | | |
| Industrial effluent: FK Plant Effluent | 4/19/01 | | 0.31 | 30.7 |
| Industrial effluent: FK Plant Effluent | 7/25/01 | | 0.45 | 8.49 |
| Landfill leachate: 1A-HMDC | 6/22/00 | | | |
| Landfill leachate: 1D-HMDC | 6/22/00 | 235 | | |
| Landfill leachate: 1D-HMDC | 9/14/01 | | | |
| Landfill leachate: 1E-HMDC | 6/22/00 | 430 | | |
| Landfill leachate: 1E-HMDC | 9/14/01 | | | |
| Landfill leachate: FK Influent | 4/5/99 | | | |
| Landfill leachate: FK Influent | 6/3/99 | | | |
| Landfill leachate: FK LF 3/4 | 5/11/00 | 120 | | |
| Landfill leachate: FK LF, 1/9 "B" | 5/11/00 | 490 | | |
| Landfill leachate: FK LF, 1/9 "F" | 5/11/00 | 34.9 | | |
| Landfill leachate: FK LF, 1/9 Comp. | 5/11/00 | 365 | | |
| Landfill leachate: FK LF, 1/9 Comp. | 10/25/00 | 1680 | | |
| Landfill leachate: FK LF, 1/9 Comp. | 3/20/01 | | | |
| Landfill leachate: FK LF, 1/9 Comp. | 4/19/01 | | | |
| Landfill leachate: FK LF, 6/7 Comp. | 5/11/00 | 161 | | |
| Landfill leachate: FK LF, 6/7 Comp. | 10/25/00 | 821 | | |
| Landfill leachate: FK LF, 6/7 Comp. | 7/25/01 | | | |
| Landfill leachate: FK LF, 6/7 Comp. | 8/9/01 | | | 26.1 |
| Landfill leachate: Pelham Bay | 11/6/98 | 143 | 1.42 | 4.80 |
| Landfill leachate: Pelham Bay | 1/29/01 | | | |

CSOs and SWOs

Locations of the CSO and SWO samples are indicated on Figure 15.

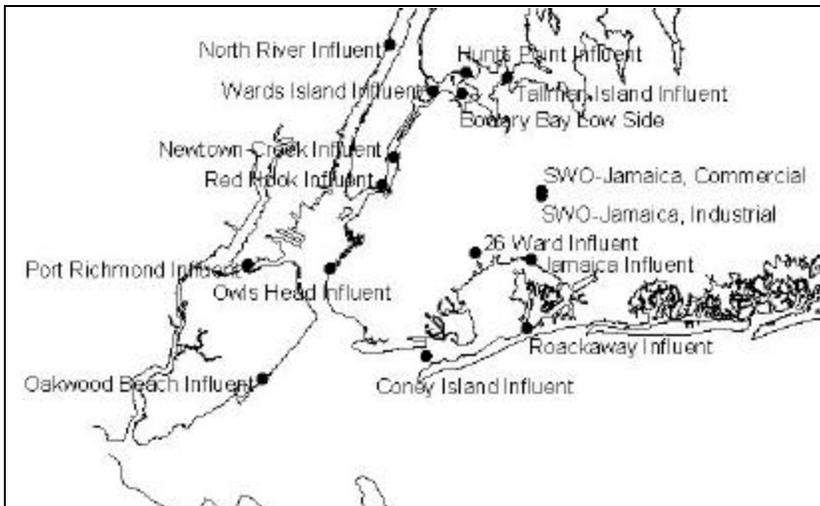


Figure 15. CSO and SWO sampling sites.

Table 10. CSOs and SWOs, names, dates, DOC, POC, and SS.

| Short_Name | MGD | Date | DOC mg/L | POC mg/L | SS mg/L |
|----------------------------------|-----|----------|-------------|-------------|------------|
| 26 th Ward, High Side | 12 | 6/2/01 | | | |
| 26 th Ward, Low Side | 12 | 5/21/01 | | | |
| Bowery Bay High Side | 13 | 3/21/01 | | | |
| Bowery Bay Low Side | 13 | 2/25/01 | | | |
| Coney Island Influent | 10 | 11/26/00 | | 65.7 | 180 |
| Hunts Point Influent | 15 | 7/8/01 | | | |
| Jamaica Influent | 31 | 9/20/01 | | | |
| Manhattan Grit Chamber | 11 | 9/24/01 | | | 92 |
| Manhattan Pump Station | 14 | 2/5/01 | | 0.667 | |
| Newtown Creek Influent | 14 | 1/30/01 | | | |
| North River Influent | 5.0 | 6/23/01 | | | |
| Owls Head Influent | 9.3 | 11/9/00 | | | 169 |
| Port Richmond Influent | 1.0 | 12/16/00 | 342 | | 298 |
| Red Hook Influent | 3.7 | 8/27/01 | | | 404 |
| SWO-Jamaica, Commercial | | 6/22/00 | 260 | 28 | |
| SWO-Jamaica, Industrial | | 10/16/00 | | 0.390 | 158 |

THE CHEMICALS

The CARP chemicals are polychlorinated biphenyls (PCBs), dioxins/furans, chlorinated pesticides, polynuclear aromatic hydrocarbons (PAHs), and the metals mercury and cadmium. Accessory parameters of particulate organic carbon (POC), dissolved organic carbon (DOC), and suspended sediment (SS) were also measured.

PCBs

PCBs and pesticides samples were usually acquired by TOPS. Extracts came from XAD resin and the glass fiber filters. On some occasions, samples were also taken from whole water grab samples, hexane (PISCES samples), sludges, and, for purposes of quality control, sediments. Details of the sampling procedures are to be found in the TOPS Standard Operating Procedure (SOP).

PCBs were measured by USEPA Method 1668A. The original Method 1668 was developed to measure the “co-planar” or “toxic” PCBs. At the outset of CARP one of the participating labs, Axys Analytical Services, suggested using an advanced version of 1668 called 1668A to measure all 209 PCB congeners. Methods 1668 and 1668A are descendants of 1613 in that they are isotopic dilution HRGC/HRMS methods. The modifications used were (a) using a single GC column (SPB-Octyl) which resulted in not all of the 209 congeners being resolved and (b) a 100 μ L final extract volume which resulted in a 5-fold increase in the detection levels for each congener. The SPB-Octyl chromatographic column is short-lived and less familiar to many labs. This method was still experimental and has proved to be difficult for some of the labs to use.

Method 1668A does not resolve each of the 209 PCB congeners. During CARP Axys usually reported 159 domains consisting of from one (126 congeners) to 6 congeners or coelutions. Coeluted congeners are virtually identical. With two exceptions, each of the co-planar PCBs is resolved. The exceptions, IUPACs 156 and 157, have the same WHO98 TEF. Since these are HRMS data, all the coelutions have the same molecular weight. By convention, CARP reports all the coelutions under the name of the congener with the numerically lowest IUPAC designation.

PCBs can be treated as dioxins by summing the products of all congeners and their TEFs or they can be summed to obtain a total PCB. The New York State Ambient Water Quality Standard (for protection of humans eating fish) sums all PCBs and is 1 μ g/L. NYSDEC WQS do not recognize the co-planar PCBs. PCBs may have from one to 10 chlorine atoms. These result in one to 10 homologues. The relative abundances of the homologues can be useful in determining the source of the PCB.

PCBs were intentionally manufactured in the United States under the “Aroclor” trademark. Table 11 relates percent homologue abundance patterns to four Aroclors⁴.

⁴ Shultz, D.E., Petrick, G., and Duniker, J.C. 1989. Complete characterization of polychlorinated biphenyls in commercial Aroclor and Clophen mixtures by multidimensional gas chromatography-electron capture detection. *ES&T* 23, 852-859.

Table 12, also using data from Shultz et al, shows the percent abundances of congeners unique (some overlap, less than 10%, was permitted) to the lighter Aroclors (1016/1242) and heavier Aroclors (1254/1260). Typical Axys coelutions are indicated.

Table 11. Percent homologue abundances in four Aroclors.

| Homologues | 1016 | 1242 | 1254 | 1260 |
|------------|-------|-------|-------|-------|
| 1-mono | | | | |
| 2-di | 21.47 | 14.95 | | |
| 3-tri | 49.76 | 35.33 | 1.21 | 0.1 |
| 4-tetra | 27.83 | 32.64 | 16.61 | 0.99 |
| 5-penta | 0.99 | 13.16 | 50.96 | 13.51 |
| 6-Hexa | 0.19 | 2.39 | 23.86 | 46.98 |
| 7-Hepta | | 0.22 | 4.38 | 33.83 |
| 8-Octa | | | 0.68 | 7.27 |
| 9-Nona | | | | 0.67 |
| 10-Deca | | | | 0.05 |

Table 12. PCB congeners “uniquely” characteristic of Aroclors 1016/1242 (“Light”) and 1254/1260 (“Heavy”). In percent abundance.

| IUPAC | Group | 1016 | 1242 | 1254 | 1260 | Coelution | IUPAC | Group | 1016 | 1242 | 1254 | 1260 | Coelution |
|-------|-------|------|------|------|------|-----------|-------|-------|------|------|------|------|-----------|
| 4 | Light | 3.89 | 3.01 | | | | 131 | Heavy | | | 0.16 | 0.16 | |
| 5 | Light | 0.13 | 0.06 | | | | 134 | Heavy | | | 0.49 | 0.62 | 143 |
| 6 | Light | 1.83 | 1.38 | | | | 135 | Heavy | | 0.08 | 1.62 | 2.56 | 151, 154 |
| 7 | Light | 0.6 | 0.6 | | | | 136 | Heavy | | 0.07 | 1.12 | 2.23 | |
| 8 | Light | 10.8 | 7.65 | | | | 137 | Heavy | | | 0.25 | 0.06 | |
| 9 | Light | 0.95 | 0.54 | | | | 138 | Heavy | 0.19 | 0.54 | 3.2 | 6.13 | 129 |
| 10 | Light | 0.37 | 0.2 | | | | 141 | Heavy | | | 1.04 | 2.56 | |
| 15 | Light | 2.9 | 1.51 | | | | 143 | Heavy | | | | | 134 |
| 16 | Light | 2.86 | 2.01 | | | | 146 | Heavy | | | 0.83 | 1.49 | |
| 17 | Light | 3.84 | 2.88 | 0.19 | | | 147 | Heavy | | | | | 149 |
| 18 | Light | 9.03 | 6.28 | 0.41 | | 30 | 151 | Heavy | | | 1.17 | 3.67 | 135 |
| 19 | Light | 0.96 | 0.53 | | | | 153 | Heavy | | 0.68 | 4.26 | 10.8 | 168 |
| 20 | Light | 1 | 0.29 | | | 28 | 154 | Heavy | | | | | 135 |
| 21 | Light | | | | | 33 | 156 | Heavy | | 0.09 | 1.62 | 0.88 | 157 |
| 22 | Light | 4.8 | 3.41 | | | | 157 | Heavy | | | | 0.14 | 156 |
| 24 | Light | 0.3 | 0.22 | | | | 158 | Heavy | | | 0.77 | 1.55 | |
| 25 | Light | 1.19 | 0.79 | | | | 160 | Heavy | | | | 0.05 | 129 |
| 26 | Light | 1.92 | 1.33 | | | 29 | 163 | Heavy | | | | | 129 |
| 27 | Light | 0.47 | 0.28 | | | | 166 | Heavy | | | | | 128 |

Table 12 continued.

| IUPAC Group | 1016 | 1242 | 1254 | 1260 | Coelution | IUPAC Group | 1016 | 1242 | 1254 | 1260 | Coelution |
|-------------|-------|------|------|------|-----------|---------------|------|-------|------|------|-----------|
| 28 | Light | 8.71 | 6.52 | 0.25 | 0.05 | 20 | 167 | Heavy | 0.21 | 0.26 | |
| 29 | Light | 0.19 | 0.1 | | | 26 | 170 | Heavy | 0.11 | 0.31 | 3.91 |
| 30 | Light | | | | | 18 | 171 | Heavy | 0.05 | 0.5 | 2.16 |
| 32 | Light | 1.34 | 0.88 | | | | 172 | Heavy | 0.05 | 0.75 | |
| 33 | Light | 6.25 | 4.79 | 0.14 | | 21 | 173 | Heavy | 0.09 | 0.36 | 171 |
| 34 | Light | 0.12 | 0.05 | | | | 174 | Heavy | 0.34 | 3.85 | |
| 35 | Light | 0.08 | 0.11 | | | | 175 | Heavy | 0.05 | 0.23 | |
| 37 | Light | 0.3 | 0.27 | | | | 176 | Heavy | 0.32 | 0.95 | |
| 45 | Light | 1.66 | 1.16 | | | 51 | 177 | Heavy | 0.21 | 2.21 | |
| 46 | Light | 0.7 | 0.49 | | | | 178 | Heavy | 1.35 | 1.62 | |
| 51 | Light | 0.36 | 0.23 | | | 45 | 179 | Heavy | 0.21 | 1.79 | |
| 59 | Light | 0.29 | 0.34 | | | 62, 75 | 180 | Heavy | 0.06 | 0.38 | 7.12 |
| 62 | Light | | | | | 59 | 183 | Heavy | 0.17 | 1.76 | 185 |
| 69 | Light | | 0.11 | | | 49 | 185 | Heavy | | 1.34 | 183 |
| 75 | Light | 0.08 | 0.11 | | | 59 | 187 | Heavy | 0.32 | 3.97 | |
| 122 | Heavy | | | 0.5 | 0.3 | | 190 | Heavy | 0.08 | 0.79 | |
| 128 | Heavy | | | 2.07 | 1.06 | 166 | 193 | Heavy | | 0.66 | 180 |
| 129 | Heavy | | | 0.23 | 1.11 | 138, 160, 163 | 201 | Heavy | 0.68 | 0.99 | |
| 130 | Heavy | | | 0.63 | 0.08 | | | | | | |

Dioxins/Furans

Dioxins and furans were usually quantified only from suspended materials recovered by filters. Some XAD samples were analyzed for the dioxins but the margin between the detection limit and the amount recovered was usually uncomfortably small. There were also some whole water samples analyzed for the dioxins. Towards the end of the project a number of experiments were performed using metered surrogates of dioxins to examine the efficiency of XAD for these chemicals. Details of the sampling procedure for dioxins can be found in the TOPS SOP.

In the lab, the chlorinated dioxins and furans are measured using EPA Method 1613. This isotopic high-resolution gas chromatography/high resolution mass spectrometry procedure is well established and familiar to all the project labs.

Seven chlorinated dioxins and 10 chlorinated furans are considered. Each of these 17 chemicals is regarded as having a similar toxicological mode of action but also to have greatly differing potencies. These potencies are expressed as Toxic Equivalency Factors (TEFs). They also have differing potentials for bioaccumulation that are expressed as Bioaccumulation Equivalency Factors (BEF). The NYS Ambient Water Quality

Standard for the chlorinated dioxins and furans is the sum of the products of the observed concentrations and their TEFs and BEFs. The result is called the dioxin equivalents (TEQ). The NYSDEC Ambient Water Quality Standard (for protection of humans eating fish) for TEF, BEF chlorinated dioxins and furans is 0.6 femtograms/L (parts per quintillion).

Table 13. Dioxin/furan TEFs and BEFs

| PARAMETER | WHO94 | BEF |
|---------------------|-------|------|
| 2,3,7,8-TCDD | 1 | 1 |
| 1,2,3,7,8-PeCDD | 0.5 | 0.9 |
| 1,2,3,4,7,8-HxCDD | 0.1 | 0.3 |
| 1,2,3,6,7,8-HxCDD | 0.1 | 0.1 |
| 1,2,3,7,8,9-HxCDD | 0.1 | 0.1 |
| 1,2,3,4,6,7,8-HpCDD | 0.01 | 0.05 |
| OCDD | 0.001 | 0.01 |
| 2,3,7,8-TCDF | 0.1 | 0.8 |
| 1,2,3,7,8-PeCDF | 0.05 | 0.2 |
| 2,3,4,7,8-PeCDF | 0.5 | 1.6 |
| 1,2,3,4,7,8-HxCDF | 0.1 | 0.08 |
| 1,2,3,6,7,8-HxCDF | 0.1 | 0.2 |
| 2,3,4,6,7,8-HxCDF | 0.1 | 0.7 |
| 1,2,3,7,8,9-HxCDF | 0.1 | 0.6 |
| 1,2,3,4,6,7,8-HpCDF | 0.01 | 0.01 |
| 1,2,3,4,7,8,9-HpCDF | 0.01 | 0.4 |
| OCDF | 0.001 | 0.02 |

Pesticides

Chlorinated pesticides were analyzed using a modification of USEPA Method 1613B. This is also a method high-resolution gas chromatography/high-resolution mass-spectrometry, combined with partial isotope dilution. A DB-5 column was used with a 200 µL final volume. Twenty-seven chlorinated pesticides were determined using 5 C-13 labeled and one deuterium labeled standards.

Table 14. CARP pesticides.

| PARAMETER | WQS (ug/L) |
|-------------------------|------------|
| 2,4'-DDD | NA |
| 2,4'-DDE | NA |
| 2,4'-DDT | NA |
| 4,4'-DDD | 0.00008 |
| 4,4'-DDE | 0.000007 |
| 4,4'-DDT | 0.00001 |
| Aldrin | 0.001 |
| HCH, alpha | 0.002 |
| HCH, beta | 0.007 |
| HCH, gamma | 0.008 |
| Chlordane,alpha (cis) | 1 |
| Chlordane,gamma (trans) | 1 |
| Chlordane,oxy- | NA |
| Heptachlor | 0.0002 |
| Hexachlorobenzene | 0.00003 |
| Mirex | 0.000001 |
| Nonachlor, cis- | NA |
| Nonachlor, trans- | NA |
| Dieldrin | 6E-07 |
| Endosulfan sulfate | NA |
| Endosulfan, alpha | 0.001 |
| Endosulfan, beta | 0.001 |
| Endrin | 0.002 |
| Endrin aldehyde | NA |
| Endrin ketone | NA |
| Heptachlor epoxide | 0.0003 |
| Methoxychlor | 0.03 |

PAHs

PAHs were determined using high resolution gas chromatography with Selected Ion Monitoring low resolution mass spectrometry. A DB-5 column was used with a final volume of 500 μ L.

The XAD resins used in TOPS release some of the PAHs targeted in the CARP. Therefore, dissolved phase PAHs were taken from the effluent of the TOPS cartridge filters. PAHs attached to particles were measured from glass fiber cartridge extracts.

The list of PAHs CARP uses includes a few where there are one or two methyl substitutions. These are called "C1" or "C2" as in "C1-Naphthalene". PAHs can be summed but they may also be summed as a molar concentration. In this approach the individual chemical concentrations are divided by their molar weight and then added together.

Table 15. CARP PAHs.

| PARAMETER | WQS (ug/L) | Molecular Wt. |
|------------------------------|------------|---------------|
| 1-Methylnaphthalene | NA | 142.2 |
| 2-Methylnaphthalene | NA | 142.2 |
| Acenaphthene | 6.6 | 154.2 |
| Acenaphthylene | NA | 152.2 |
| Anthracene | NA | 178.2 |
| Benz[a]anthracene | NA | 228.29 |
| Benzo[a]pyrene | 0.0006 | 252.3 |
| Benzo[b/j/k]fluoranthenes | NA | 252.3 |
| Benzo[b]fluoranthene | NA | 252.3 |
| Benzo[e]pyrene | NA | 228.3 |
| Benzo[ghi]perylene | NA | 228.3 |
| Benzo[k]fluoranthene | NA | 252.3 |
| Biphenyl | NA | 154.2 |
| C1 Naphthalenes | NA | 142.2 |
| C1 Phenanthrenes/Anthracenes | NA | 192.26 |
| C2 Naphthalenes | NA | 156.23 |
| C3 Naphthalenes | NA | 170.26 |
| Chrysene | NA | 228.3 |
| Dibenz[a,h]anthracene | NA | 278.4 |
| Fluoranthene | NA | 202.3 |
| Fluorene | 2.5 | 166.2 |
| Indeno[1,2,3-cd]pyrene | NA | 276.3 |
| Naphthalene | 16 | 128.2 |
| Perylene | NA | 252.32 |
| Phenanthrene | 1.5 | 178.2 |
| Pyrene | NA | 202 |

Metals

Field contamination has posed a substantial problem in sampling trace levels of metals, particularly mercury. Overcoming this problem requires a great deal of vigilance and the procedure has been formalized into an act of ritual cleanliness called “Clean Hands/Dirty Hands”. It’s also called USEPA Method 1669. Early in the life of the project we brought Michelle Gauthier from Frontier Geosciences in Seattle, WA to help us with the sampling technique. We took her out on the East River and to the Newtown Creek WPCF to see two typical and different sampling environments.

Details of the procedure to avoid contaminating the sample in the field are given in the TOPS SOP. Essentially, one person handles the placement of water into the bottles and a second helper opens Ziploc bags holding the sampling equipment while a third takes field notes. Filtration for dissolved metals is performed in the field.

During trackdown, metals were collected by duct taping a clean sample bottle to a weighted line, submersing the bottle, capping it immediately on recovery, and re-double bagging it.

All metals sample processing took place using ultra-clean handling techniques in a class 100 clean area known to be low in atmospheric mercury. Reagents, gases, and reagent water were all reagent or ultra-pure grade and previously analyzed for trace metals to ensure very low blanks.

Water samples were prepared according to Frontier Geosciences SOP #FGS-032. Metals (Ag, Cd, and Pb) preserved to pH 1.8 with HNO₃ are extracted with Co-APDC, and the precipitate is collected by filtration. The precipitate is then dissolved in concentrated HNO₃, then diluted in 5% HNO₃ to 10 mL. This method allows for the removal of the analytes of interest from the sample matrix, and makes possible up to 20-fold concentration of the sample.

Silver, cadmium, and lead were determined using inductively coupled-mass spectrometry (ICP-MS, US EPA Method 1640, modified) with a Perkin-Elmer Elan 6000. All results are reported instrument and preparation blank corrected.

Mercury analyses were performed using cold vapor atomic fluorescence spectrometry (CV-AFS). Total mercury standards are prepared by direct dilution of NIST-certified NBS-3133 mercury standard solution and results are independently verified by analysis of NIST 1641d. For the digestion/oxidation of water samples, BrCl was added to an aliquot of the sample at a level of 1-5 mL/ 100 mL of sample depending on apparent level of organics and turbidity of the samples. Sample oxidation took place on the same day of sample receipt. Samples were allowed to digest overnight at room temperature. Digests were analyzed for total Hg by CV-AFS. Aliquots of each digest were reduced in pre-purged reagent water to Hg⁰ with SnCl₂ and then the Hg⁰ purged onto gold traps as a preconcentration step. The Hg contained ion the gold traps was then analyzed by thermal desorption into an atomic fluorescence detector using the dual amalgamation technique.

For methyl mercury analysis water samples were distilled to liberated MeHg (US EPA Draft Method 1630). For water samples, 45 mL of 0.4% (v/v) HCl acidified sample was distilled using 50 mL Teflon distillation tubes. To each sample, 0.2 mL of 1% APDC solution was added prior to distillation, to enhance reproducibility and recovery. The distillate was received into a tube containing 5.0 mL of DDW to start, and distilled to 40.0 mL. Thus 35 mL out of 45 mL of sample was distilled over for analysis. The historic mean MeHg distillation recovery is 90.6 %. All net MeHg results are corrected for this efficiency factor.

Distilled samples were analyzed using aqueous phase ethylation, purging onto a Carbotrap, isothermal GC separation, and CV-AFS detection. Prior to ethylation, the distillate was diluted to 55 mL with DIW, and the pH brought to 4.9 with acetate buffer. Samples were ethylated by the addition of sodium tetraethyl borate, and then the volatile ethyl analogs purged with N₂ onto Carbotrap. After a trap drying step, the mercury ethyl

analogs were thermally desorbed into a 1 m isothermal GC column (15% OV-3 on Chromosorb WAW-DMSC) held at 100°C for separation. The column resolves elemental Hg, dimethyl Hg, methyl ethyl Hg, and diethyl Hg. Only methyl ethyl Hg, the MeHg analog, is quantified. The organo-Hg compounds are pyrolytically decomposed to Hg⁰ prior to entering the CV-AFS detector.

Water quality standards for the metals appear in Table 16.

Table 16. Water Quality Standards for metals.

| | ng/L | Water Class | Type | form |
|---------|-------|------------------------|-------|-----------|
| Cadmium | 7,700 | SA, SB, SC, I | A(C) | |
| Lead | 8,000 | SA, SB, SC, I | A(C) | |
| Mercury | 0.7 | SA, SB, SC, I, SD | H(FC) | dissolved |
| Silver | 100 | A, A-S, AA, AA-S, B, C | A(C) | ionic |

SS/DOC/POC

Suspended solids (SS) measures the total amount of particulate material suspended in the water column. At the beginning of the project we collected bottles of water that would be passed through pre-weighed filter paper, dried, and re-weighed. Some of the sites had very low suspended solids concentrations resulting in non-detections. We changed the procedure and brought pre-weighed filters and filtration equipment into the field where sufficient water was passed through the filters to obtain noticeable plugging. Particulate Organic Carbon (POC) sampling was similar. For POC, we tried to collect and filter water continuously over the span of time that TOPS was operated. Both SS and POC samples were kept frozen before being sent for analysis. The POC filters were also cored. David Hirschberg, the POC/DOC expert at SUNY Stony Brook, asked us to send him 10 mm disks. Initially, we sent him one 10 mm disk for each sample but later we sent three. This was due to apparent inhomogeneity of deposition on the filters. Dissolved organic carbon (DOC) was field filtered and acidified. Details of all three procedures appear in the TOPS SOP.

POC was measured on a Carlo Erba EA1108 CNS Analyzer. DOC was quantitated on Shimadzu TOC-5000, a high temperature oxidation type instrument. POC accuracy was assessed through use of a variety of internal and NIST standard reference materials. DOC analyses are intercalibrated through the use of internationally distributed intercalibration materials supplied by Dr. Jon Sharp at the University of Delaware and supported by the National Science Foundation.

Determination of suspended sediments, POC, and DOC at the USGS stations was more complicated. Large numbers of samples were taken across a changing hydrograph. These observations were combined to yield loads. We took the loads and back-calculated concentrations. The methods for combining the observations are described in Potterfield

(1972) - Techniques of Water Resources Investigations (TWRI) of the United States Geological Survey, Book 3, Chapter C3 - Computation of Fluvial-Sediment Discharge. The TWRI's are a series of "How To" publications the USGS puts out on a wide variety of things done on a regular basis. All survey offices computing suspended sediment loads use the same method - how they perform the computation (by hand, spreadsheet, USGS daily loads software, etc.) may vary based on the available data set.

Steps for calculating suspended sediment loads:

- 1 - Compile the discharge data
 - 1a) review and correct daily values
 - 1b) determine which 15 minute and hourly raw values are good/bad (USGS doesn't correct raw 15 min data, only the daily data).
- 2 - Compile the concentration data
 - 2a) review QC samples
 - 2b) track down missing data and any discrepancies.
- 3 - Determine a "box coefficient"
 - 3a) plot equal width vs. point samples
 - 3b) determine if coefficient is flow or seasonally dependent - other factors are possible.
- 4 - Apply the coefficient to the data.
- 5 - Generate a plot of concentration vs. discharge (and/or other parameters) for possible use in estimating periods with no record.
- 6 - Plot the adjusted data on a trace of stage and/or discharge.
- 7 - Generate a continuous concentration curve - this involves some art and a feel for how the constituent behaves at the site, it's basically an educated guess as to what the concentration was between samples. Obviously, the quality of the curve is highly dependant on the sample frequency - during TOPS event we were collecting a lot of samples, so the quality of the concentration curve is pretty good during these periods.
- 8 - Determine if the day should be "subdivided" - i.e. if the concentration curve and or the discharge changes dramatically over the course of the day ("dramatically" is more clearly defined in Potterfield (1972),
 - 8a) subdivision involves segmenting the day into smaller, more uniform parts, computing the load for those periods, and summing the periods for the day,
 - 8b) if required, the day was broken down into hour segments and the "mid-interval" method described by Potterfield (1972) was used.
- 9 - If the day doesn't need to be subdivided, determine the average concentration for the day from the continuous concentration curve.

For POC and DOC many of the same steps used in the sediment computation apply - here are some of the highlights:

POC

- 1 - Apply sediment box coefficient to POC data.
- 2 - For all 3 sites, reasonable relations exist between POC and suspended sediment these

- are used to help estimate POC concentrations during periods with no data.
- 3 - Plot adjusted POC, model POC, and anything else that might help determine the continuous concentration curve on a trace of discharge and or stage.
- 4 - Computations are identical to sediment - including subdividing days.

DOC

- 1 - The variability of DOC between samples is generally small, so linear interpolation between samples (computation of a noon value) was used to estimate many of the daily values.
- 2 - Plot DOC, model DOC (interpolated values), and anything else that might help determine the continuous concentration curve on a trace of discharge and or stage;
 - 2a - if the model wasn't responding to the discharge (for example – samples bracketing a discharge event) the concentration curve was adjusted by hand based on knowledge of how the constituent behaves at the site.
- 3 - Subdivision was rarely necessary because of the low variability in DOC concentrations, otherwise the computations were identical to POC and sediment.

QUALITY CONTROL

Data generated by the CARP are initially loaded into a data management system operated by Battelle Ocean Sciences of Duxbury, MA under the direction of Tom Gulbransen. Battelle screens incoming data for conformity to the rules of the Electronic Data Deliverable and admits those that are properly formatted. A second step of validation is being performed by Booze Allan Hamilton, a contractor to the Hudson River Foundation. The validation will consist of passing all data through a screen to determine compliance with certain QC parameters. At this time, none of the data have been “validated”. “Validated” data will supplant unvalidated data. NYSDEC does not normally use third part validation for its regulatory or enforcement work.

The amount of data collected by CARP is enormous; the water portion alone has produced a quarter of a million records. Many interested users will want to ask questions of the data that we have not anticipated. Others may want to have a simpler contact with the project and will be satisfied to have us paint a broad picture of the findings. The first group of users will go to our website (www.carpweb.org). This site includes maps, a variety of data screens, and a metadata document that explains the structure of the database.

Quality Control – Field QC

Previous experience has indicated that there may be significant lab to lab differences in measurement of trace organic chemicals. In CARP, three organics laboratories produced data for the NYSDEC water program. In order to determine the degree of inter-laboratory variability, a 5-gallon sediment sample was taken from the Arthur Kill near the northern side of Pralls Island at the beginning of the project. This material was

thoroughly homogenized and divided into vials that were kept frozen. From time to time, these vials were sent in to the labs and analyzed for the suite of CARP substances. Ultimately, Axys (AAS) received 26, Severn Trent (QTS) saw 10, and Wright State University (WSU) got 4. Each lab should have been receiving essentially the same material and should have been producing the same results. The actual results (average/standard deviation) are indicated below graphically (Figures 16-21).

Table 17. Interlab variability, Pralls Island sediments.

| | AAS | QTS | WSU |
|--------------------------|------------|------------|-----------|
| Dieldrin, ug/kg | 5/1.3 | 2.3/1.1 | 10/4.4 |
| Dioxin/Furans, ug/kg TEQ | 0.18/0.076 | 0.05/0.014 | 0.14/0.12 |
| PAH, umoles/kg | 91/28 | 24/2.8 | 55/48 |
| PCB, ug/kg | 1700/380 | 880/94 | 1100/160 |
| Total Chlordane, ug/kg | 31/9 | 10/4.4 | 50/46 |
| Total DDT, ug/kg | 980/400 | 840/1200 | 1000/340 |

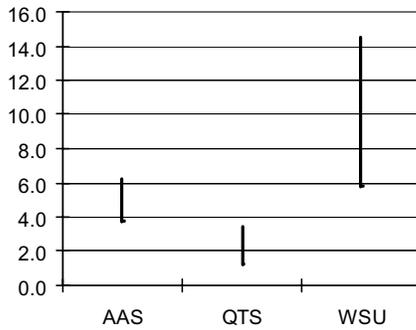


Figure 16. Field QC comparisons, Dieldrin in ug/kg. Mean +/- one standard deviation.

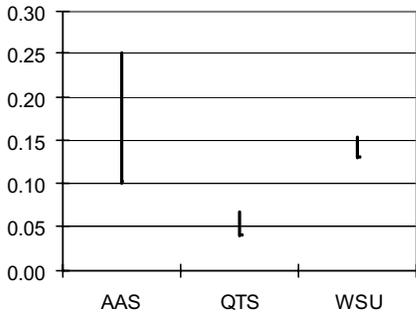


Figure 17. Field QC comparisons, TEQ dioxin/furan in ng/kg, mean +/- one standard deviation.

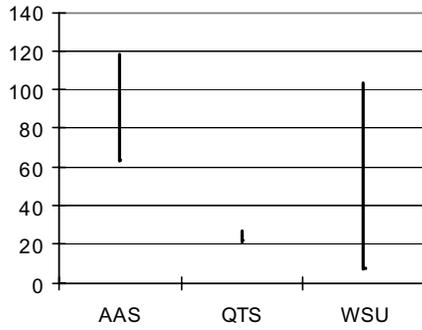


Figure 18. Field QC comparisons, PAHs in umoles/kg. Mean +/- one standard deviation.

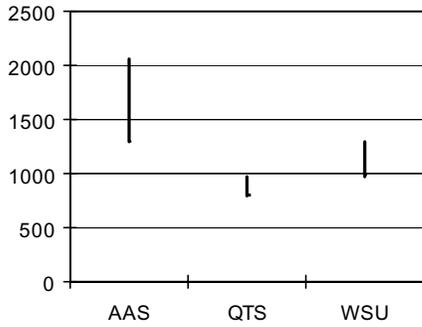


Figure 19. Field QC comparisons, PCBs in ug/kg. Mean +/- one standard deviation.

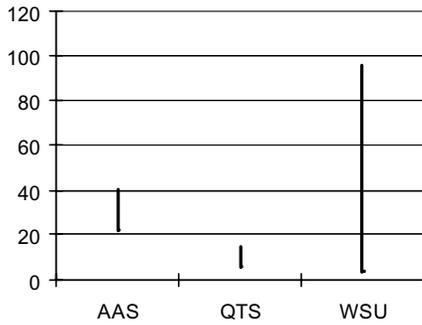


Figure 20. Field QC comparisons. Total chlordane in ug/kg. Mean +/- one standard deviation.

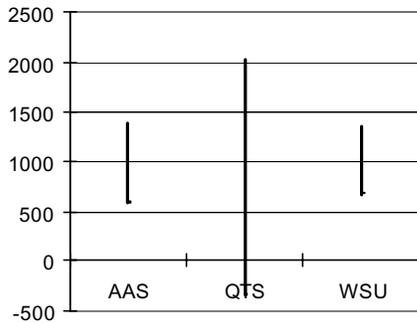


Figure 21. Field QC comparisons. Total DDT in ug/kg. Mean +/- one standard deviation.

Quality Control – Field Blanks and Equipment Blanks

A variety of experiments were performed to assess inadvertent contamination of the samples. The experiments fall into two broad groups, Field Blanks (FB) and Equipment Blanks (EB). FB are samples of media, glass fiber filters or XAD columns that are brought into the field unused, and returned to the lab. EB are samples of filters, XAD, or water that had been run through TOPS after routine cleaning.

The results of the blank experiments should be compared with the samples on the basis of total mass of recovered analyte rather than concentration.

- 1) unfired filter
- 2) processed after lower East River cruise
- 3) 96 L of Colonie, New York tap water transported to a sampling site on Staten Island and back to Colonie, New York and then processed via TOPS-Next Generation.

The significance of the field and equipment blank values is in how they measure against actual sample observations. The relevant units from the blanks are mass, not concentration. Therefore, the comparison with the sample observations must be the mass of analyte recovered from the medium.

Table 18 summarizes the results of field and equipment blanks. Table 18. Field and equipment blanks.

| Samp ID | Field Code | Lab ID | Medium | dieldrin ng | diox-F pg | PAH TEQ | PCB nmoles | T Chlor. ng | DDT ng |
|------------------------|------------|--------|-----------------------|----------------|--------------|------------|---------------|----------------|-----------|
| 1SPL00015 | FB | AAS | filt. Water | | | 0.1 | | | |
| 1SPL00521 | FB | AAS | glass fiber cart. | <0.32 | 1.57 | 1.4 | 1.3 | 0.2 | 0.24 |
| 1SPL00588 | FB | AAS | reagent water | | | 0.24 | | | |
| 1SPL00595 | FB | AAS | glass fiber cart. | <0.37 | 2.1 | 0.75 | 2.6 | <3.8 | <11 |
| 1SPL00596 | FB | AAS | XAD | <0.59 | 1.6 | | 25 | <4.5 | <14 |
| 1SPL00944 | FB | AAS | XAD | <0.82 | | 0.12 | 3.7 | <3.4 | <3.6 |
| 1SPL01625 ¹ | FB | AAS | glass fiber cart. | 0.13 | 11 | 72 | 74 | 0.91 | 0.27 |
| 1SPL01780 | FB | AAS | XAD | <0.22 | | | 1.4 | <0.41 | 0.48 |
| 1SPL01781 | FB | AAS | glass fiber cart. | <0.096 | 8 | 1 | 1.5 | <0.35 | <1.7 |
| 1SPL01814 | FB | AAS | glass fiber cart. | <0.077 | 7 | 0.67 | 8.3 | <0.26 | <1.2 |
| 1SPL01815 | FB | AAS | XAD | <0.15 | | | 1.6 | 0.38 | 3.4 |
| 1SPL01858 | FB | AAS | XAD | <0.083 | | | 0.83 | 0.33 | <6.1 |
| 1SPL01859 | FB | AAS | glass fiber cart. | <0.16 | 0.24 | | 0.31 | <0.76 | <4.8 |
| 1SPL01896 | EB | AAS | filt. water (AE,GF/F) | | 0.3 | | 0.17 | | |
| 1SPL01897 | EB | AAS | filt. AE+GF/F | | | | 0.58 | | |
| 1SPL02123 | FB | WSU | XAD | <4.6 | | | 6.4 | <12 | <6.2 |
| 1SPL02124 | FB | WSU | glass fiber cart. | <2.7 | 7.1 | 3.6 | 15 | <11 | 280 |
| 1SPL02145 ² | FB | WSU | glass fiber cart. | <3.1 | | 2.6 | 15 | <16 | 15 |
| 1SPL02159 ² | FB | WSU | XAD | <3.1 | | | 11 | <8.3 | <5.4 |

Table 18 continued.

| Samp_ID | Field Code | Lab_ID | Medium | Dieldrin ng | Diox-F pg TEQ | PAH nmoles | PCB ng | T Chlor. ng | TDDT ng |
|------------------------|------------|--------|-----------------------|----------------|------------------|---------------|-----------|----------------|------------|
| 1SPL02182 | EB | WSU | filt. water (AE,GF/F) | | | 0.48 | | | |
| 1SPL02183 | EB | WSU | filt. water (AE,GF/F) | | | 0.34 | | | |
| 1SPL02193 | EB | WSU | filt. water (AE,GF/F) | | | 0.42 | | | |
| 1SPL02194 | EB | WSU | filt. water (AE,GF/F) | | | 0.47 | | | |
| 1SPL02210 | EB | WSU | XAD | <1.6 | | | 13 | <10 | <5.1 |
| 1SPL02212 | EB | WSU | glass fiber cart. | <5.3 | 28 | 5.9 | 11 | <21 | <6.1 |
| 1SPL02285 | EB | WSU | filt. water (AE,GF/F) | | | 0.87 | | | |
| 1SPL02294 | EB | WSU | filt. water (AE,GF/F) | | | 1.3 | | | |
| 1SPL02295 | EB | WSU | filt. water (AE,GF/F) | | | 1.3 | | | |
| 1SPL02309 | EB | WSU | glass fiber cart. | <3.5 | 0.26 | 5.7 | 13 | <15 | <7.8 |
| 1SPL02310 | EB | WSU | XAD | <2.5 | | | 9.1 | <17 | <6.8 |
| 1SPL02317 | EB | WSU | filt. water (AE,GF/F) | | | 0.26 | | | |
| 1SPL02318 | EB | WSU | filt. water (AE,GF/F) | | | 0.19 | | | |
| 1SPL02373 | EB | AAS | filt. water, Post XAD | <0.0045 | 1.5 | 1.2 | 11 | <0.031 | <180 |
| 1SPL02374 | EB | AAS | XAD | 0.036 | | | 4.8 | 0.062 | <1.5 |
| 1SPL02375 | EB | AAS | glass fiber cart. | 0.046 | 1.7 | 0.79 | 1.4 | <0.37 | <2.2 |
| 1SPL02577 | FB | AAS | glass fiber cart. | <0.030 | 2.9 | 1.4 | 0.78 | <0.41 | <3.1 |
| 1SPL02578 | FB | AAS | XAD | <0.019 | | | 0.8 | | |
| 1SPL02621 | FB | AAS | XAD | <0.0041 | | | 1.6 | <1.1 | <2.7 |
| 1SPL02622 | FB | AAS | glass fiber cart. | <0.057 | 4 | 1.3 | 0.8 | <0.34 | <2.6 |
| 1SPL02820 ³ | EB | AAS | XAD | <0.056 | | | 5.6 | | |
| 1SPL02821 ³ | EB | AAS | XAD | 2.6 | | | 68 | | |
| 1SPL02873 ³ | EB | AAS | glass fiber cart. | <0.086 | | | 0.67 | | |
| 1SPL02953 | EB | AAS | reagent water | | 12 | | 1.2 | | |
| 1SPL03012 | EB | AAS | XAD | <0.32 | 5.6 | | 48 | <2.3 | <7.9 |
| 1SPL03013 | EB | AAS | glass fiber cart. | <0.19 | 7.3 | 6.7 | 1.4 | <2.1 | <8.1 |

PCB Blanks

In the case of PCBs, the highest blank value for XAD was 68 ng. The lowest sample mass from either a primary XAD column (first in the series) or a combined XAD (both columns extracted and analyzed together) was 80 ng and the 1th percentile (99 percent of observations were greater) was 100 ng. That 68 ng blank was from a sample of 96 L of tap water that had been put in large glass carboys, driven to Staten Island and back, and then processed in the lab. Had it been treated as a sample, the resultant concentration would have been 0.7 ng/L. The blank XAD with the next lowest PCB mass came from a

column with visible discoloration and had last been used to process landfill leachate. The homologue distribution of that blank does not look like a commercial PCB.

The highest levels of PCB found in a filter cartridge (74 ng) occurred from one that had not been fired in a furnace for 4 hours at 450°C. All glass fiber filter media used in the project for samples had been fired. The next highest blank filter PCB value was 15 ng. This value does begin to impinge on the data.

Dioxin/Furan Blanks

Dioxins/furans were measured mostly from glass fiber cartridges. The maximum blank was 28 pg TEQ and the next highest value (11 pg TEQ) was from the unfired filter. The average glass fiber filter blank was 6.2 pg TEQ. These blanks impinge on sample data; the 25th percentile for glass fiber cartridges was 23 pg TEQ and the 10th percentile was 10 pg TEQ.

Data from 99 dioxin/furan analyses on XADs (first or combined) had a median value of 3.4 pg. The blanks had a mean of 3.6 pg TEQ.

PAH Blanks

The average blank for PAHs from cartridges expressed as summation of moles, (ignoring the unfiltered cartridge with its 72 nmoles) was 2.7 nmoles. This is compared against the 5th percentile value for samples of 6.9 nmoles. The maximum cartridge blank contamination level was 6.7 nmoles.

The average blank value for water analyzed for PAHs was 0.63 nmoles. The 30th percentile value for the filtered water samples was 0.6 nmoles. Dissolved PAHs had no field concentration due to the potential of contamination by XAD resin.

Total DDT Blanks

Total DDT (2,4'-DDT, DDD, and DDE and 4,4'-DDT, DDD, and DDE) samples were much like the PCBs where they were measured from cartridges and XAD. Ten cartridge blanks had total DDT masses below the detection limit (max detection limit was 8.1 ng). Four samples had quantifiable TDDTs and ranged from 280 to 0.24 ng. The 1th percentile TDDT mass in the samples was 3.6 ng. There is no ready explanation for the very high blank (280 ng); the next highest, 15 ng, was from the unbaked filter. It is possible that the large value was a decimal error but there were values for five of the six analytes.

The 5th percentile of the first and combined XAD columns for TDDT was 2.8 ng. Ten XAD blanks were non-detect and two were quantitated values of 3.4 and 0.48 ng.

Dieldrin Blanks

Dieldrin was non-detect on 13 blank cartridges and detected twice at 0.046 and 0.13 ng. The 1th percentile for dieldrin on samples was 0.37 ng.

Blanks for dieldrin were non-detect from 13 XADs and quantitated on two at 0.036 and 2.6 ng. Of the first and combined sample XAD columns, the 5th percentile was 4.6 ng.

Total Chlordane Blanks

Total chlordane (alpha-, trans, and oxy-chlordane) had non-detect blanks on 12 cartridges and detected values in two of 0.2 and 0.91 ng. The 5th percentile for total chlordane on cartridges was 2.3 ng.

Nine XAD blanks had no measurable chlordane. Three had values ranging from 0.38 to 0.062 ng. The 1th percentile for total chlordane in first or combined sample XADs was 0.77 ng.

SAMPLE RESULTS

Sample results are shown below for PCBs, dioxins/furans, pesticides, PAHs, metals, and the accessory parameters SS/DOC/POC.

PCBs

PCBs are treated here as the summation of homologues. Data were evaluated against laboratory blanks and sample specific detection limits. Data collected by TOPS were also adjusted by factors derived from research more completely discussed in the paper, XAD in the Real World.

PCB Data Quality

PCB homologue data were evaluated against two tests;

Is the sum of the analyte masses in a sample 5 times greater than the sum of the analyte masses in its associated SDG method blank?

Is the sum of the analyte masses in a sample exceeded by the 10 times the sum of the sample specific detection limits from that sample?

Table 19 shows the success of the sampling program in obtaining adequate PCB samples. Attaining adequate data is a function of the amount of chemical present, the size of the sample, the laboratory detection level, and the lab's cleanliness. Samples from Severn-Trent have not been formally accepted into the database and lack method blanks. These SDGs with missing blanks are designated "M MB". Samples where the analyte was not detected are "ND". Samples with high detection limit and high method blank are "Hi DL" and "Hi MB". Samples meeting the criteria are "USE DL" and "USE MB".

Table 19. Number of analyzed PCB homologs meeting certain data quality criteria.

| | ND | Hi DL, Hi Bk | Hi DL, M MB | Hi DL, USE MB | USE DL, Hi Bk | USE DL, M MB | USE DL, MB |
|---------|-----|--------------|-------------|---------------|---------------|--------------|------------|
| 1-mono | 36 | 42 | 27 | 121 | 61 | 12 | 472 |
| 2-di | 42 | 26 | 25 | 183 | 5 | 23 | 468 |
| 3-tri | 9 | 34 | 19 | 80 | 24 | 29 | 576 |
| 4-tetra | 0 | 25 | 8 | 117 | 23 | 43 | 555 |
| 5-penta | 7 | 47 | 32 | 116 | 27 | 15 | 527 |
| 6-Hexa | 6 | 73 | 38 | 118 | 19 | 9 | 508 |
| 7-Hepta | 22 | 68 | 34 | 176 | 10 | 7 | 454 |
| 8-Octa | 90 | 87 | 19 | 174 | 11 | 6 | 384 |
| 9-Nona | 220 | 23 | 17 | 181 | 4 | 7 | 319 |
| 10-Deca | 178 | 67 | 12 | 98 | 50 | 6 | 360 |

The overall success rate of analyses is shown in Table 20. Inadequate detection limits resulted in non-detections or in observed masses insufficiently separated from SPDLS in 35% of the homologues. Only 5% of homologues were problematic due to method

blanks being either high or missing. The major data quality problem was insufficient sample size relative to the detection limits available in HRGC/HRMS.

Table 21 gives the average volumes (in liters) of water passed through the glass fiber filter and the XAD columns by sample type. As apparent, a significant effort was made to avoid under sampling. The actual amount of water processed for PCBs was less than indicated. Extracts were split into 4 (XAD) or 5 (cartridge filter samples); one for reserve in case of accident, one for dioxins/furans, one for pesticides, one for PAHs (cartridge filters only) and one for PCBs.

Table 20. Percent of high quality PCB homologues.

| | |
|---------|-----|
| 1-mono | 61% |
| 2-di | 61% |
| 3-tri | 75% |
| 4-tetra | 72% |
| 5-penta | 68% |
| 6-Hexa | 66% |
| 7-Hepta | 59% |
| 8-Octa | 50% |
| 9-Nona | 41% |
| 10-Deca | 47% |

Table 21. Volumes of water (L) processed

| sample_type | cartridge filter | XAD |
|---------------------|------------------|-----|
| Ambient, clean | 3000 | 710 |
| Ambient, Hudson R. | 680 | 200 |
| Ambient, Kills | 750 | 170 |
| Ambient, Non Kills | 720 | 180 |
| CSO/SWO | 91 | 82 |
| Industrial effluent | 470 | 100 |
| Landfill leachate | | 70 |
| Major tributary | 770 | 210 |
| Minor tributary | 740 | 180 |
| WPCF | 330 | 120 |

While these statistics show a large number of homologues failing to meet the criteria for being good data, a comparison between PCB concentrations from sample sites (Table 22) using all data versus “high quality” data reveals little difference in most cases.

PCB Results

Table 22 shows average PCB concentrations (ng/L) by sites where homologues are screened for inclusion by method blank and detection limit exceedences (censored data) and where all data were used. In subsequent analyses, all the data were used. The highest concentrations for each set of observations are highlighted.

The three highest concentrations of PCB were seen from a leachate sample (1E-HMDC), final effluent from the Passaic Valley Sewerage Commissioners (PVSC) in Newark, NJ, and a wet weather raw sewage influent used to simulate CSO samples (26th Ward, High Side). The PVSC sample was dominated by a single homologue; actually, a single congener. The 26th Ward, High Side shows Aroclor 1260. 1E-HMDC shows evidence of multiple Aroclors (Figure 21).

Table 22. Average PCB concentrations at each site without and with quality censoring (ng/L).

| Sample | raw | censored | Censored/raw |
|--|--------|----------|--------------|
| Ambient-clean: Long Island Sound | 0.47 | 0.285 | 61% |
| Ambient-clean: New York Bight | 0.0732 | 0.0286 | 39% |
| Ambient-Hudson: Haverstraw Bay | 25.8 | 25 | 97% |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 92.4 | 91 | 98% |
| Ambient-Hudson: Hudson R. below Kingston | 19.6 | 19.4 | 99% |
| Ambient-Hudson: Hudson R. below Tappen Zee | 31.8 | 31.1 | 98% |
| Ambient-Hudson: Hudson R. South of Harlem R. | 14.5 | 14 | 97% |
| Ambient-Kills: Hackensack R., Mid-Tidal | 28.3 | 27.5 | 97% |
| Ambient-Kills: Hackensack R., Mouth | 13.3 | 12.3 | 93% |
| Ambient-Kills: Newark Bay | 9.54 | 8.86 | 93% |
| Ambient-Kills: Northern Arthur Kill | 15.9 | 14.9 | 94% |
| Ambient-Kills: Passaic R., Mouth, Bottom | 22.4 | 21.9 | 98% |
| Ambient-Kills: Passaic R., Mouth, Surface | 33.7 | 33.2 | 98% |
| Ambient-Kills: Passaic River, Mid-Tidal | 38.5 | 37.7 | 98% |
| Ambient-Non_Kills: Jamaica Bay | 1.45 | 0.601 | 41% |
| Ambient-Non_Kills: Lower Bay | 3.69 | 2.77 | 75% |
| Ambient-Non_Kills: Lower East R. | 11.6 | 10.3 | 89% |
| Ambient-Non_Kills: Raritan Bay | 3.82 | 3.04 | 79% |
| Ambient-Non_Kills: Upper Bay | 8.02 | 7.95 | 99% |
| Ambient-Non_Kills: Upper East R. | 4.35 | 4.09 | 94% |
| CSO: 26 th Ward, High Side | 3500 | 3500 | 100% |
| CSO: 26 th Ward, Low Side | 851 | 851 | 100% |
| CSO: Bowery Bay High Side | 297 | 297 | 100% |
| CSO: Bowery Bay Low Side | 10.2 | 10.2 | 100% |
| CSO: Coney Island Influent | 43.7 | 43.7 | 100% |
| CSO: Hunts Point Influent | 57.7 | 57.7 | 100% |
| CSO: Jamaica Influent | 65 | 65 | 100% |
| CSO: Manhattan Grit Chamber | 130 | 130 | 100% |
| CSO: Manhattan Pump Station | 153 | 153 | 100% |
| CSO: Newtown Creek Influent | 261 | 261 | 100% |
| CSO: North River Influent | 351 | 351 | 100% |
| CSO: Owls Head Influent | 65.8 | 65.8 | 100% |
| CSO: Port Richmond Influent | 561 | 561 | 100% |
| CSO: Red Hook Influent | 1310 | 1310 | 100% |
| CSO: SWO-Jamaica, Commercial | 47.6 | 47.6 | 100% |
| CSO: SWO-Jamaica, Industrial | 69.8 | 69.8 | 100% |

Table 22 (continued).

| Sample | raw | censored | censored/raw |
|--|-------|----------|--------------|
| Industrial effluent: Clean Waters of New York | 0.046 | 0 | 0% |
| Industrial effluent: Fresh Kills Landfill Plant Effluent | 16.6 | 16.5 | 100% |
| Landfill leachate: 1A-HMDC | 946 | 946 | 100% |
| Landfill leachate: 1D-HMDC | 91.2 | 91.2 | 100% |
| Landfill leachate: 1E-HMDC | 1490 | 1490 | 100% |
| Landfill leachate: Fresh Kills LF 3/4 | 97.3 | 97.3 | 100% |
| Landfill leachate: Fresh Kills LF, 6/7 Composite | 275 | 275 | 100% |
| Landfill leachate: Fresh Kills LF, 1/9 "B" | 186 | 186 | 100% |
| Landfill leachate: Fresh Kills LF, 1/9 "F" | 87 | 87 | 100% |
| Landfill leachate: Fresh Kills LF, 1/9 Composite | 545 | 545 | 100% |
| Landfill leachate: Pelham Bay Landfill Holding Tank | 9.03 | 11.9 | 132% |
| Major_tributary: Hudson R. (Pleasantdale) | 30.4 | 26.8 | 88% |
| Major_tributary: Mohawk R. (Cohoes) | 8.42 | 3.96 | 47% |
| Major_tributary: Walkill (New Paltz) | 1.82 | 1.31 | 72% |
| Minor_tributary: Bronx River | 4.52 | 4.35 | 96% |
| Minor_tributary: Gowanus Canal | 7.94 | 6.38 | 80% |
| Minor_tributary: Saw Mill River (Yonkers) | 4.73 | 4.5 | 95% |
| Trackdown: Mill Creek at Arthur Kill Rd | 2.97 | 2.97 | 100% |
| WPCF: 26th Ward | 40.7 | 40.6 | 100% |
| WPCF: Bowery Bay | 5.74 | 5.22 | 91% |
| WPCF: Coney Island | 2.25 | 1.93 | 86% |
| WPCF: Edgewater | 6.51 | 6.51 | 100% |
| WPCF: Hunts Point | 4.4 | 4.33 | 98% |
| WPCF: Jamaica | 5.5 | 5.13 | 93% |
| WPCF: Newtown Creek | 12.6 | 12.2 | 97% |
| WPCF: North River | 3.77 | 3.43 | 91% |
| WPCF: Oakwood Beach | 9.22 | 8.96 | 97% |
| WPCF: Owls Head | 3.41 | 3.11 | 91% |
| WPCF: Port Richmond | 137 | 137 | 100% |
| WPCF: Poughkeepsie, City | 15.3 | 15.3 | 100% |
| WPCF: PVSC | 334 | 334 | 100% |
| WPCF: Red Hook | 3.71 | 3.36 | 90% |
| WPCF: Rensselaer | 5.93 | 5.68 | 96% |
| WPCF: Rockaway | 4.44 | 4.1 | 92% |
| WPCF: Rockland County | 4.42 | 4.36 | 99% |
| WPCF: Tallman Island | 5.33 | 5.02 | 94% |
| WPCF: Wards Island | 2.39 | 2.23 | 93% |
| WPCF: Yonkers | 8.24 | 4.61 | 56% |

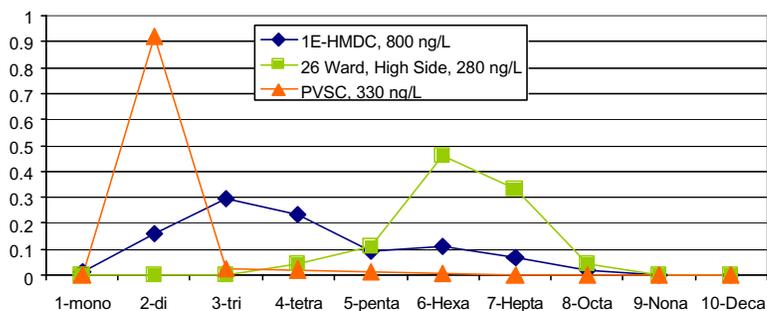


Figure 22. Relative abundances of PCB homologues from the three samples with the highest PCB concentrations.

Samples from tributaries, CSOs, and WPCFs have associated discharges that can be multiplied with concentration to obtain loads. Table 23 shows the average loads (g/hour) from the CSOs, tributaries, and WPCFs. The loads shown here are averages of the observed events and are not attempts to compute yearly loads encompassing unsampled times. The tributary loads are biased in that the samples were generally taken during hydrological events. Samples from wastewater treatment plants and minor tributaries were taken during different seasons but were not specifically intended to reflect wet or dry days. Samples called “CSO” were wet weather influents to treatment plants.

Table 23 shows four dominant PCB sources, the Hudson and Mohawk Rivers, the PVSC wastewater treatment plant, and 26th Ward CSOs. PVSC, the Newark, NJ treatment plant, was sampled once as part of a bi-state inter-comparison program. That program also involved DEC sampling at Edgewater, NJ.

The upper Hudson PCB source is well known to be from General Electric’s manufacturing of capacitors at Hudson Falls and Fort Edward. PCBs in the capacitor facility varied over the years but the largest type was Aroclor 1242. This is apparent in the homologue fingerprints from the top four loading events (Figure 23)

Table 23. Average loads (g/hr) from tributaries and point sources.

| Name | PCB load in g/hr |
|---|------------------|
| CSO-26 th Ward | 4.3 |
| CSO-Red Hook Influent | 0.76 |
| CSO-Newtown Creek Influent | 0.45 |
| CSO-Bowery Bay | 0.42 |
| CSO-Jamaica Influent | 0.32 |
| CSO-North River Influent | 0.28 |
| CSO-Wards Island | 0.22 |
| CSO-Hunts Point Influent | 0.14 |
| CSO-Owls Head Influent | 0.097 |
| CSO-Port Richmond Influent | 0.092 |
| CSO-Coney Island Influent | 0.069 |
| INDEF-Fresh Kills Landfill Plant Effluent | 0.0017 |
| Major_TRIB-Hudson R. (Pleasantdale) | 63 |
| Major_TRIB-Mohawk R. (Cohoes) | 33 |
| Major_TRIB-Walkkill (New Paltz) | 1.0 |
| Minor_TRIB-Bronx River | 0.033 |
| Minor_TRIB-Saw Mill River (Yonkers) | 0.02 |
| WPCF-PVSC | 17 |
| WPCF-Port Richmond | 0.96 |
| WPCF-Newtown Creek | 0.53 |
| WPCF-26th Ward | 0.5 |
| WPCF-Yonkers | 0.12 |
| WPCF-Bowery Bay | 0.11 |
| WPCF-Hunts Point | 0.1 |
| WPCF-North River | 0.093 |
| WPCF-Wards Island | 0.08 |
| WPCF-Jamaica | 0.075 |
| WPCF-Owls Head | 0.062 |
| WPCF-Tallman Island | 0.044 |
| WPCF-Oakwood Beach | 0.042 |
| WPCF-Coney Island | 0.035 |
| WPCF-Red Hook | 0.022 |
| WPCF-Rensselaer | 0.016 |
| WPCF-Rockaway | 0.015 |
| WPCF-Rockland County | 0.014 |
| WPCF-Poughkeepsie | 0.011 |
| WPCF-Edgewater | 0.0029 |

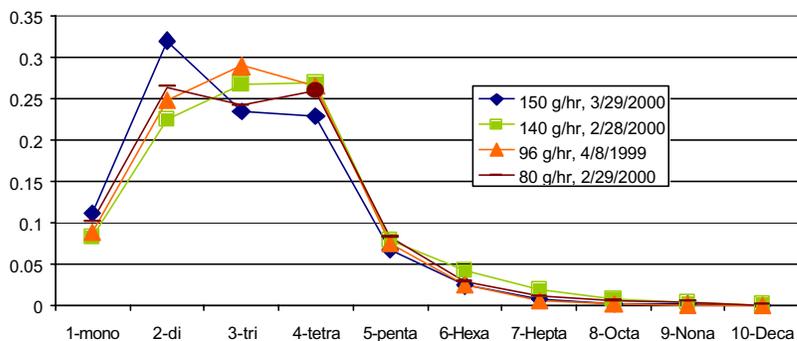


Figure 23. Relative abundances of PCB homologues from the top four loading events at Pleasantdale in the upper Hudson River.

These samples were deliberately biased toward large hydrological events to capture suspended sediments. Some researchers believe that a significant amount of PCB loading in the Hudson comes from biological effects that are at greatest intensity in the late spring or early summer. We do not have samples from that period and may have underestimated the load.

The apparent significance of the Mohawk River was investigated. Much of the weight of the average came from a single event, on February 2, 2000 where 260 g/hr were noted. The homologue pattern was indicative of Aroclor 1254. Pentachlorobiphenyl congeners are abundant in most Mohawk samples and were the most abundant group in seven out of 11 samples. February 28, 2000 was the day with the greatest concentration (54 ng/L) and the day of the greatest discharge of those sampled (48,000 CFS).

The fourth largest load was the 26th Ward CSO. This is due to Aroclor 1260 found in the sewers. High concentrations of Aroclor 1260 were seen in two separate wet weather influents as well as in PISCES and grab samples taken from the service area. Specific sources have not been discovered in this formerly industrial area.

The Wallkill, the third major tributary, had relatively low PCB concentrations (1.9 ng/L) but a high volume of discharge.

The third and sixth largest sources are the PVSC and Port Richmond wastewater treatment plants. In the cases of both treatment plants (90% at PVSC and 93% at Port Richmond), the overwhelmingly dominant PCB congener is the inadvertently produced 3,3'-dichlorobiphenyl (IUPAC 11). This congener is not routinely measured because it is not associated with commercial PCBs and it is not thought to be particularly toxic. Figure 24 shows the average relative abundances of IUPAC 11 in ambient sites. While IUPAC 11 is the single most abundant congener in New York Harbor ambient waters, it does not strongly bioaccumulate.

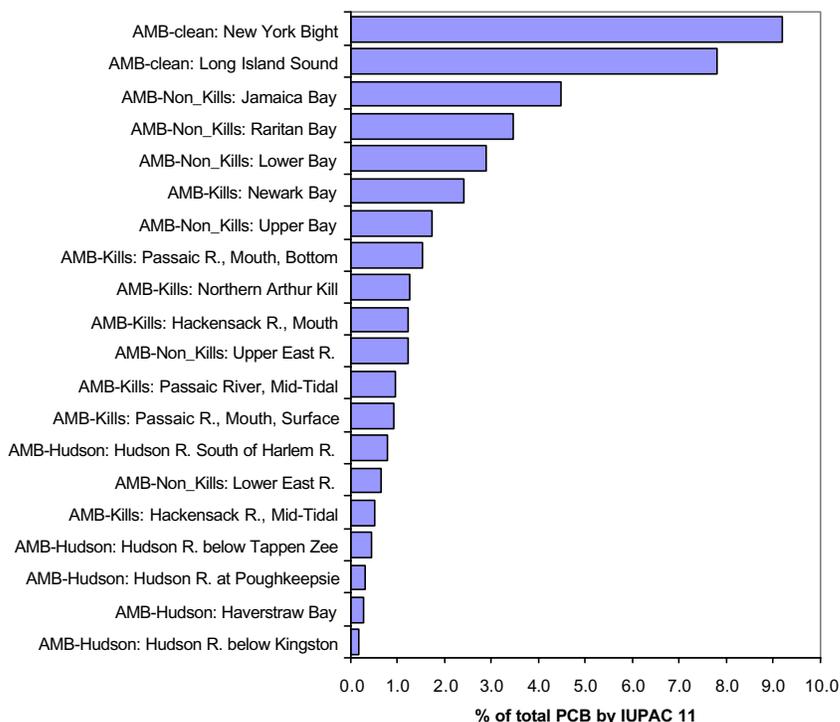


Figure 24. Percent relative abundances of IUPAC 11 in ambient water column sites. Area averages.

IUPAC 11 is produced in a small number of pigment factories in New York and New Jersey. The NYS facility has ceased using the IUPAC 11 precursor (3,3'-dichlorobenzidine) but inadvertently produced PCBs (principally IUPACs 11, 35, 77, and 126) continue to be emitted from the site. Discharge of inadvertently produced PCBs is covered by TSCA and the discharge from the facility falls far below the TSCA level. Elimination of these sources would be the easiest way to reduce total PCBs in New York Harbor water. But such a measure would have very little impact on PCBs in sediments or biota. An appended paper, Identification of a novel PCB source through analysis of 209 PCB congeners by US EPA Modified Method 1668, discussed the subject further.

1E-HMDC, a leachate sample from the New Jersey Meadowlands Commission, breaks out of mound 1E, flows through an area of fill and marsh, and enters the lower Passaic River. The total amount of flow is probably very small and its contribution to loading is probably insignificant. Most of the leachate from the HMDC is captured by PVSC and treated before being discharged into the Upper Bay.

Detailed discussions of PCB concentrations, homologue abundances, and, where appropriate, loads, appear below. The note "DU" indicates that a duplicate sample was taken. Duplicates and samples (SA) drew water from the same point over the same time.

Ambient Concentrations

Ambient samples were taken from 20 sites which, with two exceptions, were taken from slowly cruising boats. The samples were composited over as much of the area as was practical to go. Two sites, in mid-tidal Passaic and Hackensack Rivers, were taken from a bridge or from a dock. A 20th sample, Poughkeepsie, was also taken from a fixed location, the City of Poughkeepsie water treatment plant. This site was sampled primarily during periods of high flow. The other samples were taken to represent each of the four seasons. They were not specifically taken during high or low flow times.

Table 24. PCB concentrations and relative homologue abundances from samples composited between Kingston and Poughkeepsie.

| sample ng/L | 5/25/1999 - DU 23 | 5/25/1999 - SA 22 | 6/28/2000 - SA 16 | 10/7/1999 - SA 13 |
|----------------|----------------------|----------------------|----------------------|----------------------|
| 1-mono | 1.0% | 1.0% | 1.8% | 0.88% |
| 2-di | 17% | 17% | 20% | 15% |
| 3-tri | 40% | 38% | 32% | 38% |
| 4-tetra | 28% | 28% | 31% | 31% |
| 5-penta | 9.7% | 10% | 10% | 9.5% |
| 6-Hexa | 3.3% | 4.2% | 3.3% | 3.6% |
| 7-Hepta | 0.730% | 1.1% | 1.1% | 1.1% |
| 8-Octa | 0.22% | 0.35% | 0.38% | 0.32% |
| 9-Nona | 0.082% | 0.13% | 0.13% | 0.13% |
| 10-Deca | 0.028% | 0.053% | 0.088% | 0.068% |

USGS used TOPS to sample at Poughkeepsie in the Poughkeepsie water intake and found the highest average concentration, 92 ng/L. Most samples were taken during times of high flow at Waterford, New York. Table 25 shows for each of the Poughkeepsie samples the total PCB concentration and the percent abundance of each of the homologues. These concentrations are much greater than those found at Pleasantdale.

Table 25. PCB concentrations and relative homologue abundances from Poughkeepsie water intake samples.

| sample ng/L | 4/17/99 300 | 4/18/99 140 | 4/16/99 120 | 5/17/00 69 | 3/28/99 59 | 3/1/99 41 | 6/15/00 40 | 10/23/99 40 | 3/18/00 25 |
|----------------|----------------|----------------|----------------|---------------|---------------|--------------|---------------|----------------|---------------|
| 1-mono | 2.2% | 1.9% | 2.1% | 2.6% | 2.4% | 2.1% | 3.8% | 2.3% | 3.2% |
| 2-di | 16% | 16% | 16% | 23% | 17% | 18% | 17% | 21% | 19% |
| 3-tri | 35% | 35% | 36% | 26% | 36% | 36% | 30% | 35% | 34% |
| 4-tetra | 28% | 29% | 29% | 32% | 28% | 27% | 28% | 28% | 24% |
| 5-penta | 10% | 11% | 10% | 8.8% | 10% | 10% | 9.0% | 8.5% | 11% |
| 6-Hexa | 5.2% | 4.3% | 4.9% | 5.7% | 4.8% | 4.3% | 11% | 3.6% | 5.2% |
| 7-Hepta | 1.6% | 2.1% | 1.5% | 1.3% | 1.5% | 1.3% | 1.2% | 1.2% | 1.9% |
| 8-Octa | 0.61% | 0.87% | 0.65% | 0.32% | 0.59% | 0.58% | 0.28% | 0.36% | 0.64% |
| 9-Nona | 0.22% | 0.22% | 0.20% | 0.16% | 0.23% | 0.19% | 0.12% | 0.10% | 0.30% |
| 10-Deca | 0.093% | 0.076% | 0.089% | 0.053% | 0.10% | 0.092% | 0.045% | 0.059% | 0.12% |

Table 26. PCB concentrations and relative homologue abundances on samples collected from continuous tows between the Tappen Zee Bridge and Bear Mountain Bridge.

| Sample ng/L | 7/11/1999 34 | 2/10/1999 31 | 11/24/1998 19 | 4/4/2000 12 |
|----------------|-----------------|-----------------|------------------|----------------|
| 1-mono | 0.61% | 4.6% | 0.75% | 1.4% |
| 2-di | 10% | 24% | 11% | 16% |
| 3-tri | 32% | 31% | 36% | 34% |
| 4-tetra | 37% | 26% | 32% | 28% |
| 5-penta | 13% | 8.8% | 13% | 12% |
| 6-Hexa | 5.0% | 4.1% | 5.3% | 6.5% |
| 7-Hepta | 1.5% | 1.3% | 2.1% | 1.4% |
| 8-Octa | 0.56% | 0.45% | 0.68% | 0.33% |
| 9-Nona | 0.19% | 0.15% | 0.22% | 0.13% |
| 10-Deca | 0.12% | 0.074% | 0.12% | 0.063% |

Table 27. PCB concentrations and relative homologue abundances from the Hudson River between Tappen Zee Bridge and the Harlem River.

| sample ng/L | 2/19/99 65 | 4/4/00 23 | 12/1/98 18 | 7/10/99 16 |
|----------------|---------------|--------------|---------------|---------------|
| 1-mono | 2.0% | 0.96% | 0.81% | 0.59% |
| 2-di | 17% | 15% | 11% | 9.0% |
| 3-tri | 35% | 34% | 33% | 29% |
| 4-tetra | 27% | 29% | 28% | 34% |
| 5-penta | 9.9% | 12% | 14% | 17% |
| 6-Hexa | 5.6% | 6.2% | 7.1% | 7.3% |
| 7-Hepta | 2.0% | 1.6% | 3.2% | 2.5% |
| 8-Octa | 0.94% | 0.41% | 1.2% | 1.0% |
| 9-Nona | 0.30% | 0.19% | 0.51% | 0.31% |
| 10-Deca | 0.14% | 0.09% | 0.28% | 0.20% |

The trend toward lower concentrations but heavier congeners continues in the set of samples taken in cruises off Manhattan between the Battery and the Harlem River. Tides may play a role in congener abundances. Table 28 shows the cosine tides. See Table 6 for the cosine tides associated with all the ambient samples.

Table 28. Hudson River between the Harlem River and the Battery. Cosine tide is positive during ebb tide and negative during flood tide.

| sample | 3/16/99 | 12/17/98 | 6/14/00 | 10/5/01 | 8/12/99 | 12/14/99 | 8/12/99 | 12/14/99 |
|-------------|---------|----------|---------|---------|---------|----------|---------|----------|
| field QC | SA | SA | SA | SA | SA | DU | DU | SA |
| cosine tide | -1.1 | 0.47 | 1.1 | 1.4 | 1.9 | 1.9 | 1.9 | 1.9 |
| ng/L | 31 | 19 | 14 | 12 | 12 | 8.3 | 7.5 | 5.5 |
| 1-mono | 0.89% | 0.75% | 0.93% | 0.66% | 0.67% | 0.70% | 0.85% | 0.67% |
| 2-di | 12% | 12% | 12% | 10% | 8.8% | 11% | 10% | 12% |
| 3-tri | 33% | 31% | 31% | 30% | 30% | 31% | 29% | 30% |
| 4-tetra | 30% | 31% | 33% | 32% | 34% | 33% | 28% | 33% |
| 5-penta | 13% | 12% | 13% | 14% | 15% | 14% | 16% | 15% |
| 6-Hexa | 6.8% | 8.2% | 7.0% | 8.5% | 7.8% | 7.0% | 9.5% | 6.7% |
| 7-Hepta | 2.6% | 3.3% | 1.8% | 3.2% | 3.0% | 2.5% | 4.2% | 2.2% |
| 8-Octa | 1.1% | 1.5% | 0.66% | 1.0% | 0.92% | 1.0% | 1.4% | 0.77% |
| 9-Nona | 0.35% | 0.48% | 0.29% | 0.41% | 0.37% | 0.29% | 0.56% | 0.26% |
| 10-Deca | 0.18% | 0.32% | 0.17% | 0.22% | 0.25% | 0.17% | 0.38% | 0.13% |

Table 29. PCB concentrations and relative homologue abundances for samples taken on cruises around the Upper Bay.

| sample | 3/18/1999 - SA | 6/15/2000 - SA | 8/11/1999 - SA | 8/11/1999 - DU | 12/15/1998 - SA |
|---------|----------------|----------------|----------------|----------------|-----------------|
| ng/L | 12 | 7.9 | 7.8 | 7.2 | 5.4 |
| 1-mono | 0.80% | 0.79% | 0.60% | 0.77% | 0.64% |
| 2-di | 12% | 11% | 8.7% | 9.6% | 10% |
| 3-tri | 29% | 27% | 28% | 29% | 31% |
| 4-tetra | 28% | 30% | 30% | 30% | 29% |
| 5-penta | 16% | 14% | 17% | 16% | 15% |
| 6-Hexa | 8.3% | 14% | 9.4% | 8.8% | 8.3% |
| 7-Hepta | 3.6% | 2.4% | 3.4% | 3.2% | 3.4% |
| 8-Octa | 1.5% | 0.75% | 1.1% | 0.98% | 1.1% |
| 9-Nona | 0.51% | 0.24% | 0.37% | 0.34% | 0.43% |
| 10-Deca | 0.29% | 0.14% | 0.36% | 0.22% | 0.27% |

Table 30 shows concentrations and relative homologue abundances from samples taken on cruises between the Verrazano Narrows, the Sandy Hook-Rockaway line, and a line drawn south from Great Bay on Staten Island.

Table 30. PCB concentrations and homologue abundances in the Lower Bay.

| sample ng/L | 3/2/1999 - SA 3.3 | 6/1/2000 - DU 3.3 | 12/3/1998 – SA 3.1 | 6/1/2000 - SA 2.7 | 7/28/1999 - SA 1.8 |
|----------------|----------------------|----------------------|-----------------------|----------------------|-----------------------|
| 1-mono | 0.65% | 0.71% | 0.50% | 1.0% | 0.54% |
| 2-di | 13% | 9.2% | 10% | 10% | 12% |
| 3-tri | 27% | 21% | 28% | 21% | 28% |
| 4-tetra | 29% | 34% | 31% | 33% | 28% |
| 5-penta | 15% | 18% | 18% | 18% | 20% |
| 6-Hexa | 9.4% | 10% | 7.5% | 10% | 8.23% |
| 7-Hepta | 3.6% | 4.4% | 3.2% | 4.0% | 2.7% |
| 8-Octa | 1.3% | 1.0% | 1.1% | 1.0% | 0.79% |
| 9-Nona | 0.36% | 0.44% | 0.44% | 0.40% | 0.22% |
| 10-Deca | 0.21% | 0.42% | 0.25% | 0.30% | 0.15% |

Samples from the New York Bight were taken beyond the Sandy Hook – Rockaway line (Table 31). The sample taken on April 26, 1999 had a significant contribution from IUPAC 11 (20%) but usually Bight samples had heavier congeners than harbor samples.

Table 31. PCB concentrations and homologue abundances in New York Bight.

| sample field QC ng/L | 12/9/98 SA 0.27 | 2/1/99 SA 0.15 | 3/13/00 SA 0.11 | 4/26/99 SA 0.084 | 1/30/99 SA 0.077 | 1/29/99 DU 0.07 | 2/1/99 DU 0.068 | 1/30/99 DU 0.06 | 1/29/99 SA 0.059 |
|----------------------------|-----------------------|----------------------|-----------------------|------------------------|------------------------|-----------------------|-----------------------|-----------------------|------------------------|
| 1-mono | 0.54% | 3.1% | 5.6% | 6.7% | 5.0% | 8.6% | 6.8% | 7.0% | 7.1% |
| 2-di | 5.2% | 12% | 14% | 28% | 12% | 16% | 13% | 12% | 16% |
| 3-tri | 28% | 16% | 17% | 19% | 18% | 21% | 22% | 21% | 20% |
| 4-tetra | 34% | 26% | 20% | 22% | 19% | 22% | 20% | 21% | 24% |
| 5-penta | 17% | 24% | 25% | 15% | 31% | 17% | 22% | 25% | 19% |
| 6-Hexa | 12% | 15% | 15% | 7.1% | 11% | 12% | 12% | 12% | 11% |
| 7-Hepta | 3.4% | 3.1% | 3.2% | 1.3% | 2.4% | 2.7% | 3.0% | 1.7% | 2.7% |
| 8-Octa | 0.51% | 0.63% | 0.29% | 0.43% | 0.34% | 0.33% | 0.16% | 0.05% | 0.45% |
| 9-Nona | 0.043% | 0.068% | 0.084% | 0.12% | 0.054% | | 0.16% | | 0.10% |
| 10-Deca | 0.03% | 0.13% | 0.28% | 0.08% | 0.11% | 0.16% | 0.35% | 0.11% | 0.34% |

Table 32. PCB samples and relative homologue abundances in Long Island Sound.

| sample ng/L | 5/27/99 - SA 0.6 | 11/18/98 - SA 0.55 | 10/19/99 - SA 0.49 | 5/27/99 - DU 0.48 | 3/2/99 - SA 0.4 |
|----------------|---------------------|-----------------------|-----------------------|----------------------|--------------------|
| 1-mono | 1.8% | 0.44% | 0.71% | 2.3% | 3.3% |
| 2-di | 15% | 5.9% | 19% | 13% | 8.1% |
| 3-tri | 19% | 17% | 11% | 15% | 13% |
| 4-tetra | 26% | 26% | 27% | 22% | 20% |
| 5-penta | 18% | 26% | 19% | 21% | 21% |
| 6-Hexa | 13% | 15% | 14% | 17% | 20% |
| 7-Hepta | 4.7% | 6.0% | 5.4% | 7.0% | 8.9% |
| 8-Octa | 1.6% | 2.4% | 1.8% | 2.5% | 3.6% |
| 9-Nona | 0.48% | 0.68% | 0.58% | 0.67% | 0.89% |
| 10-Deca | 0.23% | 0.46% | 0.39% | 0.45% | 0.59% |

Table 33 shows sample results from cruises on the East River between Hell Gate and the Throgs Neck Bridge. Concentrations in the upper East River are much higher than those in central Long Island Sound and the congeners are lighter.

Table 33. Upper East River PCB concentrations and homologue abundances.

| Sample ng/L | 12/10/98 5.8 | 3/7/00 4.2 | 3/3/99 4.1 | 8/10/99 3.9 |
|----------------|-----------------|---------------|---------------|----------------|
| 1-mono | 0.46% | 1.8% | 0.88% | 0.49% |
| 2-di | 7.8% | 8.6% | 13% | 6.8% |
| 3-tri | 27% | 19% | 29% | 25% |
| 4-tetra | 28% | 30% | 26% | 29% |
| 5-penta | 19% | 18% | 14% | 19% |
| 6-Hexa | 10% | 14% | 11% | 12% |
| 7-Hepta | 4.8% | 6.6% | 3.9% | 4.6% |
| 8-Octa | 1.5% | 0.72% | 1.6% | 1.3% |
| 9-Nona | 0.63% | 0.35% | 0.52% | 0.55% |
| 10-Deca | 0.43% | 0.48% | 0.23% | 0.39% |

Table 34. Lower East River samples taken on cruises between Hell Gate and the Brooklyn Bridge, PCB concentrations and homologue abundances.

| Sample ng/L | 6/2/2000 - DU 13 | 9/18/1998 - SA 12 | 6/2/2000 - SA 9.3 | 3/11/1999 - SA 8 | 7/27/1999 - SA 7 |
|----------------|---------------------|----------------------|----------------------|---------------------|---------------------|
| 1-mono | 0.77% | 0.64% | 0.75% | 0.80% | 0.50% |
| 2-di | 7.7% | 7.2% | 7.5% | 12% | 7.1% |
| 3-tri | 19% | 23% | 21% | 31% | 25% |
| 4-tetra | 33% | 34% | 32% | 29% | 29% |
| 5-penta | 18% | 17% | 16% | 13% | 22% |
| 6-Hexa | 14% | 12% | 16% | 9.2% | 10% |
| 7-Hepta | 5.2% | 4.3% | 4.5% | 3.4% | 3.9% |
| 8-Octa | 1.0% | 1.0% | 0.83% | 1.2% | 1.5% |
| 9-Nona | 0.60% | 0.44% | 0.64% | 0.37% | 0.42% |
| 10-Deca | 0.48% | 0.31% | 0.54% | 0.24% | 0.32% |

Table 35. Jamaica Bay PCB concentrations and relative homologue abundances.

| sample ng/L | 2/23/99 2.3 | 7/9/99 1.7 | 5/4/00 1.4 | 10/14/98 0.74 |
|----------------|----------------|---------------|---------------|------------------|
| 1-mono | 0.68% | 1.3% | 0.46% | 0.61% |
| 2-di | 12% | 13% | 14% | 6.9% |
| 3-tri | 26% | 22% | 25% | 20% |
| 4-tetra | 25% | 28% | 29% | 25% |
| 5-penta | 18% | 20% | 18% | 22% |
| 6-Hexa | 12% | 11% | 12% | 15% |
| 7-Hepta | 5.2% | 3.5% | 1.8% | 6.9% |
| 8-Octa | 1.0% | 1.1% | | 2.0% |
| 9-Nona | 0.48% | 0.26% | | 0.68% |
| 10-Deca | 0.00% | 0.21% | | 0.39% |

Samples were taken at three sites in the Passaic River, cruises near the mouth at the surface; cruises near the mouth and 1 meter above the bottom, and from a bridge at Nutley, NJ.

Table 36. Passaic River, mouth surface PCB concentrations and homologue abundances.

| Sample ng/L | 6/17/99 87 | 6/27/00 21 | 2/3/99 14 | 11/13/98 11 |
|----------------|---------------|---------------|--------------|----------------|
| 1-mono | 0.28% | 0.23% | 0.38% | 0.21% |
| 2-di | 4.8% | 4.6% | 7.8% | 5.0% |
| 3-tri | 19% | 21% | 24% | 28% |
| 4-tetra | 29% | 35% | 27% | 40% |
| 5-penta | 21% | 18% | 16% | 14% |
| 6-Hexa | 15% | 13% | 14% | 8.0% |
| 7-Hepta | 7.0% | 6.4% | 7.7% | 3.2% |
| 8-Octa | 2.7% | 1.8% | 2.5% | 0.68% |
| 9-Nona | 0.69% | 0.37% | 0.52% | 0.19% |
| 10-Deca | 0.40% | 0.27% | 0.28% | 0.10% |

Table 37. Passaic River, mouth bottom PCB concentrations and homologue abundances.

| sample ng/L | 5/2/00 31 | 7/21/99 24 | 6/26/00 20 | 2/5/99 14 |
|----------------|--------------|---------------|---------------|--------------|
| 1-mono | 0.23% | 0.21% | 0.30% | 0.40% |
| 2-di | 5.5% | 4.3% | 4.3% | 12% |
| 3-tri | 20% | 21% | 18% | 26% |
| 4-tetra | 32% | 35% | 34% | 29% |
| 5-penta | 17% | 22% | 20% | 14% |
| 6-Hexa | 14% | 11% | 12% | 11% |
| 7-Hepta | 7.7% | 4.4% | 7.6% | 5.1% |
| 8-Octa | 2.1% | 1.4% | 2.4% | 1.8% |
| 9-Nona | 0.63% | 0.27% | 0.55% | 0.44% |
| 10-Deca | 0.26% | 0.17% | 0.31% | 0.25% |

Table 38. Passaic River, mid-tidal PCB concentrations and homologue abundances.

| Sample ng/L | 10/18/00 71 | 8/25/99 64 | 5/9/00 12 | 3/16/99 7.1 |
|----------------|----------------|---------------|--------------|----------------|
| 1-mono | 0.17% | 0.25% | 0.65% | 1.5% |
| 2-di | 3.1% | 4.2% | 6.5% | 8.9% |
| 3-tri | 15% | 21% | 20% | 17% |
| 4-tetra | 30% | 32% | 35% | 28% |
| 5-penta | 22% | 21% | 16% | 20% |
| 6-Hexa | 16% | 13% | 15% | 14% |
| 7-Hepta | 9.7% | 6.1% | 5.0% | 7.2% |
| 8-Octa | 3.0% | 1.9% | 0.98% | 2.4% |
| 9-Nona | 0.65% | 0.44% | 0.26% | 0.52% |
| 10-Deca | 0.36% | 0.24% | 0.17% | 0.23% |

The Hackensack River was sampled from cruises at its mouth onto Newark Bay (Hackensack, mouth) and from a dock at the foot of Plank Road (Hackensack mid-tidal).

Table 39. Hackensack River mouth, PCB concentrations and homologue abundances.

| sample ng/L | 11/12/98 11 | 2/8/99 8.1 | 7/7/99 25 | 4/11/00 6.8 |
|----------------|----------------|---------------|--------------|----------------|
| 1-mono | 0.17% | 0.35% | 1.2% | 0.33% |
| 2-di | 3.5% | 10% | 8.0% | 9.3% |
| 3-tri | 24% | 27% | 20% | 30% |
| 4-tetra | 44% | 33% | 35% | 42% |
| 5-penta | 16% | 14% | 19% | 13% |
| 6-Hexa | 8.4% | 8.7% | 10% | 4.4% |
| 7-Hepta | 3.1% | 4.0% | 4.2% | 1.3% |
| 8-Octa | 0.63% | 1.4% | 1.6% | 0.05% |
| 9-Nona | 0.19% | 0.31% | 0.38% | 0.00% |
| 10-Deca | 0.12% | 0.17% | 0.23% | 0.00% |

Table 40. Hackensack, mid-tidal PCB concentrations and homologue abundances.

| sample ng/L | 3/17/99 14 | 9/2/99 27 | 10/12/99 29 | 5/10/00 41 |
|----------------|---------------|--------------|----------------|---------------|
| 1-mono | 0.45% | 0.19% | 0.16% | 0.13% |
| 2-di | 5.1% | 3.8% | 3.5% | 3.8% |
| 3-tri | 24% | 27% | 20% | 20% |
| 4-tetra | 41% | 42% | 41% | 43% |
| 5-penta | 18% | 18% | 20% | 17% |
| 6-Hexa | 7.0% | 5.8% | 8.6% | 12% |
| 7-Hepta | 2.9% | 2.3% | 4.0% | 3.4% |
| 8-Octa | 0.95% | 0.57% | 1.4% | 0.80% |
| 9-Nona | 0.23% | 0.15% | 0.38% | 0.19% |
| 10-Deca | 0.13% | 0.09% | 0.23% | 0.11% |

Table 41. Newark Bay from Shooter's Island to the NJ Turnpike. PCB concentrations and homologue abundances.

| sample ng/L | 8/11/99 - SA 14 | 11/25/98 - SA 10 | 1/27/99 - SA 9.2 | 12/15/99 - SA 8.2 | 12/15/99 - DU 6.8 | 4/12/00 - SA 4.9 |
|----------------|--------------------|---------------------|---------------------|----------------------|----------------------|---------------------|
| 1-mono | 0.29% | 0.35% | 0.43% | 0.30% | 0.32% | 0.58% |
| 2-di | 5.8% | 8.3% | 11% | 8.2% | 8.7% | 9.3% |
| 3-tri | 25% | 27% | 27% | 24% | 25% | 34% |
| 4-tetra | 35% | 30% | 29% | 34% | 35% | 30% |
| 5-penta | 20% | 18% | 15% | 18% | 16% | 16% |
| 6-Hexa | 9.4% | 9.2% | 11% | 9.7% | 9.0% | 7.3% |
| 7-Hepta | 3.7% | 4.9% | 4.3% | 3.8% | 3.9% | 2.4% |
| 8-Octa | 1.1% | 1.5% | 1.6% | 1.4% | 1.4% | 0.32% |
| 9-Nona | 0.32% | 0.49% | 0.41% | 0.26% | 0.36% | 0.12% |
| 10-Deca | 0.24% | 0.30% | 0.24% | 0.16% | 0.20% | |

Table 42. Arthur Kill from the Goethals Bridge to the northern mouth of Fresh Kills, PCB concentrations and homologue abundances.

| sample ng/L | 11/17/1998 25 | 2/17/1999 18 | 7/8/1999 13 | 4/18/2000 7 |
|----------------|------------------|-----------------|----------------|----------------|
| 1-mono | 0.34% | 0.50% | 0.27% | 0.18% |
| 2-di | 5.3% | 3.6% | 7.1% | 7.2% |
| 3-tri | 20% | 16% | 18% | 22% |
| 4-tetra | 36% | 23% | 26% | 33% |
| 5-penta | 19% | 19% | 17% | 16% |
| 6-Hexa | 11% | 19% | 17% | 13% |
| 7-Hepta | 5.2% | 15% | 10% | 7.2% |
| 8-Octa | 1.8% | 3.9% | 3.2% | 1.3% |
| 9-Nona | 0.41% | 0.55% | 0.48% | 0.29% |
| 10-Deca | 0.34% | 0.30% | 0.28% | 0.16% |

Table 43. Raritan Bay west of a line dropped from Great Kills south, PCB concentrations and homologue abundances.

| sample ng/L | 7/12/1999 6.4 | 2/24/1999 4.2 | 5/3/2000 3.8 | 11/16/1998 2.4 |
|----------------|------------------|------------------|-----------------|-------------------|
| 1-mono | 0.36% | 0.28% | 0.25% | 0.31% |
| 2-di | 6.7% | 13% | 8.6% | 7.2% |
| 3-tri | 25% | 19% | 26% | 27% |
| 4-tetra | 40% | 27% | 35% | 31% |
| 5-penta | 17% | 19% | 16% | 16% |
| 6-Hexa | 6.9% | 13% | 11% | 10% |
| 7-Hepta | 2.2% | 5.9% | 2.6% | 5.1% |
| 8-Octa | 0.56% | 1.6% | 0.46% | 1.5% |
| 9-Nona | 0.17% | 0.39% | 0.24% | 0.45% |
| 10-Deca | 0.12% | 0.22% | 0.00% | 0.28% |

Tributaries

Two kinds of tributary samples were taken, major tributaries (Hudson and Pleasantdale, Mohawk at Cohoes, and Wallkill at New Paltz) and minor tributaries (Bronx River at Botanical Garden and below the Bronz Zoo, Saw Mill River in Yonkers, and the Gowanus Canal from the Carroll Street Bridge). The major tributaries were sampled by the USGS to capture high flow events and a few base flows. The minor tributaries were sampled seasonally. Discharges are shown as cubic feet per second (CFS).

Table 44. Hudson River at Pleasantdale PCB concentrations and homologue abundances.

| sample | 3/29/00 | 4/8/99 | 2/28/00 | 4/1/99 | 9/7/01 | 2/29/00 | 3/22/99 | 9/20/99 | 3/4/99 | 8/29/00 | 2/25/00 | 4/4/00 |
|---------|---------|--------|---------|--------|--------|---------|---------|---------|--------|---------|---------|--------|
| CFS | 26,600 | 19,000 | 33,600 | 21,100 | | 28,600 | 19,100 | 6,470 | 17,000 | 4,970 | 13,100 | 29,300 |
| ng/L | 56 | 50 | 41 | 38 | 32 | 28 | 25 | 20 | 20 | 18 | 14 | 11 |
| 1-mono | 11% | 8.8% | 8.4% | 7.6% | 5.0% | 10% | 7.8% | 8.6% | 7.9% | 8.0% | 17% | 11% |
| 2-di | 32% | 25% | 22% | 23% | 36% | 26% | 22% | 38% | 23% | 38% | 39% | 31% |
| 3-tri | 23% | 29% | 27% | 28% | 31% | 24% | 25% | 16% | 27% | 26% | 22% | 21% |
| 4-tetra | 23% | 26% | 27% | 28% | 20% | 26% | 27% | 26% | 23% | 22% | 15% | 27% |
| 5-penta | 6.9% | 7.6% | 7.9% | 9.2% | 5.4% | 8.3% | 12% | 6.8% | 11% | 4.3% | 4.1% | 6.8% |
| 6-Hexa | 2.5% | 2.5% | 4.3% | 3.2% | 1.6% | 2.9% | 4.8% | 2.6% | 5.1% | 1.3% | 2.3% | 2.5% |
| 7-Hepta | 0.74% | 0.64% | 2.0% | 0.82% | 0.39% | 1.0% | 1.2% | 1.1% | 1.4% | 0.19% | 0.93% | 0.77% |
| 8-Octa | 0.24% | 0.21% | 0.84% | 0.45% | 0.12% | 0.66% | 0.59% | 0.19% | 0.71% | 0.06% | 0.28% | 0.24% |
| 9-Nona | 0.19% | 0.08% | 0.46% | 0.22% | 0.05% | 0.49% | 0.26% | 0.02% | 0.28% | 0.06% | 0.12% | 0.14% |
| 10-Deca | 0.04% | 0.02% | 0.14% | 0.04% | 0.01% | 0.11% | 0.09% | 0.01% | 0.12% | 0.01% | 0.04% | 0.04% |

Table 45. Mohawk River at Cohoes PCB concentrations and homologue abundances.

| Sample | 2/28/00 | 4/4/00 | 3/4/99 | 4/1/99 | 9/17/99 | 2/26/00 | 3/12/00 | 3/28/00 |
|---------|---------|--------|--------|--------|---------|---------|---------|---------|
| CFS | 48,000 | 38,200 | 17,000 | 17,500 | 21,500 | 18,400 | 23,700 | 31,500 |
| ng/L | 54 | 9.9 | 6.7 | 6.1 | 5.1 | 2.2 | 2 | 1.7 |
| 1-mono | 0.25% | 7.50% | 0.19% | 1.20% | 0.10% | 0.00% | 1.30% | 0.42% |
| 2-di | 2.30% | 20% | 2.50% | 13% | 1.60% | 2.10% | 4.50% | 3.40% |
| 3-tri | 15% | 28% | 9.90% | 23% | 12% | 15% | 21% | 15% |
| 4-tetra | 24% | 27% | 19% | 27% | 23% | 27% | 21% | 24% |
| 5-penta | 31% | 10% | 38% | 20% | 29% | 28% | 24% | 29% |
| 6-Hexa | 19% | 4.90% | 21% | 11% | 22% | 18% | 17% | 19% |
| 7-Hepta | 6.30% | 1.60% | 6.70% | 3.00% | 8.80% | 7.30% | 7.00% | 6.50% |
| 8-Octa | 1.30% | 0.54% | 2.00% | 0.70% | 2.20% | 2.10% | 2.50% | 1.50% |
| 9-Nona | 0.28% | 0.29% | 0.35% | 0.14% | 0.50% | 0.35% | 1.20% | 0.76% |
| 10-Deca | 0.25% | 0.07% | 0.21% | 0.08% | 0.00% | 0.33% | 0.30% | 0.48% |

Table 46. Walkkill at New Paltz PCB concentrations and homologues.

| Sample | 3/21/01 | 3/30/01 | 9/17/99 | 5/26/01 | 2/15/00 | 6/29/01 | 6/17/01 | 10/13/99 |
|---------|---------|---------|---------|---------|---------|---------|---------|----------|
| CFS | 1,030 | 6,140 | 6,350 | 6,270 | 589 | 1,470 | 589 | 608 |
| ng/L | 4.4 | 3.6 | 3 | 1.6 | 1.2 | 0.53 | 0.47 | 0.26 |
| 1-mono | 0.16% | 0.18% | 0.00% | 0.21% | 0.00% | 0.42% | 0.51% | 0.47% |
| 2-di | 2.60% | 2.30% | 0.00% | 4.70% | 0.00% | 6.00% | 7.00% | 4.00% |
| 3-tri | 7.60% | 7.60% | 7.60% | 11% | 11% | 10% | 11% | 11% |
| 4-tetra | 17% | 16% | 18% | 17% | 16% | 18% | 18% | 18% |
| 5-penta | 31% | 32% | 29% | 27% | 31% | 29% | 29% | 29% |
| 6-Hexa | 22% | 24% | 24% | 22% | 31% | 22% | 22% | 21% |
| 7-Hepta | 9.90% | 11% | 13% | 9.40% | 8.80% | 8.50% | 6.80% | 10% |
| 8-Octa | 3.10% | 3.30% | 7.30% | 3.00% | 1.30% | 2.90% | 2.30% | 3.00% |
| 9-Nona | 1.30% | 0.96% | 1.40% | 1.20% | 0.00% | 1.00% | 0.87% | 0.90% |
| 10-Deca | 4.20% | 3.00% | 0.00% | 3.80% | 0.78% | 1.90% | 1.70% | 1.80% |

Table 47. Bronx River PCB concentrations and homologue abundances.

| sample | 3/8/99 | 7/27/99 | 10/29/98 | 10/26/99 |
|---------|--------|---------|----------|----------|
| ng/L | 5.3 | 5.2 | 4.7 | 2.9 |
| 1-mono | 0.13% | 0.16% | 0.17% | 0.36% |
| 2-di | 1.7% | 1.6% | 1.6% | 2.9% |
| 3-tri | 5.0% | 4.9% | 3.9% | 7.6% |
| 4-tetra | 11% | 14% | 14% | 18% |
| 5-penta | 17% | 21% | 16% | 19% |
| 6-Hexa | 35% | 31% | 35% | 27% |
| 7-Hepta | 23% | 22% | 24% | 19% |
| 8-Octa | 6.7% | 5.2% | 4.4% | 5.1% |
| 9-Nona | 0.66% | 0.53% | 0.44% | 0.52% |
| 10-Deca | 0.20% | 0.17% | 0.11% | 0.16% |

Table 48. Saw Mill River (at Yonkers) PCB concentrations and homologue abundances.

| sample | 11/10/1998 | 3/10/1999 | 5/5/1999 | 8/20/1999 |
|---------|------------|-----------|----------|-----------|
| ng/L | 3.2 | 1.9 | 11 | 4.3 |
| 1-mono | 0.25% | 0.40% | 1.00% | 0.34% |
| 2-di | 3.8% | 5.3% | 5.8% | 7.2% |
| 3-tri | 13% | 17% | 19% | 32% |
| 4-tetra | 23% | 24% | 28% | 24% |
| 5-penta | 29% | 24% | 27% | 21% |
| 6-Hexa | 21% | 19% | 14% | 11% |
| 7-Hepta | 8.5% | 7.0% | 3.6% | 3.5% |
| 8-Octa | 1.9% | 2.5% | 0.92% | 1.2% |
| 9-Nona | 0.40% | 0.46% | 0.16% | 0.18% |
| 10-Deca | 0.13% | 0.17% | 0.07% | 0.07% |

Table 49 shows PCB concentrations and homologue abundances from the Gowanus Canal at the Carroll Street Bridge.

Table 49. Gowanus Canal PCB concentrations and homologue abundances.

| Sample ng/L | 3/17/1999 11 | 8/24/1999 5.6 | 3/21/2000 3.2 | 9/28/2000 4.5 |
|----------------|-----------------|------------------|------------------|------------------|
| 1-mono | 0.96% | 0.66% | 2.0% | 1.0% |
| 2-di | 12% | 8.2% | 1.4% | 11% |
| 3-tri | 29% | 25% | 23% | 27% |
| 4-tetra | 29% | 30% | 21% | 27% |
| 5-penta | 15% | 18% | 26% | 16% |
| 6-Hexa | 9.1% | 12% | 18% | 11% |
| 7-Hepta | 3.5% | 4.3% | 7.7% | 4.9% |
| 8-Octa | 1.2% | 1.3% | 1.1% | 1.8% |
| 9-Nona | 0.32% | 0.46% | 0.20% | 0.71% |
| 10-Deca | 0.16% | 0.31% | 0.10% | 0.47% |

Wastewater Pollution Control Facilities

Wastewater treatment plants were usually sampled three times. Only two results are reported from Red Hook due to sampling errors and four plants (Newtown Creek, 26th Ward, Hunts Point, and Port Richmond) were visited more often. Discharges are given in million gallons per day (MGD).

Table 50. 26th Ward PCB concentrations and homologue abundances.

| sample MGD ng/L | 1/27/99 53 6.9 | 5/5/99 60 3.3 | 9/20/00 64 127 | 6/11/01 83 26 |
|-----------------------|----------------------|---------------------|----------------------|---------------------|
| 1-mono | 0.27% | 1.2% | 0.02% | 0.26% |
| 2-di | 5.1% | 5.8% | 0.26% | 1.5% |
| 3-tri | 16% | 16% | 0.49% | 2.1% |
| 4-tetra | 23% | 20% | 1.2% | 5.9% |
| 5-penta | 28% | 31% | 15% | 16% |
| 6-Hexa | 18% | 18% | 48% | 46% |
| 7-Hepta | 7.1% | 5.7% | 30% | 25% |
| 8-Octa | 2.0% | 1.8% | 4.4% | 3.1% |
| 9-Nona | 0.42% | 0.32% | 0.20% | |
| 10-Deca | 0.27% | 0.14% | 0.01% | |

Table 51. Bowery Bay PCB concentrations and homologue abundances.

| sample | 11/5/98 | 4/21/99 | 9/22/99 |
|---------|---------|---------|---------|
| MGD | 101 | 138 | 103 |
| ng/L | 4.7 | 7.3 | 5.3 |
| 1-mono | 0.47% | 0.25% | 0.30% |
| 2-di | 5.2% | 4.8% | 14% |
| 3-tri | 17% | 13% | 18% |
| 4-tetra | 34% | 25% | 22% |
| 5-penta | 22% | 26% | 20% |
| 6-Hexa | 15% | 21% | 15% |
| 7-Hepta | 5.7% | 7.6% | 8.2% |
| 8-Octa | 0.66% | 1.9% | 1.5% |
| 9-Nona | 0.082% | 0.28% | 0.17% |
| 10-Deca | 0.023% | 0.081% | 0.057% |

Table 52. Coney Island WPCF PCB concentrations and homologue abundances.

| Sample | 10/4/00 | 3/17/99 | 7/28/99 |
|---------|---------|---------|---------|
| MGD | 87 | 105 | 103 |
| ng/L | 1.4 | 3.0 | 2.4 |
| 1-mono | 1.5% | 0.30% | 0.76% |
| 2-di | 14% | 8.8% | 9.4% |
| 3-tri | 18% | 16% | 21% |
| 4-tetra | 19% | 20% | 24% |
| 5-penta | 26% | 29% | 26% |
| 6-Hexa | 15% | 18% | 13% |
| 7-Hepta | 5.9% | 6.8% | 5.0% |
| 8-Octa | 0.75% | 1.84% | 0.53% |
| 9-Nona | 0.17% | 0.36% | 0.066% |
| 10-Deca | 0.08% | 0.09% | 0.05% |

Table 53. Hunts Point WPCF PCB concentrations and homologue abundances.

| Sample | 2/1/01 | 2/19/99 | 3/19/01 | 3/28/01 | 4/11/01 | 4/18/01 | 4/30/99 |
|---------|--------|---------|---------|---------|---------|---------|---------|
| MGD | 142 | 149 | 120 | 181 | 146 | 125 | 133 |
| ng/L | 4.0 | 14 | 0.4 | 3.2 | 2.2 | 3.3 | 6.3 |
| 1-mono | 3.9% | 11% | 4.9% | 3.2% | 3.3% | 5.5% | 3.2% |
| 2-di | 8.9% | 21% | 20% | 13% | 17% | 36% | 15% |
| 3-tri | 17% | 24% | 31% | 18% | 22% | 23% | 22% |
| 4-tetra | 19% | 19% | 19% | 21% | 24% | 15% | 26% |
| 5-penta | 24% | 14% | 15% | 24% | 20% | 12% | 18% |
| 6-Hexa | 17% | 7.6% | 7.6% | 16% | 9.7% | 5.8% | 11% |
| 7-Hepta | 7.7% | 2.9% | 2.4% | 3.9% | 2.6% | 1.4% | 3.7% |
| 8-Octa | 2.2% | 0.68% | 0.52% | 1.03% | 0.57% | 0.29% | 0.98% |
| 9-Nona | 0.58% | 0.18% | | | | | 0.19% |
| 10-Deca | 0.19% | 0.05% | | | | | 0.075% |

Table 54. Jamaica WPCF PCB concentrations and homologue abundances.

| Sample | 2/5/99 | 6/30/99 | 2/15/01 |
|---------|--------|---------|---------|
| MGD | 88 | 84 | 90 |
| ng/L | 7.6 | 7.5 | 4.2 |
| 1-mono | 0.16% | 0.55% | 0.80% |
| 2-di | 3.8% | 6.6% | 5.6% |
| 3-tri | 11% | 18% | 11% |
| 4-tetra | 21% | 28% | 18% |
| 5-penta | 36% | 26% | 34% |
| 6-Hexa | 20% | 15% | 21% |
| 7-Hepta | 5.6% | 4.9% | 6.6% |
| 8-Octa | 1.4% | 0.89% | 1.7% |
| 9-Nona | 0.28% | 0.14% | 0.39% |
| 10-Deca | 0.21% | 0.07% | 0.15% |

Table 55. Newtown Creek WPCF PCB concentrations and homologue abundances.

| sample field QC | 1/5/00 | 1/5/00 | 3/11/99 | 3/11/99 | 6/22/99 | 6/22/99 | 9/28/99 | 9/28/99 |
|--------------------|--------|--------|---------|---------|---------|---------|---------|---------|
| | DU | SA | DU | SA | DU | SA | DU | SA |
| MGD | 249 | 249 | 249 | 249 | 249 | 249 | 249 | 249 |
| ng/L | 25 | 11 | 6.7 | 3.4 | 21 | 12 | 7.5 | 10 |
| 1-mono | 2.0% | 4.1% | 0.37% | 11% | 5.1% | 8.8% | 8.4% | 5.0% |
| 2-di | 5.6% | 5.6% | 3.1% | 14% | 8.6% | 11% | 9.6% | 7.2% |
| 3-tri | 12% | 12% | 9.7% | 21% | 15% | 16% | 17.15% | 15% |
| 4-tetra | 21% | 20% | 18% | 22% | 24% | 21% | 21% | 23% |
| 5-penta | 27% | 25% | 33% | 21% | 26% | 25% | 24% | 27% |
| 6-Hexa | 20% | 21% | 22% | 9.0% | 14% | 13% | 14% | 16% |
| 7-Hepta | 8.4% | 8.9% | 9.4% | 2.0% | 5.4% | 4.7% | 5.0% | 5.7% |
| 8-Octa | 3.0% | 2.8% | 3.5% | | 1.2% | 1.3% | 0.88% | 1.0% |
| 9-Nona | 0.39% | 0.40% | 0.58% | | 0.20% | 0.25% | 0.15% | 0.16% |
| 10-Deca | 0.083% | 0.086% | 0.14% | | 0.047% | 0.054% | 0.047% | 0.044% |

Table 56. North River WPCF, PCB concentrations and homologue abundances.

| sample | 3/24/99 | 9/1/99 | 1/25/01 |
|---------|---------|--------|---------|
| MGD | 153 | 167 | 152 |
| ng/L | 4.7 | 4.4 | 2.4 |
| 1-mono | 0.47% | 0.99% | 0.47% |
| 2-di | 6.8% | 7.6% | 16% |
| 3-tri | 19% | 18% | 24% |
| 4-tetra | 27% | 22% | 20% |
| 5-penta | 24% | 28% | 22% |
| 6-Hexa | 15% | 17% | 12% |
| 7-Hepta | 5.7% | 5.5% | 3.9% |
| 8-Octa | 1.4% | 0.77% | 0.77% |
| 9-Nona | 0.30% | 0.13% | 0.15% |
| 10-Deca | 0.053% | 0.032% | 0.032% |

Oakwood Beach WPCF receives sludge from Port Richmond WPCF. The Port Richmond WPCF receives waste from a pigment manufacturer that inadvertently generates certain PCB congeners, particularly 3,3'-dichlorobiphenyl. This congener accounts for the high proportion of dichlorobiphenyl seen at Port Richmond and Oakwood Beach WPCFs.

Table 57. Oakwood Beach WPCF, PCB concentrations and homologue abundances.

| sample | 2/11/99 | 8/18/99 | 10/13/99 |
|---------|---------|---------|----------|
| MGD | 25 | 25 | 36 |
| ng/L | 5.5 | 13 | 9.6 |
| 1-mono | 0.36% | 0.88% | 0.33% |
| 2-di | 57% | 87% | 80% |
| 3-tri | 12% | 3.2% | 5.5% |
| 4-tetra | 13% | 3.6% | 5.0% |
| 5-penta | 9.2% | 3.4% | 4.8% |
| 6-Hexa | 6.1% | 1.5% | 2.8% |
| 7-Hepta | 1.8% | 0.44% | 0.91% |
| 8-Octa | 0.45% | 0.074% | 0.28% |
| 9-Nona | 0.11% | 0.014% | 0.055% |
| 10-Deca | 0.020% | 0.007% | 0.020% |

Table 58. Owls Head WPCF PCB concentrations and homologue abundances.

| sample | 9/15/98 | 7/7/99 | 8/23/00 |
|---------|---------|--------|---------|
| MGD | 113 | 119 | 115 |
| ng/L | 2.9 | 3.4 | 3.9 |
| 1-mono | 0.75% | 0.65% | 0.77% |
| 2-di | 7.4% | 6.9% | 8.5% |
| 3-tri | 16% | 23% | 18% |
| 4-tetra | 18% | 24% | 24% |
| 5-penta | 30% | 25% | 28% |
| 6-Hexa | 19% | 14% | 15% |
| 7-Hepta | 7.1% | 5.1% | 4.6% |
| 8-Octa | 1.2% | 0.60% | 0.56% |
| 9-Nona | 0.25% | 0.098% | 0.14% |
| 10-Deca | 0.098% | 0.053% | 0.055% |

Table 59. Port Richmond WPCF PCB concentration and homologue abundances.

| sample | 2/24/99 | 8/25/99 | 10/20/99 | 4/11/01 | 4/30/01 |
|---------|---------|---------|----------|---------|---------|
| MGD | 31 | 35 | 78 | 49 | 29 |
| ng/L | 29 | 213 | 143 | 160 | 103 |
| 1-mono | 0.50% | 0.049% | 0.40% | 0.072% | 0.14% |
| 2-di | 65% | 97% | 92% | 98% | 97% |
| 3-tri | 7.5% | 1.1% | 2.5% | 0.57% | 1.2% |
| 4-tetra | 6.3% | 0.78% | 2.3% | 0.38% | 0.64% |
| 5-penta | 5.3% | 0.57% | 1.4% | 0.35% | 0.52% |
| 6-Hexa | 7.9% | 0.24% | 0.87% | 0.22% | 0.35% |
| 7-Hepta | 5.8% | 0.070% | 0.33% | 0.075% | 0.098% |
| 8-Octa | 1.7% | 0.012% | 0.091% | 0.015% | 0.023% |
| 9-Nona | 0.16% | 0.003% | 0.017% | | |
| 10-Deca | 0.026% | 0.0006% | 0.0092% | | |

Table 60. Poughkeepsie WPCF PCB concentrations and homologue abundances.

| sample | 12/5/00 - SA | 4/1/99 - SA | 8/19/99 - DU | 8/19/99 - SA |
|---------|--------------|-------------|--------------|--------------|
| MGD | 4.3 | 7.2 | 4.5 | 4.5 |
| ng/L | 6.4 | 5.9 | 27 | 22 |
| 1-mono | 1.3% | 5.8% | 8.9% | 8.2% |
| 2-di | 11% | 8.2% | 3.5% | 3.8% |
| 3-tri | 11% | 9.2% | 12% | 13% |
| 4-tetra | 15% | 14% | 16% | 17% |
| 5-penta | 32% | 28% | 29% | 28% |
| 6-Hexa | 22% | 21% | 20% | 20% |
| 7-Hepta | 6.0% | 12% | 9.4% | 9.0% |
| 8-Octa | 1.5% | 1.4% | 1.3% | 1.2% |
| 9-Nona | 0.39% | 0.17% | 0.15% | 0.15% |
| 10-Deca | 0.12% | 0.061% | 0.041% | 0.042% |

Table 61. Red Hook (Brooklyn) WPCF PCB concentrations and homologue abundances.

| sample | 2/3/99 | 4/14/99 |
|---------|--------|---------|
| MGD | 39.73 | 30.2 |
| ng/L | 5.4 | 2.1 |
| 1-mono | 0.47% | 0.82% |
| 2-di | 4.9% | 8.9% |
| 3-tri | 16% | 22% |
| 4-tetra | 26% | 31% |
| 5-penta | 21% | 21% |
| 6-Hexa | 19% | 12% |
| 7-Hepta | 8.9% | 4.1% |
| 8-Octa | 2.8% | 0.73% |
| 9-Nona | 0.54% | 0.09% |
| 10-Deca | 0.12% | 0.03% |

Table 62. Renesselaer WPCF PCB concentrations and homologue abundances.

| Sample | 1/12/99 | 3/30/99 | 8/11/99 |
|---------|---------|---------|---------|
| MGD | 16 | 23 | 14 |
| ng/L | 9.3 | 3.0 | 5.5 |
| 1-mono | 1.5% | 1.8% | 1.8% |
| 2-di | 33% | 41% | 25% |
| 3-tri | 10% | 12% | 20% |
| 4-tetra | 13% | 14% | 19% |
| 5-penta | 21% | 17% | 20% |
| 6-Hexa | 15% | 9.3% | 9.5% |
| 7-Hepta | 4.7% | 2.9% | 2.8% |
| 8-Octa | 1.8% | 0.89% | 1.2% |
| 9-Nona | 0.62% | 0.29% | 0.88% |
| 10-Deca | 0.16% | 0.08% | 0.26% |

Table 63. Rockaway WPCF PCB concentrations and homologue abundances.

| Sample | 4/1/99 | 8/11/99 | 11/3/99 |
|---------|--------|---------|---------|
| MGD | 21 | 22 | 19 |
| ng/L | 3.7 | 7.1 | 2.4 |
| 1-mono | 0.50% | 0.34% | 0.90% |
| 2-di | 4.6% | 3.9% | 7.6% |
| 3-tri | 14% | 15% | 26% |
| 4-tetra | 27% | 22% | 23% |
| 5-penta | 25% | 33% | 21% |
| 6-Hexa | 16% | 18% | 13% |
| 7-Hepta | 11% | 6.2% | 5.4% |
| 8-Octa | 1.7% | 1.6% | 1.7% |
| 9-Nona | 0.38% | 0.25% | 0.23% |
| 10-Deca | 0.10% | 0.12% | 0.07% |

Table 64. Rockland County WPCF PCB concentrations and homologue abundances.

| sample | 3/8/00 | 4/20/99 | 8/19/99 |
|---------|--------|---------|---------|
| MGD | 22 | 20 | 17 |
| ng/L | 4.6 | 4.3 | 5.1 |
| 1-mono | 1.5% | 8.6% | 1.7% |
| 2-di | 4.4% | 22% | 8.1% |
| 3-tri | 9.1% | 16% | 20% |
| 4-tetra | 20% | 21% | 22% |
| 5-penta | 28% | 22% | 32% |
| 6-Hexa | 28% | 8.4% | 12% |
| 7-Hepta | 7.7% | 2.2% | 2.8% |
| 8-Octa | 0.81% | 0.31% | 0.32% |
| 9-Nona | 0.13% | 0.04% | 0.10% |
| 10-Deca | 0.10% | 0.02% | 0.03% |

Table 65. Tallman Island WPCF PCB concentrations and homologue abundances.

| Sample | 2/12/99 | 7/20/99 | 9/6/00 |
|---------|---------|---------|--------|
| MGD | 56 | 59 | 41 |
| ng/L | 5.9 | 5.1 | 4.9 |
| 1-mono | 0.29% | 0.43% | 0.47% |
| 2-di | 6.7% | 9.4% | 6.0% |
| 3-tri | 9.2% | 25% | 14% |
| 4-tetra | 15% | 25% | 22% |
| 5-penta | 23% | 20% | 29% |
| 6-Hexa | 28% | 13% | 20% |
| 7-Hepta | 14% | 6.7% | 7.6% |
| 8-Octa | 3.3% | 0.76% | 1.7% |
| 9-Nona | 0.32% | 0.086% | 0.28% |
| 10-Deca | 0.08% | 0.021% | 0.055% |

Table 66. Wards Island WPCF PCB concentrations and homologue abundances.

| Sample | 1/20/99 - SA | 4/28/99 - SA | 8/10/00 - DU | 8/10/00 - SA |
|---------|--------------|--------------|--------------|--------------|
| MGD | 221 | 179 | 220 | 220 |
| ng/L | 3.2 | 1.8 | 2.6 | 2.2 |
| 1-mono | 0.94% | 0.49% | 1.5% | 1.7% |
| 2-di | 8.1% | 4.3% | 7.0% | 7.2% |
| 3-tri | 17% | 15% | 13% | 14% |
| 4-tetra | 22% | 23% | 23% | 22% |
| 5-penta | 25% | 29% | 30% | 27% |
| 6-Hexa | 17% | 19% | 17% | 19% |
| 7-Hepta | 7.0% | 7.5% | 6.0% | 7.3% |
| 8-Octa | 1.8% | 1.2% | 1.5% | 1.4% |
| 9-Nona | 0.61% | 0.25% | 0.35% | 0.36% |
| 10-Deca | 0.17% | 0.12% | 0.10% | 0.10% |

Table 67. Yonkers WPCF PCB concentrations and homologue abundances.

| Sample | 4/22/99 | 8/18/99 | 3/22/00 |
|---------|---------|---------|---------|
| MGD | 89 | 85 | 95 |
| ng/L | 2.0 | 2.5 | 8.6 |
| 1-mono | 0.28% | 0.85% | 0.55% |
| 2-di | 4.6% | 7.2% | 2.3% |
| 3-tri | 16% | 25% | 12% |
| 4-tetra | 24% | 24% | 19% |
| 5-penta | 32% | 27% | 23% |
| 6-Hexa | 17% | 12% | 28% |
| 7-Hepta | 4.0% | 4.0% | 12.6% |
| 8-Octa | 1.1% | 0.34% | 1.9% |
| 9-Nona | 0.25% | 0.075% | 0.28% |
| 10-Deca | 0.066% | 0.020% | 0.18% |

NYSDEC sampled NJ treatment plants at Edgewater and PVSC. Samples (34 L at Edgewater and 47L at PVSC) were composited over 24 hours beginning on May 21, 2001. They were brought back to the NYSDEC lab and processed identically to the ways of the routine NYSDEC samples. 3,3'-DiCB accounts for almost all of the PCB at PVSC.

Table 68. Edgewater and PVSC PCB concentrations and homologue abundances.

| Name | Edgewater | PVSC |
|---------|-----------|-------|
| Ng/L | 6.5 | 330 |
| 1-mono | 0.72% | 0.03% |
| 2-di | 5.6% | 92% |
| 3-tri | 8.1% | 2.8% |
| 4-tetra | 13% | 2.3% |
| 5-penta | 19% | 1.5% |
| 6-Hexa | 29% | 0.95% |
| 7-Hepta | 20% | 0.53% |
| 8-Octa | 4.2% | 0.14% |
| 9-Nona | 0.44% | 0.03% |
| 10-Deca | 0.11% | 0.01% |

CSO/SWO

Sixteen samples were taken to represent combined sewer overflows (CSOs) and storm water overflows (SWOs). The CSO samples were wet weather influents to treatment plants. Sampling details are shown elsewhere. SWO samples were taken in two locations in the Jamaica section of Queens. One represents a commercial district and the other represents an industrial area. Table 69 lists the names of the sites and the abbreviations used in Table 70. The order in the list is the same as in the numerical table, in order of ascending concentration. Discharges (MGD) shown on Table 70 are modeled daily total releases from all the individual CSOs in the WPCF's drainage. Three facilities (26th Ward, Bowery Bay, and Newtown Creek) were sampled at two influent points. The two Newtown Creek points are "Manhattan Pump Station" and "Newtown Creek". In these cases, the same MGD value is used for both. Readers may wish to average the concentrations in calculating a load. Manhattan Grit Chamber was used to evaluate raw influent entering the Wards Island facility. Similar values are not available for the two SWOs.

The predominance of dichlorobiphenyls at Port Richmond is due to the inadvertently synthesized PCB congener 3,3'DiCB. The predominance of hexachlorobiphenyls at both 26th Ward samples reflects the Aroclor 1260 previously noted there. It is interesting that high concentrations of Aroclor 1260 occur in both influents reflecting the widespread contamination of this formerly industrial area.

Table 69. Names and abbreviations used in Table 70.

| name | abbreviation |
|----------------------------------|-------------------|
| 26 th Ward, High Side | 26 High Inf |
| Red Hook Influent | RH Inf |
| 26 th Ward, Low Side | 26, Low Inf |
| Port Richmond Influent | PR Inf |
| North River Influent | NR Inf |
| Bowery Bay High Side | BB, High Inf |
| Newtown Creek Influent | NC Inf |
| Manhattan Pump Station | Man. PS |
| Manhattan Grit Chamber | Man. Grit Chamber |
| Bowery Bay Low Side | BB, Low Inf |
| SWO-Jamaica, Industrial | SWO-Jam Ind. |
| Owls Head Influent | OH Inf |
| Jamaica Influent | JA Inf |
| Hunts Point Influent | HP Inf |
| SWO-Jamaica, Commercial | SWO-Jam Com. |
| Coney Island Influent | CI Inf |

Table 70. CSO/SWO PCB concentrations and homologue abundances.

| Sample | 26 High Inf | RH Inf | 26, Low Inf | PR Inf | NR Inf | BB, High Inf | NC Inf | Man. PS |
|---------|----------------|--------|----------------|--------|--------|-----------------|--------|---------|
| MGD | 12 | 3.7 | 12 | 1.0 | 5.0 | 13 | 14 | 14 |
| ng/L | 3500 | 1300 | 850 | 560 | 350 | 300 | 260 | 150 |
| 1-mono | 0.004% | 0.077% | 0.040% | 0.73% | 0.40% | 0.090% | 0.81% | 0.33% |
| 2-di | 0.07% | 2.1% | 0.50% | 87% | 1.5% | 1.2% | 4.2% | 2.0% |
| 3-tri | 0.38% | 19% | 1.7% | 3.1% | 4.0% | 4.1% | 12% | 5.1% |
| 4-tetra | 3.4% | 37% | 3.6% | 3.59% | 11% | 9.2% | 19% | 9.5% |
| 5-penta | 9.7% | 17% | 9.4% | 2.65% | 32% | 20% | 26% | 26% |
| 6-Hexa | 42% | 13% | 43% | 2.06% | 32% | 34% | 22% | 27% |
| 7-Hepta | 36% | 8.5% | 34% | 0.89% | 13% | 23% | 12% | 20% |
| 8-Octa | 8.0% | 2.6% | 6.7% | 0.24% | 4.0% | 6.8% | 3.6% | 7.4% |
| 9-Nona | 0.38% | 0.34% | 0.34% | 0.065% | 1.2% | 0.84% | 0.68% | 1.7% |
| 10-Deca | 0.010% | 0.064% | 0.031% | 0.021% | 0.18% | 0.14% | 0.18% | 0.18% |

Table 70 continued.

| Sample | Man. Grit Chamber | BB, Low Inf | SWO-Jam Ind | OH Inf | JA Inf | HP Inf | SWO-Jam Com | CI Inf |
|---------|-------------------|-------------|-------------|--------|--------|--------|-------------|--------|
| MGD | 11 | 13 | NA | 9.3 | 31 | 15 | NA | 10.0 |
| ng/L | 130 | 110 | 70 | 66 | 65 | 58 | 48 | 44 |
| 1-mono | 0.16% | 0.35% | 0.24% | 0.30% | 0.63% | 0.39% | 0.48% | 0.80% |
| 2-di | 1.1% | 4.5% | 4.1% | 3.1% | 3.7% | 3.8% | 3.2% | 3.8% |
| 3-tri | 3.2% | 8.4% | 13% | 7.8% | 9.1% | 11% | 5.2% | 7.7% |
| 4-tetra | 13% | 17% | 22% | 16% | 17% | 17% | 12% | 15% |
| 5-penta | 40% | 31% | 24% | 29% | 32% | 23% | 28% | 32% |
| 6-Hexa | 29% | 25% | 22% | 26% | 24% | 27% | 29% | 26% |
| 7-Hepta | 8.5% | 9.8% | 10% | 13% | 9.3% | 12% | 16% | 10% |
| 8-Octa | 3.5% | 3.6% | 2.9% | 3.1% | 2.8% | 4.1% | 5.3% | 3.1% |
| 9-Nona | 2.0% | 0.84% | 0.83% | 1.1% | 0.88% | 1.1% | 1.05% | 0.94% |
| 10-Deca | 0.30% | 0.26% | 0.29% | 0.64% | 0.48% | 0.38% | 0.25% | 0.61% |

Wastewater Treatment Plant Sludges

Table 71 shows the site names and the abbreviations used in Table 72. All the sludge samples, except the one from Newtown Creek (NC) were composited from daily collections made during February 2001. Port Richmond and Oakwood Beach (PR and OB) have a predominance of 3,3'-DiCB. So, too, does Wards Island South (WI, South).

Table 71. Names used in Table 72.

| | |
|----------------------------------|-----------|
| 26th Ward WPCF, Sludge | 26W |
| Bowery Bay WPCF, Sludge | BB |
| Coney Island WPCF, Sludge | CI |
| Hunts Point WPCF #10 Sludge | HP#10 |
| Hunts Point WPCF #9 Sludge | HP#9 |
| Jamaica WPCF Sludge | JA |
| Oakwood Beach WPCF, Sludge | OB |
| Port Richmond WPCF, Sludge | PR |
| Red Hook WPCF, Sludge | RH |
| Tallman Island WPCF, Sludge | TI |
| Wards Island WPCF, North, Sludge | WI, North |
| Wards Island WPCF, South, Sludge | WI, South |

Table 72. PCB concentrations and homologue abundances in WPCF sludges.
Note that the units are mg/kg (ppm).

| Name mg/kg | PR 12 | 26W 2.6 | WI, South 2.0 | OB 1.6 | JA 1.4 | RH 0.82 | CI 0.69 | HP#10 0.61 |
|---------------|----------|------------|------------------|-----------|-----------|------------|------------|---------------|
| 1-mono | 0.27% | 0.039% | 0.37% | 0.23% | 0.12% | 0.085% | 0.14% | 0.73% |
| 2-di | 94% | 0.53% | 44% | 85% | 1.8% | 2.5% | 1.8% | 5.4% |
| 3-tri | 2.5% | 2.0% | 7.3% | 2.6% | 6.3% | 16% | 5.2% | 10% |
| 4-tetra | 1.4% | 4.1% | 9.9% | 3.5% | 14% | 25% | 9.2% | 17% |
| 5-penta | 0.97% | 12% | 16% | 4.1% | 30% | 21% | 19% | 26% |
| 6-Hexa | 0.87% | 43% | 15% | 3.1% | 29% | 21% | 36% | 26% |
| 7-Hepta | 0.35% | 31% | 5.4% | 1.3% | 15% | 9.7% | 23% | 11% |
| 8-Octa | 0.090% | 6.7% | 1.6% | 0.40% | 3.2% | 3.1% | 4.7% | 3.0% |
| 9-Nona | 0.024% | 0.43% | 0.54% | 0.11% | 0.48% | 0.91% | 0.68% | 0.95% |
| 10-Deca | 0.010% | 0.051% | 0.21% | 0.04% | 0.18% | 0.22% | 0.18% | 0.30% |

Table 72 continued.

| Sample mg/kg | BB 0.60 | HP#9 0.59 | TI 0.41 | WI, North 0.40 |
|-----------------|------------|--------------|------------|-------------------|
| 1-mono | 0.15% | 1.2% | 0.22% | 0.19% |
| 2-di | 2.4% | 4.9% | 2.4% | 7.2% |
| 3-tri | 7.0% | 12% | 8.2% | 9.5% |
| 4-tetra | 17% | 19% | 17% | 17% |
| 5-penta | 27% | 25% | 26% | 27% |
| 6-Hexa | 26% | 24% | 25% | 24% |
| 7-Hepta | 15% | 9.2% | 16% | 9.2% |
| 8-Octa | 4.0% | 2.8% | 4.5% | 3.5% |
| 9-Nona | 0.77% | 0.89% | 0.69% | 1.7% |
| 10-Deca | 0.39% | 0.30% | 0.15% | 0.41% |

Landfill Leachates

Landfill leachates were only sampled in the dissolved (filtered) phase. There are methods for estimating the amount of liquid leachate that leaves a landfill, but we lack a way to estimate transport of particles from within to outside a mound. Three landfills were sampled; Pelham Bay (PB) in the Bronx, Fresh Kills (FK) on Staten Island, and the New Jersey Meadowlands Commission (formerly Hackensack Meadowlands Development Commission, HMDC). Pelham Bay leachate is collected into holding tanks and then trucked to the Hunts Point WPCF for treatment. Most of the Fresh Kills leachate is collected and treated by a specially built treatment plant on site. Effluent from this plant was also sampled (Table 74). Some of the leachate from the HMDC is treated at Passaic Valley Sewerage Commissioners (PVSC). Pelham Bay holding tanks were sampled twice. The two other much larger operations were sampled at different points and at different times.

Table 73. Leachates, PCBs and relative homologue abundances.

| name sample ng/L | 1E-HMDC | 1A-HMDC | FK 1/9 Comp. | 1E-HMDC | FK 1/9 Comp. | FK 1/9 Comp. |
|------------------------|-----------------|----------------|-----------------|---------|--------------|--------------------|
| | 6/22/00 2200 | 6/22/00 950 | 10/25/00 950 | 800 | 9/14/01 | 3/20/01 4/19/01 |
| 1-mono | 4.6% | 8.2% | 3.2% | 1.5% | 3.6% | 0.88% |
| 2-di | 14% | 32% | 14% | 16% | 14% | 15% |
| 3-tri | 32% | 39% | 29% | 30% | 32% | 36% |
| 4-tetra | 19% | 14% | 27% | 23% | 31% | 34% |
| 5-penta | 10% | 3.9% | 14% | 9.5% | 12% | 12% |
| 6-hexa | 9.6% | 1.9% | 8.1% | 11% | 6.0% | 1.4% |
| 7-hepta | 6.1% | 0.83% | 3.0% | 6.8% | 1.6% | 0.37% |
| 8-octa | 2.5% | 0.34% | 1.16% | 1.9% | 0.40% | 0.08% |
| 9-nona | 0.86% | 0.050% | 0.36% | | | |
| 10-deca | 0.16% | 0.014% | 0.24% | | | |

Table 73 continued.

| Name Sample ng/L | FK 6/7 Comp. | FK 6/7 Comp. | FK 6/7 Comp. | FK 1/9 Comp. | FK LF 1/9 "B" | 1D-HMDC |
|------------------------|-----------------|---------------|----------------|----------------|----------------|----------------|
| | 10/25/00 430 | 8/9/01 310 | 7/25/01 260 | 5/11/00 240 | 5/11/00 190 | 9/14/01 110 |
| 1-mono | 28% | 18% | 20% | 2.3% | 3.3% | 0.32% |
| 2-di | 48% | 31% | 40% | 9.7% | 12% | 2.0% |
| 3-tri | 17% | 27% | 26% | 23% | 31% | 6.64% |
| 4-tetra | 4.8% | 17% | 9.3% | 34% | 33% | 18% |
| 5-penta | 1.0% | 4.6% | 2.3% | 16% | 11% | 23% |
| 6-hexa | 0.61% | 1.9% | 1.6% | 9.1% | 6.1% | 28% |
| 7-hepta | 0.31% | 0.75% | 0.79% | 4.4% | 2.4% | 16.63% |
| 8-octa | 0.12% | 0.25% | 0.35% | 0.73% | 0.34% | 5.0% |
| 9-nona | | | | 0.22% | 0.12% | |
| 10-deca | | | | 0.38% | 0.35% | |

Table 73 continued.

| Name | FK 6/7 Comp. | FK 3/4 | FK 1/9 "F" | 1D-HMDC | Pelham Bay | Pelham Bay-DU | Pelham Bay |
|---------|--------------|---------|------------|---------|------------|---------------|------------|
| Sample | 5/11/00 | 5/11/00 | 5/11/00 | 6/22/00 | 11/6/98 | 1/29/01 | 1/29/01 |
| ng/L | 100 | 97 | 87 | 73 | 22 | 3.7 | 1.8 |
| 1-mono | 0.16% | 5.8% | 6.1% | 0.42% | 15% | 6.7% | 0.34% |
| 2-di | 48% | 34% | 26% | 3.6% | 16% | 27% | 4.64% |
| 3-tri | 24% | 21% | 20% | 18% | 31% | 23% | 11% |
| 4-tetra | 17% | 20% | 27% | 24% | 23% | 16% | 28% |
| 5-penta | 5.6% | 9.1% | 11% | 24% | 9.0% | 11% | 24% |
| 6-hexa | 3.2% | 5.0% | 5.2% | 17% | 4.4% | 9.57% | 20% |
| 7-hepta | 1.8% | 2.8% | 3.3% | 8.4% | 1.0% | 4.9% | 8.7% |
| 8-octa | 0.46% | 0.45% | 0.68% | 3.6% | 0.16% | 2.3% | 2.1% |
| 9-nona | 0.10% | 0.24% | 0.15% | 0.95% | | | 0.45% |
| 10-deca | 0.18% | 0.95% | 0.24% | 0.46% | | | 0.46% |

Industrial Effluents

Few New York City industrial concerns discharge directly to surface waters. Two facilities were sampled, Clean Waters of New York (an industrial waste processor) and New York City Department of Sanitation's Fresh Kills Treatment Plant (FK, Eff). This state of the art facility treats only leachates from mounds 1,6,7, and 9.

Table 74. PCB concentrations and homologue abundances from two "industrial" dischargers.

| name | Clean Waters | Clean Waters | FK, Eff | FK, Eff | FK, Eff | FK, Eff |
|---------|--------------|--------------|----------|---------|---------|---------|
| sample | 4/29/99 | 9/20/99 | 10/25/00 | 3/20/01 | 4/19/01 | 7/25/01 |
| MGD | | | 0.26 | 0.56 | 0.67 | |
| ng/L | 0.022 | 0.007 | 11 | 30 | 19 | 11 |
| 1-mono | 5.2% | 2.8% | 0.89% | 0.09% | 0.25% | 0.48% |
| 2-di | 6.1% | | 9.4% | 3.4% | 12% | 12% |
| 3-tri | 17% | 17% | 18% | 7.3% | 19% | 21% |
| 4-tetra | 18% | 22% | 27% | 26% | 29% | 29% |
| 5-penta | 18% | 35% | 20% | 28% | 19% | 18% |
| 6-Hexa | 18% | 15% | 14% | 20% | 13% | 12% |
| 7-Hepta | 11% | 6.8% | 6.6% | 10% | 5.3% | 5.0% |
| 8-Octa | 4.4% | 1.1% | 2.6% | 3.0% | 2.0% | 1.9% |
| 9-Nona | 1.2% | | 0.66% | 0.80% | 0.21% | 0.12% |
| 10-Deca | 1.1% | | 0.53% | 0.70% | 0.16% | 0.08% |

Trackdown

A small level of effort was taken toward identifying PCB sources using PISCES. As noted above, Aroclor 1260 was seen entering and leaving 26th Ward. Table 75 shows the results. Samples at both Van Siclen Ave. and Hendrix St. showed unusually high

concentrations of a heavy PCB mixture. These samples were from two separate mains indicating widespread PCB contamination. The same effect was also noted in the raw wet weather influent samples.

Table 75. PISCES results for PCBs in 26th Ward, 6/7/01 – 6/22/01.

| | Van Siclen Ave. 1800 ng/L | Hendrix St. 210 ng/L | Flatlands St. 17 ng/L |
|---------|------------------------------|-------------------------|--------------------------|
| 1-mono | 0.04% | 0.24% | 2.3% |
| 2-di | 0.14% | 0.76% | 9.4% |
| 3-tri | 0.38% | 1.6% | 14% |
| 4-tetra | 1.4% | 4.5% | 17% |
| 5-penta | 9.4% | 11% | 22% |
| 6-Hexa | 44% | 42% | 19% |
| 7-Hepta | 37% | 32% | 10% |
| 8-Octa | 7.3% | 7.1% | 3.9% |
| 9-Nona | 0.35% | 0.38% | 0.78% |
| 10-Deca | 0.01% | 0.01% | 0.19% |

Two different PISCES surveys were carried out in the Newtown Creek service area on 1/18-1/29/01 and later on 6/7-6/22/01 (Figure 25). Here too, two separate mains (Greenpoint and Manhattan, south) indicate relatively high PCB concentrations (Table 76).



Figure 25. PISCES sampling points in the Newtown Creek area.

76. PISCES results for PCBs from the Newtown Creek area.

| | Greenpoint Ave. Manhattan, south 1/18-1/29/01 120 ng/L | Newtown Cr. bar screen 1/18-1/29/01 100 ng/L | Houston St. 1/18-1/29/01 63 ng/L | Maspeth Ave. 6/7-6/22/01 49 ng/L | 1/18-1/29/01 41 ng/L |
|---------|--|--|--|--|-------------------------|
| 1-mono | 14% | 7.6% | 4.8% | 4.4% | 1.9% |
| 2-di | 12% | 9.9% | 9.82% | 7.1% | 7.6% |
| 3-tri | 17% | 17% | 18% | 15% | 16% |
| 4-tetra | 19% | 18% | 20% | 17% | 19% |
| 5-penta | 18% | 20% | 19% | 21% | 24% |
| 6-Hexa | 12% | 16% | 16% | 21% | 19% |
| 7-Hepta | 4.7% | 9.1% | 8.6% | 10% | 9.9% |
| 8-Octa | 1.2% | 2.6% | 2.5% | 3.1% | 2.7% |
| 9-Nona | 0.20% | 0.58% | 0.41% | 0.73% | 0.51% |
| 10-Deca | 0.05% | 0.16% | 0.12% | 0.18% | 0.11% |

Table 76 continued.

| | Manhattan, north 1/18-1/29/01 40 ng/L | Franklin St. 6/7-6/22/01 37 ng/L | South St. 6/7-6/22/01 36 ng/L | Kent Ave. 6/7-6/22/01 26 ng/L | Nassau Ave. 1/18-1/29/01 18 ng/L | Johnson Ave. 1/18-1/29/01 17 ng/L |
|---------|---|--|-------------------------------------|-------------------------------------|--|---|
| 1-mono | 2.4% | 1.9% | 5.0% | 4.2% | 5.7% | 2.6% |
| 2-di | 8.8% | 6.5% | 7.1% | 10% | 14% | 9.3% |
| 3-tri | 15% | 14% | 14% | 16% | 23% | 13% |
| 4-tetra | 18% | 26% | 17% | 18% | 18% | 18% |
| 5-penta | 24% | 21% | 21% | 24% | 17% | 28% |
| 6-Hexa | 19% | 17% | 19% | 18% | 13% | 19% |
| 7-Hepta | 9.0% | 9.2% | 12% | 7.6% | 6.1% | 6.1% |
| 8-Octa | 2.7% | 3.0% | 4.4% | 2.2% | 1.9% | 2.2% |
| 9-Nona | 0.65% | 0.37% | 0.98% | 0.49% | 0.43% | 0.80% |
| 10-Deca | 0.11% | 0.08% | 0.10% | 0.15% | 0.11% | 0.18% |

A PISCES survey was carried out in Staten Island to positively identify the suspected source of the 3,3'-dichlorobiphenyl. Sample locations are shown in Figure 26 and results appear on Table 77. Sampling dates were July 27 to August 2, 2000. Total concentrations derived from PISCES are only moderately quantitative. Nevertheless, the concentrations from the pigment manufacturer outfall and the Port Richmond WPCF influent are extraordinary. The unusual composition of the pigment outfall material is emphasized in Table 78 which shows the three most important congeners from these two sites. The apparent loss of 3,3',4,4'-tetrachlorobiphenyl (IUPAC 77) between pigment manufacturer and Port Richmond is unexplained. Samplers at the pigment manufacturer and the WPCF were both in place for seven days.

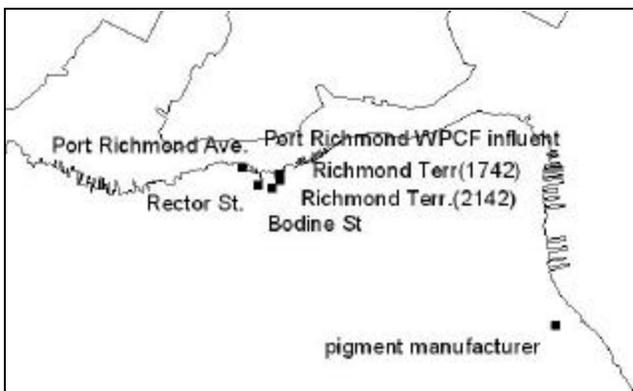


Figure 26. PISCES sampling locations on Staten Island.

Table 77. PCB trackdown (7/27/00 to 8/2/00) on Staten Island.

| | pigment outfall 5900 ng/L | PR WPCF inf 520 ng/L | 2142 Rich. T. 18 ng/L | Bodine St. 12 ng/L | Rector St. 12 ng/L | Pt Rich. Ave. 2.7 ng/L | 1742 Rich. T. 1.3 ng/L |
|---------|------------------------------|-------------------------|--------------------------|-----------------------|-----------------------|---------------------------|---------------------------|
| 1-mono | 0.06% | 0.17% | 2.2% | 5.7% | 3.4% | 2.5% | 3.0% |
| 2-di | 61% | 95% | 14% | 13% | 9.7% | 15% | 19% |
| 3-tri | 6.6% | 1.67% | 32% | 14% | 11% | 21% | 16% |
| 4-tetra | 32% | 1.66% | 26% | 21% | 19% | 20% | 17% |
| 5-penta | 0.10% | 0.80% | 13% | 29% | 30% | 19% | 25% |
| 6-Hexa | 0.018% | 0.41% | 7.2% | 12% | 20% | 15% | 15% |
| 7-Hepta | 0.0033% | 0.12% | 3.3% | 3.8% | 5.0% | 4.7% | 2.8% |
| 8-Octa | 0.0007% | 0.027% | 0.83% | 1.1% | 1.2% | 1.4% | 0.74% |
| 9-Nona | 0.0004% | 0.006% | 0.16% | 0.18% | 0.25% | 0.45% | 0.19% |
| 10-Deca | 0.0025% | 0.0014% | 0.035% | 0.040% | 0.067% | 0.10% | 0.066% |

Table 78. Top congeners at PR WPCF influent and pigment outfall, PISCES

| IUPAC | Percent of total PCB | PR Influent | Pigment Outfall |
|-------|-------------------------------|-------------|-----------------|
| | | 95% | 99% |
| 11 | 3,3'-dichlorobiphenyl | 490 | 3600 |
| 77 | 3,3',4,4'-tetrachlorobiphenyl | 2.5 | 1900 |
| 35 | 3,3',4-trichlorobiphenyl | 2.3 | 380 |

Congener Analysis

The preceding tables of homologue abundances show a trend toward heavier PCB mixtures in the harbor but the trends are difficult to visualize. Figure 27 uses percent abundances of congeners “unique” to Aroclors 1016/1242 and 1254/1260 (see Table 12) to highlight the patterns. “Unique” appears in quotation marks because a small degree of over-lap is permitted. Overall, the two sets of “unique” congeners account for an average of about 50% of the total mass of PCBs in all the samples. Thirty-four congeners were used to describe 1016/1242 and 41 congeners describe 1254/1260.

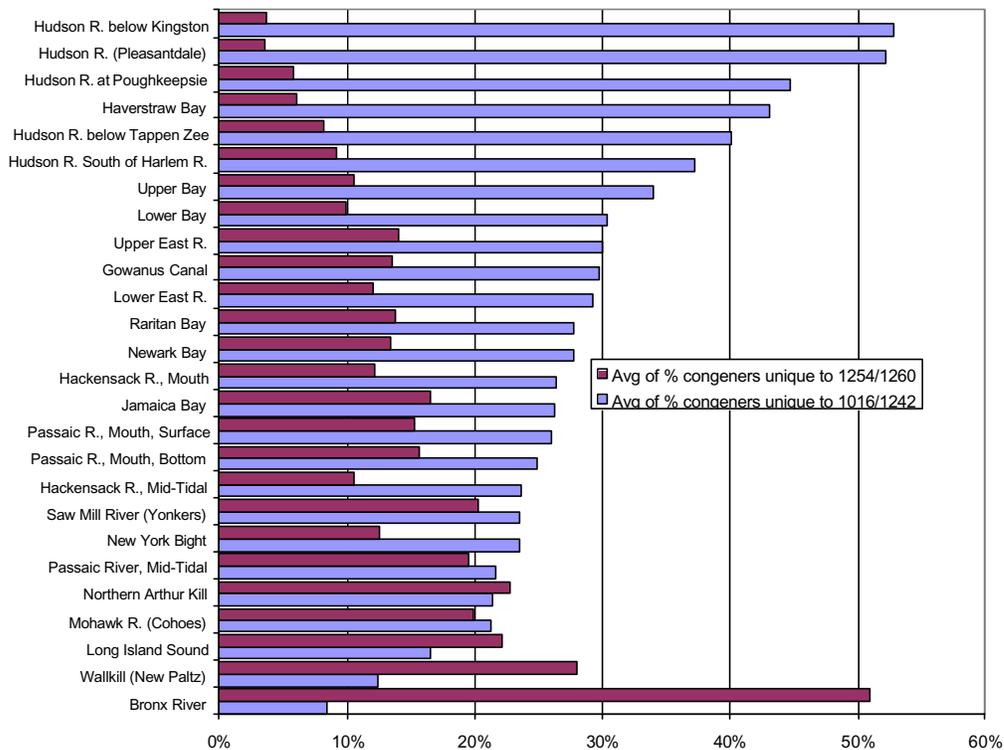


Figure 27. Average percent abundances of congeners “unique” to Aroclors 1016/1242 and 1254/1260 in ambient and tributary sites.

The change in apparent Aroclor composition requires sources of 1254/1260. Both the Bronx River and the Wallkill are such sources. The change in ambient patterns from upstream Hudson toward the harbor is unlikely to be due entirely to volatilization of lighter congeners. The heavy congeners seen to be having increasing abundance in down stream sites are not expected to occur in the lighter Aroclors from the upper Hudson.

Figure 28 shows that CSO/SWO discharges have significantly heavier Aroclors than the ambient samples. WPCFs (Figure 29) show a distribution of Aroclors heavier than those from ambient samples but lighter (a greater proportion of congeners “unique” to 1016/1242) than CSO/SWOs. Figure 30 shows the distributions of light and heavy congeners in biosolids. The very limited samples from landfills also fail to find a metropolitan source for the 1254/1260 congeners (Figure 31). Leachate samples excluded the particulate phase biasing the congener distribution.

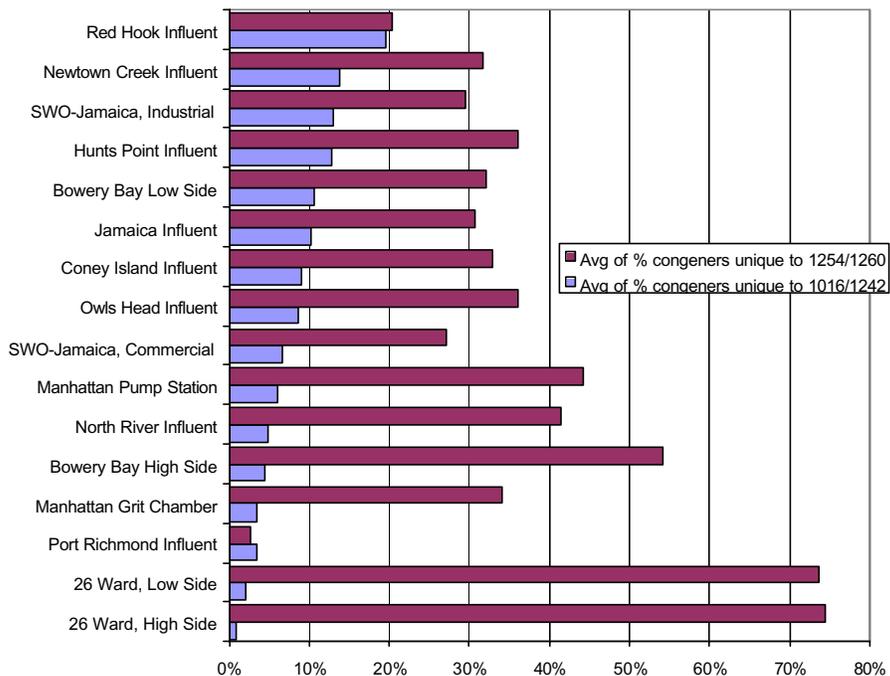


Figure 28. Average percent abundances of congeners “unique” to Aroclors 1016/1242 and 1254/1260 in CSO and SWO sites.

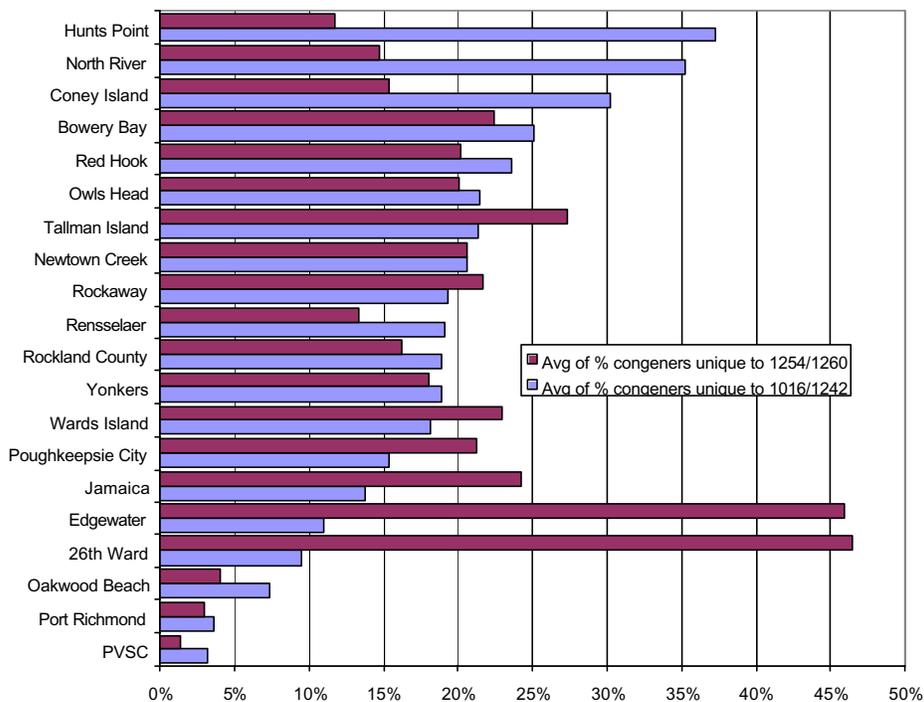


Figure 29. Average percent abundances of congeners “unique” to Aroclors 1016/1242 and 1254/1260 in WPCF sites.

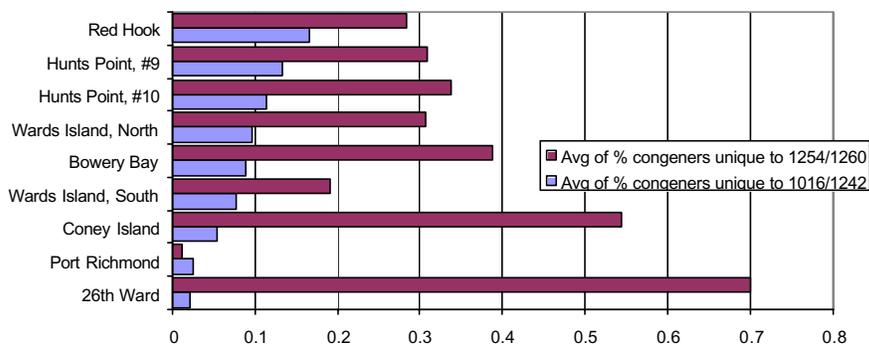


Figure 30. Average percent abundances of congeners “unique” to Aroclors 1016/1242 and 1254/1260 in sewage treatment plant biosolids.

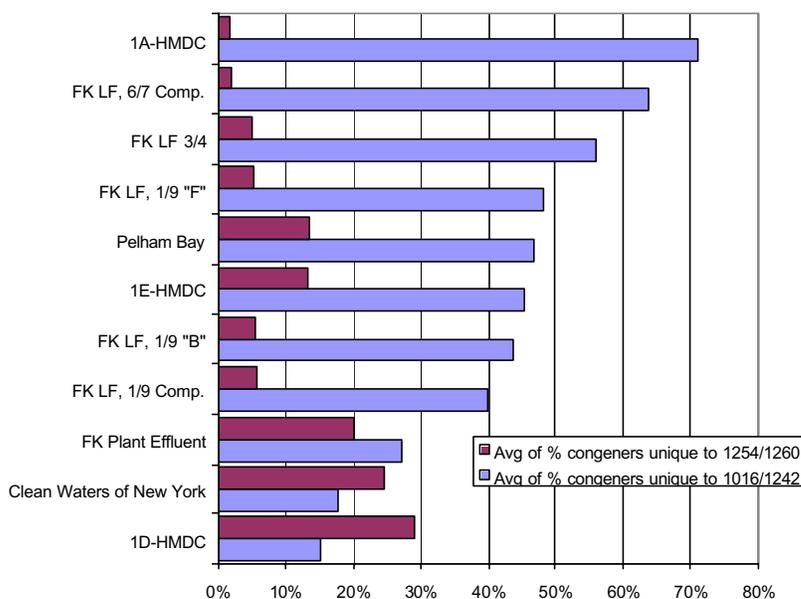


Figure 31. Average percent abundances of congeners “unique” to Aroclors 1016/1242 and 1254/1260 in landfill leachate and landfill treated leachate effluent sites.

PCDD/PCDFs

This section deals with 2,3,7,8-substituted chlorinated dibenzodioxins and dibenzofurans. A separate discussion (Dioxins and Related Compounds: Are Regulators Measuring the Right Chemicals?) discusses with broader issues of dioxin-like properties in other chemicals including the co-planar PCBs.

Dioxins/furans are treated collectively using Toxic Equivalency Factors (TEFs). Two sets of TEFs are used. NYSDEC uses WHO94 values in its Water Quality Standards. WHO98 reflects more recent science. We will use WHO98 values in calculating Toxic Equivalents (TEQs).

Table 79. Dioxin/furan names, ordering, and two TEFs.

| PARAM | Order | WHO94 | WHO98 | BEF |
|---------------------|-------|-------|--------|------|
| 2,3,7,8-TCDD | 1 | 1 | 1 | 1 |
| 1,2,3,7,8-PeCDD | 2 | 0.5 | 1 | 0.9 |
| 1,2,3,4,7,8-HxCDD | 3 | 0.1 | 0.1 | 0.3 |
| 1,2,3,6,7,8-HxCDD | 4 | 0.1 | 0.1 | 0.1 |
| 1,2,3,7,8,9-HxCDD | 5 | 0.1 | 0.1 | 0.1 |
| 1,2,3,4,6,7,8-HpCDD | 6 | 0.01 | 0.01 | 0.05 |
| OCDD | 7 | 0.001 | 0.0001 | 0.01 |
| 2,3,7,8-TCDF | 8 | 0.1 | 0.1 | 0.8 |
| 1,2,3,7,8-PeCDF | 9 | 0.05 | 0.05 | 0.2 |
| 2,3,4,7,8-PeCDF | 10 | 0.5 | 0.5 | 1.6 |
| 1,2,3,4,7,8-HxCDF | 11 | 0.1 | 0.1 | 0.08 |
| 1,2,3,6,7,8-HxCDF | 12 | 0.1 | 0.1 | 0.2 |
| 2,3,4,6,7,8-HxCDF | 13 | 0.1 | 0.1 | 0.7 |
| 1,2,3,7,8,9-HxCDF | 14 | 0.1 | 0.1 | 0.6 |
| 1,2,3,4,6,7,8-HpCDF | 15 | 0.01 | 0.01 | 0.01 |
| 1,2,3,4,7,8,9-HpCDF | 16 | 0.01 | 0.01 | 0.4 |
| OCDF | 17 | 0.001 | 0.0001 | 0.02 |

Data Quality

Calculation of TEQs require sufficient data. Since the TEF weighting factors span many orders of magnitude, it is essential that the congeners with high TEFs be detected. TEQs calculated when high TEF substances are not detected are underestimations. The rule used here was that the difference between TEQs using non-detections set to zero or to half the detection level must be less than 10%. Application of this rule ensures that sufficient analyte masses were collected and obviates issues with analyses being near the detection level or samples with lab blank interferences.

Table 80 shows the level of success in quantitating dioxin/furan congeners. The Order on Table 80 is described in Table 79.

Table 80. Success in quantitating dioxin/furan congeners.

| Order | Amb-clean | | Amb-Hud. | | Amb-Kills | | Amb-Non_Kills | | CSO | | Ind. eff. | | Trib. | | WPCF | |
|-------|-----------|----|----------|----|-----------|----|---------------|----|------|----|-----------|----|-------|-----|------|-----|
| | Det. | ND | Det. | ND | Det. | ND | Det. | ND | Det. | ND | Det. | ND | Det. | ND | Det. | ND |
| 1 | 9 | 5 | 16 | 8 | 22 | | 16 | 8 | 8 | | 4 | 1 | 17 | 30 | 26 | 44 |
| 2 | 12 | 2 | 20 | 4 | 19 | 3 | 20 | 4 | 8 | | 3 | 2 | 22 | 25 | 63 | 7 |
| 3 | 11 | 3 | 21 | 3 | 19 | 3 | 20 | 4 | 8 | | 2 | 3 | 23 | 24 | 59 | 11 |
| 4 | 13 | 1 | 23 | 1 | 20 | 2 | 21 | 3 | 8 | | 5 | | 27 | 20 | 68 | 2 |
| 5 | 12 | 2 | 23 | 1 | 19 | 3 | 22 | 2 | 8 | | 5 | | 28 | 19 | 67 | 3 |
| 6 | 14 | | 24 | | 22 | | 23 | 1 | 8 | | 5 | | 44 | 3 | 69 | 1 |
| 7 | 14 | | 24 | | 22 | | 24 | | 8 | | 5 | | 47 | 0 | 70 | |
| 8 | 12 | 2 | 23 | 1 | 22 | | 23 | 1 | 8 | | 3 | 2 | 33 | 14 | 63 | 3 |
| 9 | 11 | 3 | 20 | 4 | 19 | 3 | 21 | 3 | 8 | | 4 | 1 | 23 | 24 | 52 | 18 |
| 10 | 14 | | 20 | 4 | 19 | 3 | 20 | 4 | 8 | | 5 | | 24 | 23 | 60 | 10 |
| 11 | 13 | 1 | 20 | 4 | 21 | 1 | 21 | 3 | 8 | | 5 | | 27 | 20 | 65 | 5 |
| 12 | 12 | 2 | 21 | 3 | 20 | 2 | 21 | 3 | 8 | | 4 | 1 | 27 | 20 | 63 | 7 |
| 13 | 12 | 2 | 18 | 6 | 19 | 3 | 21 | 3 | 8 | | 4 | 1 | 27 | 20 | 62 | 8 |
| 14 | 9 | 5 | 6 | 18 | 7 | 15 | 8 | 16 | 6 | 2 | 3 | 2 | 9 | 38 | 20 | 50 |
| 15 | 14 | | 24 | | 22 | | 22 | 2 | 8 | | 5 | | 42 | 5 | 67 | 3 |
| 16 | 9 | 5 | 20 | 4 | 19 | 3 | 18 | 6 | 8 | | 4 | 1 | 24 | 23 | 56 | 14 |
| 17 | 13 | 1 | 24 | | 22 | | 22 | 2 | 8 | | 5 | | 42 | 5 | 69 | 1 |
| Total | 204 | 34 | 347 | 61 | 333 | 41 | 343 | 65 | 134 | 2 | 71 | 14 | 486 | 313 | 999 | 187 |

Some congeners are more readily found than others. Table 81 illustrates the success of detection by congener.

Table 81. Success in detecting congeners.

| PARAM | Order | Detection success rate |
|---------------------|-------|------------------------|
| 2,3,7,8-TCDD | 1 | 56% |
| 1,2,3,7,8-PeCDD | 2 | 78% |
| 1,2,3,4,7,8-HxCDD | 3 | 76% |
| 1,2,3,6,7,8-HxCDD | 4 | 86% |
| 1,2,3,7,8,9-HxCDD | 5 | 86% |
| 1,2,3,4,6,7,8-HpCDD | 6 | 98% |
| OCDD | 7 | 100% |
| 2,3,7,8-TCDF | 8 | 89% |
| 1,2,3,7,8-PeCDF | 9 | 74% |
| 2,3,4,7,8-PeCDF | 10 | 79% |
| 1,2,3,4,7,8-HxCDF | 11 | 84% |
| 1,2,3,6,7,8-HxCDF | 12 | 82% |
| 2,3,4,6,7,8-HxCDF | 13 | 80% |
| 1,2,3,7,8,9-HxCDF | 14 | 64% |
| 1,2,3,4,6,7,8-HpCDF | 15 | 95% |
| 1,2,3,4,7,8,9-HpCDF | 16 | 73% |
| OCDF | 17 | 96% |

Samples from some harbor areas were richer in detections than others (Table 82). This reflects concentration, sampling diligence (liters of water filtered in the field), and, in the case of the tributary samples, competence of the labs. Many of the major tributary samples were analyzed by labs that gave high detection limits and, therefore, are more likely to fail to detect the analytes.

Table 82. Detection success by harbor area.

| area | detection success rate |
|---------------|------------------------|
| Amb-clean | 86% |
| Amb-Hud. | 85% |
| Amb-Kills | 89% |
| Amb-Non_Kills | 84% |
| CSO | 99% |
| Ind. Effluent | 84% |
| Tributaries | 61% |
| WPCF | 84% |

Sample Data

Table 83 shows average TEQs by site in two ways. The first (WHO98) uses the WHO98 TEF. The second (NYS WQS) uses WHO94 and the bioaccumulation factors as specified by the NYS Water Quality Standard for dioxin for protection of humans eating fish. The water quality standard for this purpose is 0.0006 pg/L and is exceeded by every sample. Table 83 also shows average instantaneous sample loads from those sites having defined discharges. The loads are in milligrams of WHO98 TEQ/hr by site. The major tributaries are seen dominating the loading. However, the Passaic and Hackensack Rivers may be putting large amounts of TEQ into the harbor as well. DEC sampling of those rivers were in portions greatly influenced by tides.

Table 83. Average TEQs (pg/L), using WHO98 and the NYS WQS (WHO94 plus BAF), and average instantaneous TEQ (WHO98) loads (mg/hr).

| Sample | WHO98 | NYS WQS | mg/hr |
|--|--------|---------|-------|
| Ambient-clean: Long Island Sound | 0.039 | 0.026 | NC |
| Ambient-clean: New York Bight | 0.0069 | 0.0065 | NC |
| Ambient-Hudson: Haverstraw Bay | 0.43 | 0.3 | NC |
| Ambient-Hudson: Hudson R. at Poughkeepsie | 1.9 | 0.91 | NC |
| Ambient-Hudson: Hudson R. below Kingston | 0.14 | 0.081 | NC |
| Ambient-Hudson: Hudson R. below Tappan Zee | 0.76 | 0.55 | NC |
| Ambient-Hudson: Hudson R. South of Harlem R. | 0.48 | 0.35 | NC |
| Ambient-Kills: Hackensack R., Mid-Tidal | 2.3 | 1.8 | NC |
| Ambient-Kills: Hackensack R., Mouth | 2.1 | 1.7 | NC |
| Ambient-Kills: Newark Bay | 1.3 | 1.1 | NC |
| Ambient-Kills: Northern Arthur Kill | 1.6 | 1.4 | NC |

Table 83 continued.

| sample | WHO98 | NYS WQS | mg/hr |
|---|-------|---------|-------|
| Ambient-Kills: Passaic R., Mouth, Bottom | 3.5 | 3.1 | NC |
| Ambient-Kills: Passaic R., Mouth, Surface | 11 | 10 | NC |
| Ambient-Kills: Passaic River, Mid-Tidal | 11 | 10 | NC |
| Ambient-Non_Kills: Jamaica Bay | 0.17 | 0.081 | NC |
| Ambient-Non_Kills: Lower Bay | 0.11 | 0.087 | NC |
| Ambient-Non_Kills: Lower East R. | 0.31 | 0.21 | NC |
| Ambient-Non_Kills: Raritan Bay | 0.17 | 0.12 | NC |
| Ambient-Non_Kills: Upper Bay | 0.29 | 0.23 | NC |
| Ambient-Non_Kills: Upper East R. | 0.15 | 0.1 | NC |
| CSO: 26 th Ward, High Side | 19 | 4.2 | 38 |
| CSO: 26 th Ward, Low Side | 6.5 | 2.7 | 13 |
| CSO: Bowery Bay High Side | 17 | 6.3 | 36 |
| CSO: Hunts Point Influent | 7.5 | 2.1 | 18 |
| CSO: Jamaica Influent | 9.2 | 2.5 | 45 |
| CSO: Manhattan Grit Chamber | 3 | 1.7 | 4.9 |
| CSO: North River Influent | 8.6 | 4.2 | 6.8 |
| CSO: Red Hook Influent | 17 | 6.5 | 9.8 |
| Industrial effluent: Clean Waters of New York | 0.02 | 0.013 | NC |
| Industrial effluent: FK Plant Effluent | 0.6 | 0.19 | 0.053 |
| Major tributary: Hudson R. (Pleasantdale) | 0.22 | 0.092 | 430 |
| Major tributary: Mohawk R. (Cohoes) | 0.24 | 0.12 | 420 |
| Major tributary: Wallkill (New Paltz) | 0.43 | 0.17 | 200 |
| Minor tributary: Bronx River | 0.29 | 0.11 | 2.5 |
| Minor tributary: Gowanus Canal | 0.25 | 0.17 | NC |
| Minor tributary: Saw Mill River (Yonkers) | 0.23 | 0.089 | 1.2 |
| TRK: Mill Creek at Arthur Kill Rd | 11 | 7.7 | NC |
| WPCF: 26 th Ward | 0.31 | 0.1 | 3 |
| WPCF: Bowery Bay | 0.14 | 0.067 | 2.8 |
| WPCF: Coney Island | 0.081 | 0.025 | 1.3 |
| WPCF: Hunts Point | 0.94 | 0.29 | 25 |
| WPCF: Newtown Creek | 0.38 | 0.14 | 22 |
| WPCF: Oakwood Beach | 0.17 | 0.069 | 0.82 |
| WPCF: Owls Head | 0.13 | 0.058 | 2.4 |
| WPCF: Port Richmond | 0.28 | 0.087 | 3.5 |
| WPCF: Red Hook | 0.096 | 0.043 | 0.56 |
| WPCF: Rensselaer | 0.33 | 0.1 | 0.89 |
| WPCF: Rockaway | 0.24 | 0.11 | 0.84 |
| WPCF: Tallman Island | 0.12 | 0.041 | 1.1 |
| WPCF: Wards Island | 0.066 | 0.033 | 2.3 |

Relative abundances of the 17 dioxin congeners (in WHO98 TEQ units) are shown in the following 42 figures. Only samples where there was a small difference (less than 10%) between assigning values of zero or half the detection limit are shown. Non-detected congeners were, in the figures, assigned a value of zero. In each figure, the horizontal axis lists the dioxin and furan congeners (see Tables 76 or 78) and the vertical axis is the relative abundance of the congeners to total TEQ (WHO98). The legend gives site abbreviations, date, and total TEQ.

Ambient-Clean

Figures 32 and 33 show congener distributions from the two background stations, Long Island Sound and the New York Bight.

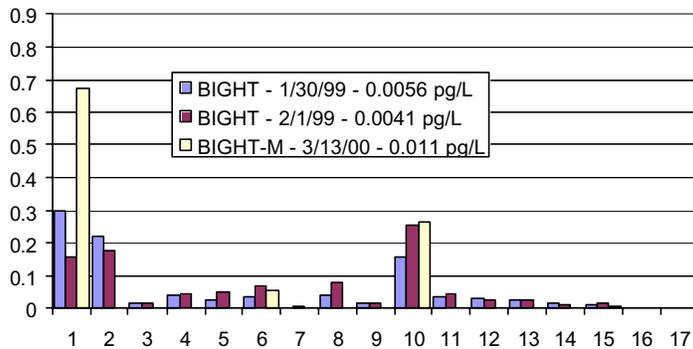


Figure 32. Three New York Bight samples show low total concentrations and dominance by 2,3,7,8-TCDD and 2,3,4,7-PeCDF.

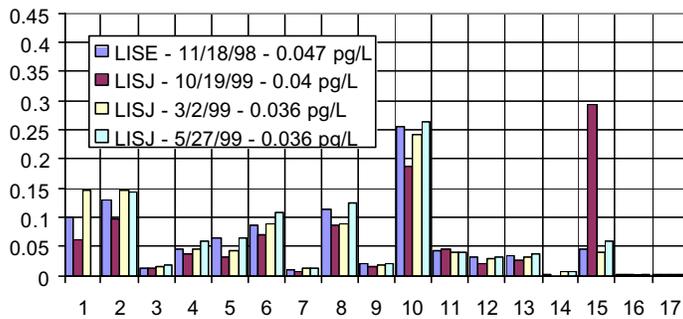


Figure 33. Long Island Sound off Eaton's Point (LISE), and Port Jefferson (LISJ). Cong. 10 is usually dominant. LIS samples have significantly more TEQ than Bight samples and a different distribution of congeners. In no other sample is cong. 15 so important.

Hudson River

Figures 34 to 44 follow the Hudson from the head of tide at Pleasantdale to the Lower Bay.

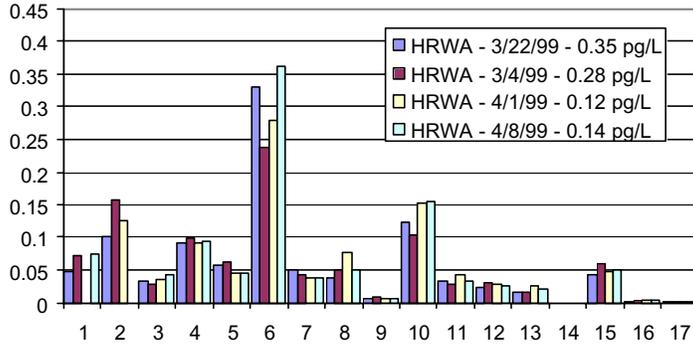


Figure 34. Hudson River at Pleasantdale samples have a fingerprint dominated by congener 6. 2,3,7,8-TCDD is a relatively minor source of TEQ.

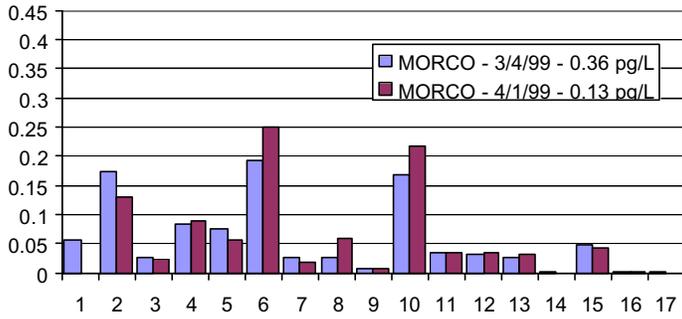


Figure 35. Mohawk River at Cohoes has concentrations similar to those in the upper Hudson. Congener 6 may be less important here.

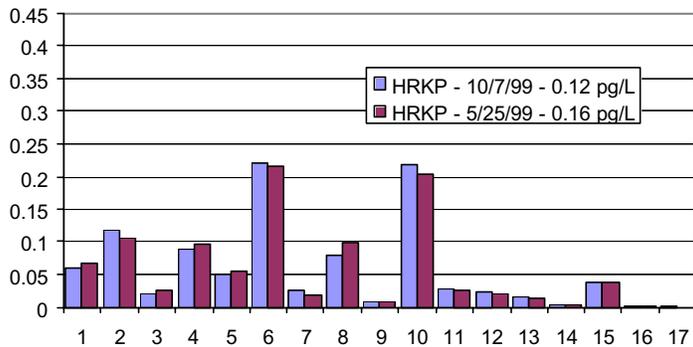


Figure 36. Ambient Hudson River samples collected between Kingston and Poughkeepsie are very similar to the Mohawk patterns. Congener 2 is less abundant at this site.

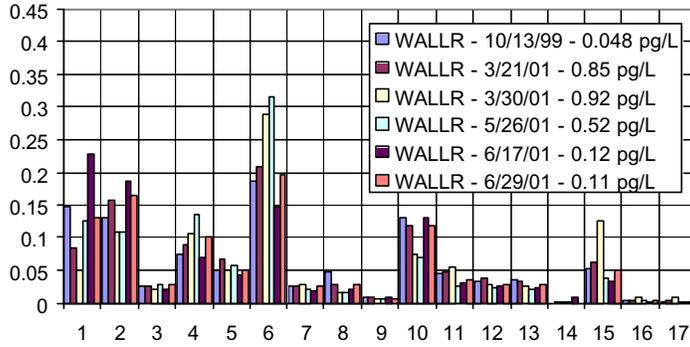


Figure 37. Wallkill at New Paltz shows an increased contribution by 1,2,3,7,8-PeCDD. This congener (#2) may be associated with municipal wastewater. Some total concentrations are higher than those from most ambient samples.

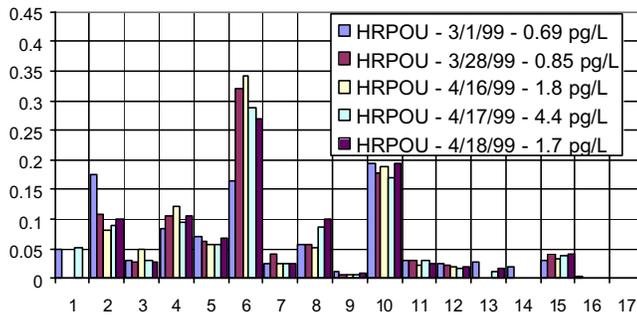


Figure 38. Hudson River at Poughkeepsie shows a typical upper Hudson River pattern (congeners 6 and 10) but some high concentrations.

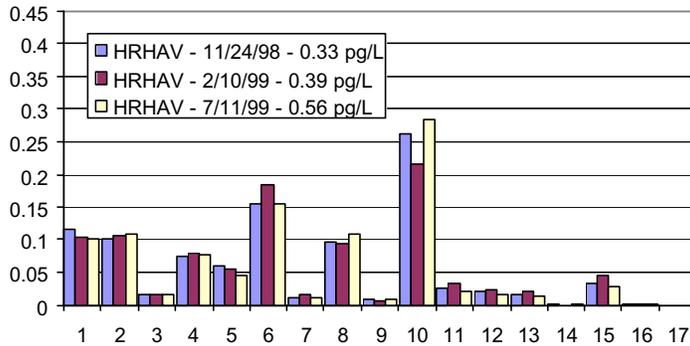


Figure 39. Hudson River, Haverstraw Bay, has a pattern dominated by 2,3,4,7,8-PeCDF. There may be a source of this congener in the Hudson. 2,3,7,8-TCDD begins to increase.

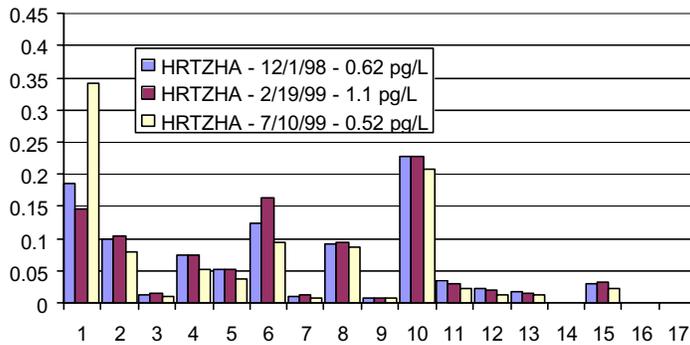


Figure 40. Hudson River, Tappan Zee Bridge to Harlem River, shows more clearly the impact of 2,3,7,8-TCDD.

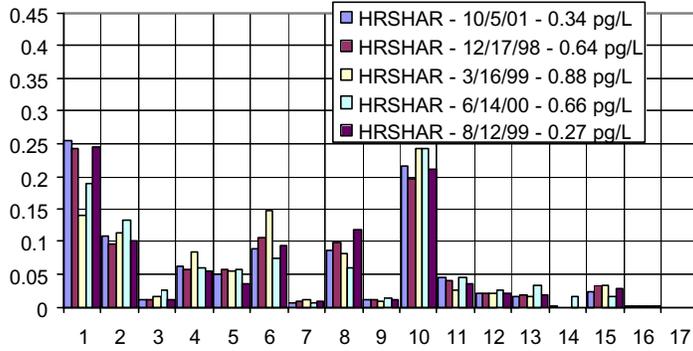


Figure 41. Hudson River, Harlem River to Battery, shows much more contribution by 2,3,4,7-TCDD.

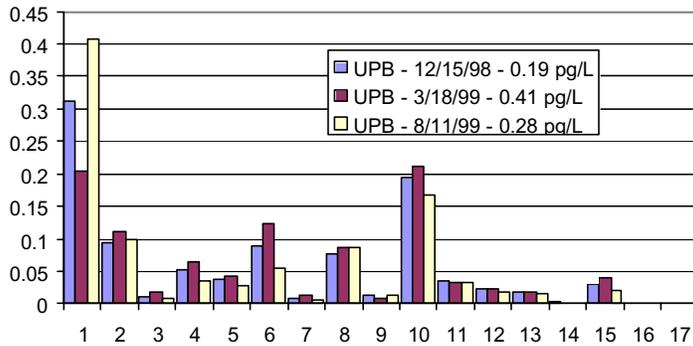


Figure 42. Upper Bay samples begin to be dominated by 2,3,7,8-TCDD.

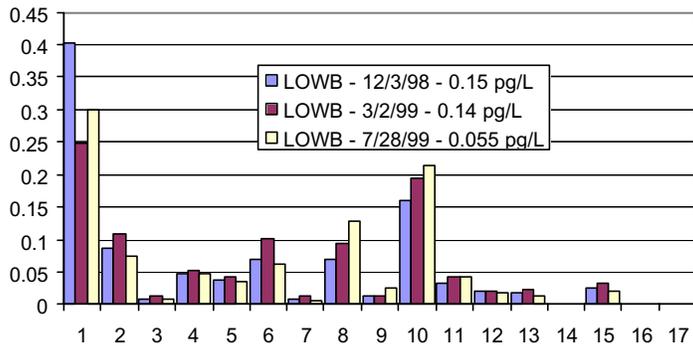


Figure 43. Lower Bay samples have the same congener pattern as the Upper Bay but lower concentrations due to dilution.

Ambient-Kills

Figures 44 to 53 follow the western side of the harbor up from Raritan Bay to the mid-tidal Passaic River.

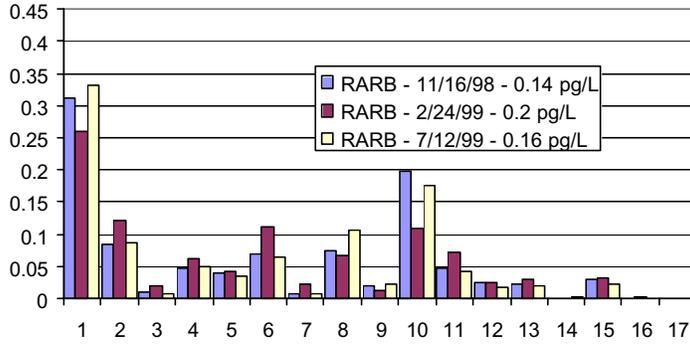


Figure 44. Raritan Bay samples again have a similar pattern but concentrations are higher.

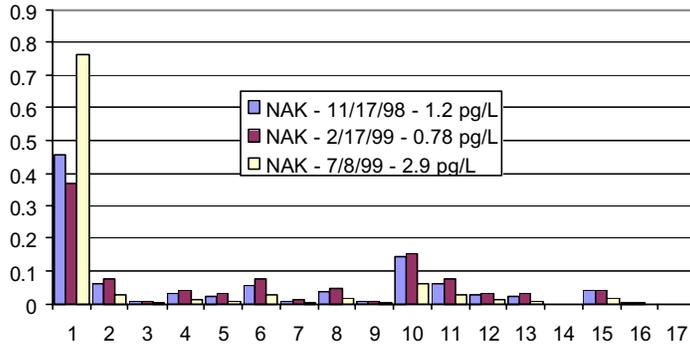


Figure 45. Northern Arthur Kill. All the Ambient-Kills samples are strongly dominated by contributions from 2,3,7,8-TCDD. The TEQ fingerprint in the Arthur Kill is affected by tides.

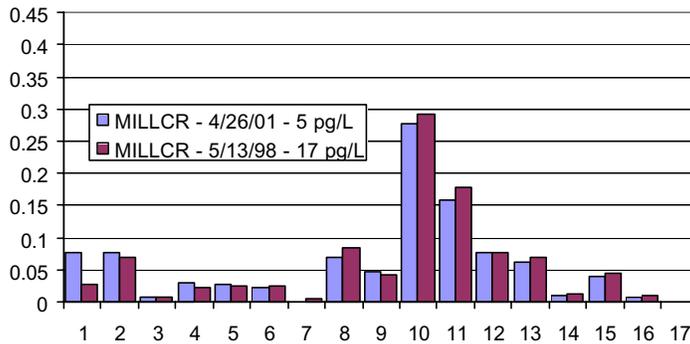


Figure 46. Mill Creek flows into the Arthur Kill and received wastewaters from a facility that incinerated obsolete electronics. TEQ concentrations are high but reflect the capture of bottom sediment. Unlike the Arthur Kill, patterns are dominated by congeners 10 and 11.

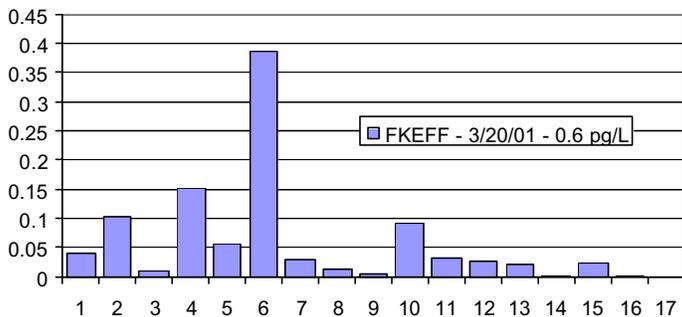


Figure 47. Leachate from the Fresh Kills Landfill is treated and discharged to the Arthur Kill. The pattern is dominated by congener 6.

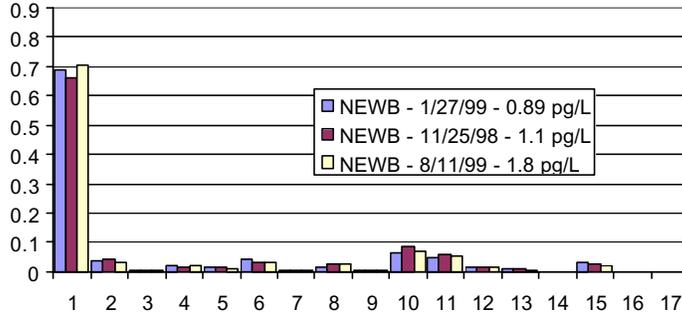


Figure 48. Newark Bay is consistently dominated by 2,3,7,8-TCDD.

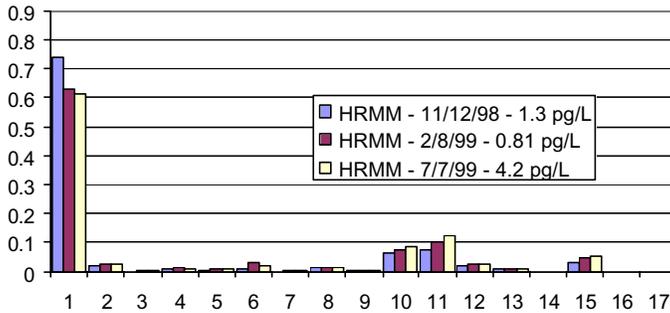


Figure 49. Hackensack River, mouth.

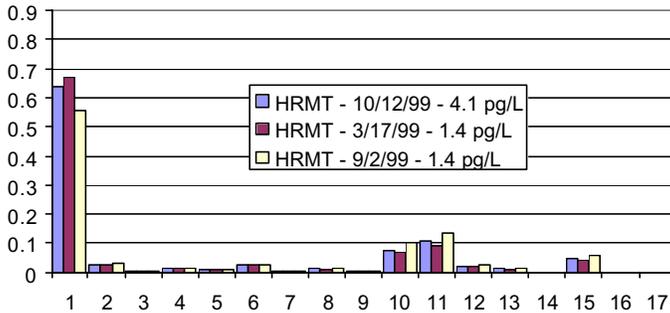


Figure 50. Hackensack River, mid tidal. The relative contribution of 2,3,7,8-TCDD is less here than at the mouth and less than in Newark Bay.

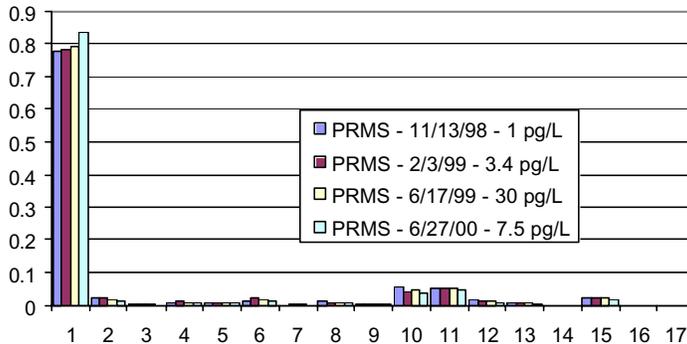


Figure 51. Passaic River, mouth, surface. The Passaic River appears to be the source of 2,3,7,8-TCDD.

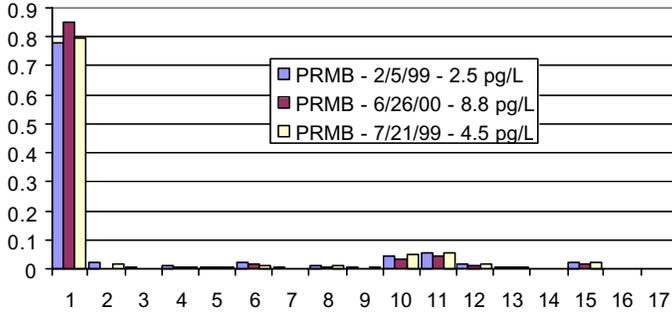


Figure 52. Passaic River, mouth, bottom meter.

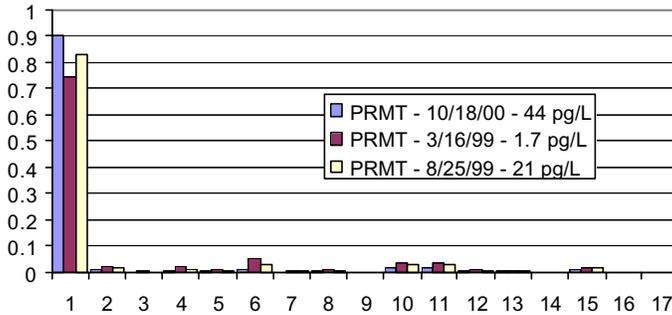


Figure 53. Passaic River, mid-tidal. 2,3,7,8-TCDD abundances and total TEQs are affected by tides. The high proportions of 2,3,7,8-TCDD and high concentrations occur on the flood tide.

Ambient- East River and the Minor Tributaries

Figures 54 to 58 show congener patterns in the East River and at Gowanus Canal, Saw Mill River, and Bronx River.

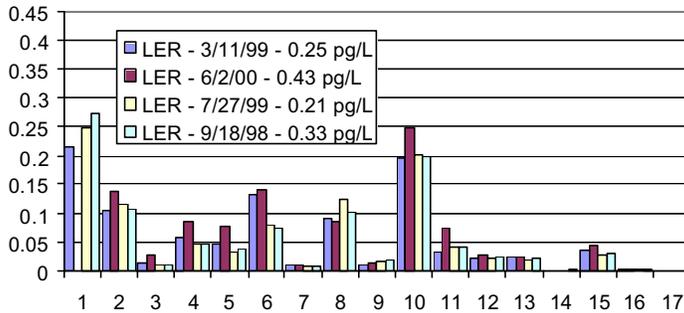


Figure 54. Lower East River. Patterns are similar to those in the Hudson south of Harlem River. 2,3,7,8-TCDD is less important here than in the Upper Bay.

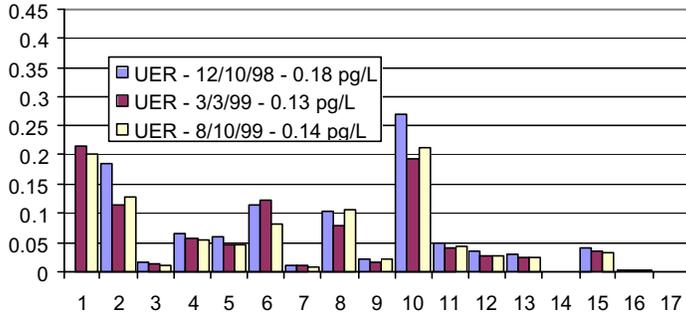


Figure 55. Upper East River. Total TEQ concentrations are lower than in the lower East River and cong. 10 is getting more important relative to 2,3,7,8-TCDD. Cong 2 seems to be associated with WPCFs.

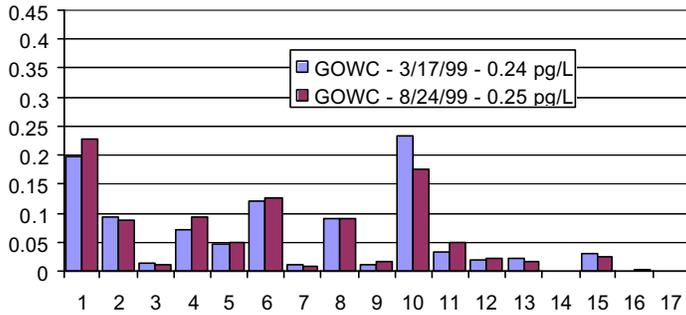


Figure 56. Gowanus Canal.

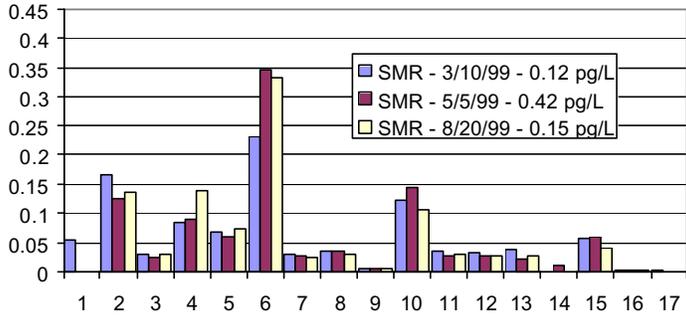


Figure 57. Saw Mill River. Both Saw Mill and the Bronx River, medium sized urban streams, have similar patterns with congeners 6, 10, and 2 dominant.

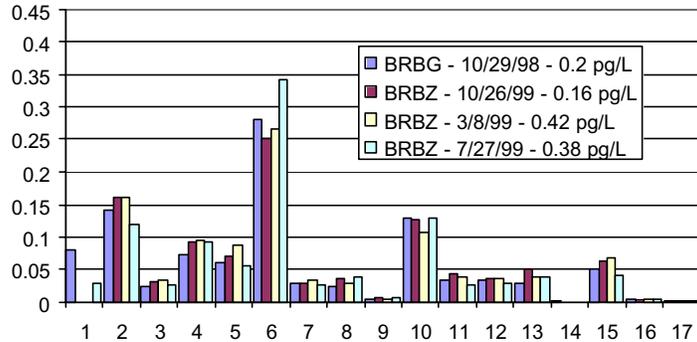


Figure 58. Bronx River at Botanical Garden (BRBG) and below the Bronx Zoo (BRBZ).

CSOs (Wet Weather Influents to Wastewater Treatment Plants)

Figures 59 and 60 show congener patterns from CSOs.

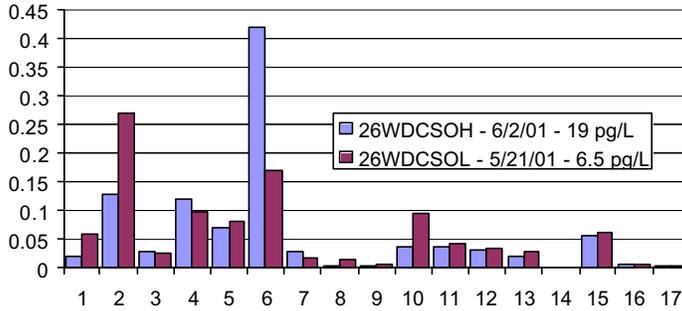


Figure 59. 26th Ward wet weather influents (CSO surrogates) are dominated by congeners 6 and 2.

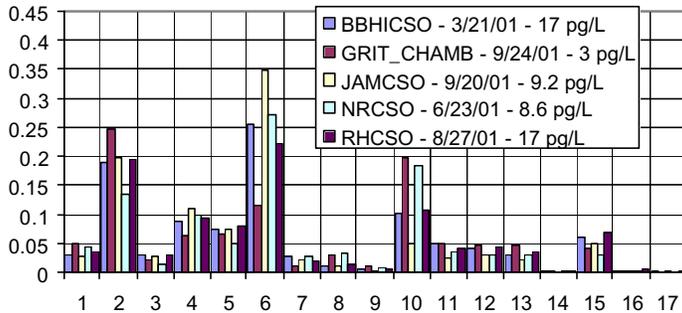


Figure 60. Wet weather influents at Bowery Bay (BB), Wards Island (Grit Chamber), Jamaica (JAM), North River (NR), and Red Hook (RH) show dominance by cong. 6, 2, and 10.

WPCF Final Effluents

Figures 61 to 69 show congener patterns from final treated effluents at wastewater treatment plants.

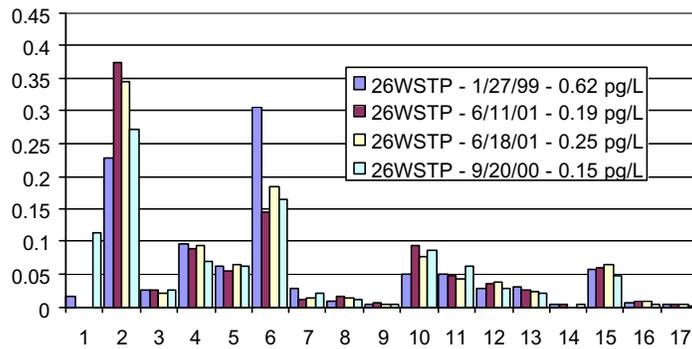


Figure 61. 26th Ward WPCF final treated effluents are dominated by congeners 2 and 6.

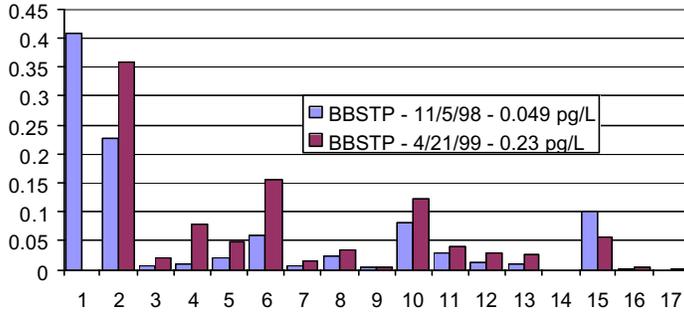


Figure 62. Bowery Bay WPCF effluent had a significant contribution from 2,3,7,8-TCDD but the total TEQ, 0.49 pg/L, was very low.

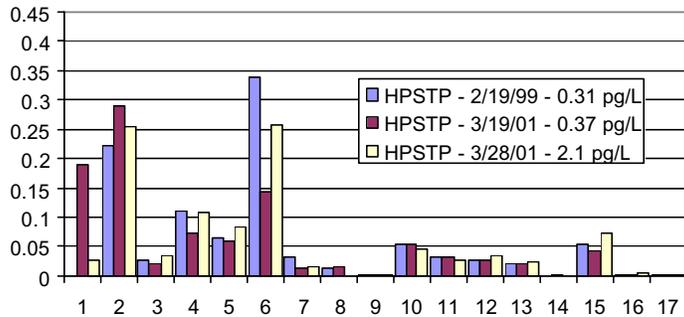


Figure 63. Hunts Point WPCF effluent also had a significant amount of 2,3,7,8-TCDD on one out of three samples.

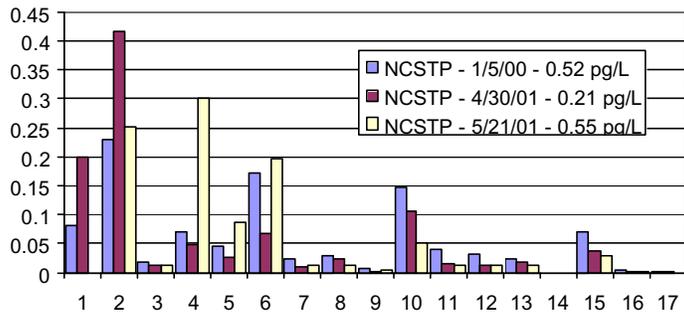


Figure 64. Newtown Creek WPCF effluent showed an unusual contribution from congener 4.

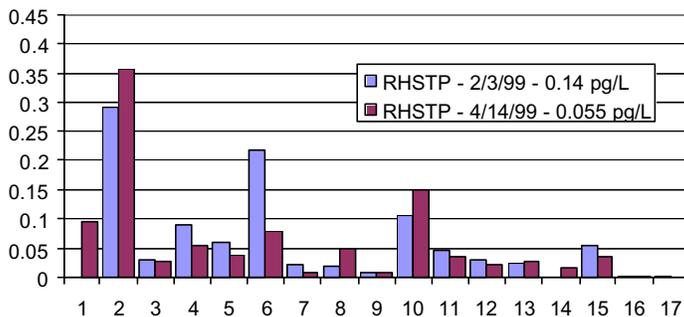


Figure 65. Red Hook WPCF effluent was, as usual, dominated by congener 2.

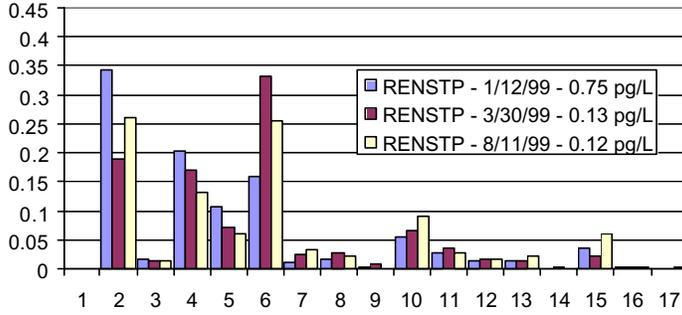


Figure 66. Rensselaer WPCF effluent is dominated by congeners 2, 6, and 4.

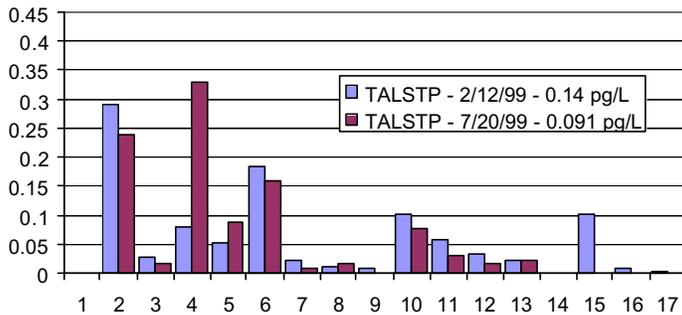


Figure 67. Tallman Island WPCF effluent is dominated by congeners 2, 4, and 6.

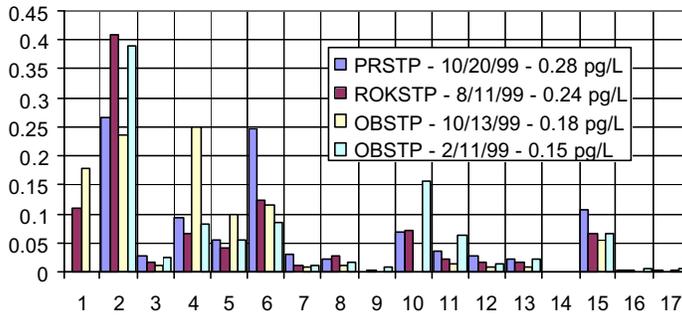


Figure 68. Samples of effluent from three facilities, Port Richmond (PR), Rockland County (ROK), and Oakwood Beach (OB) show typical abundances of congeners 2, 4, and 6.

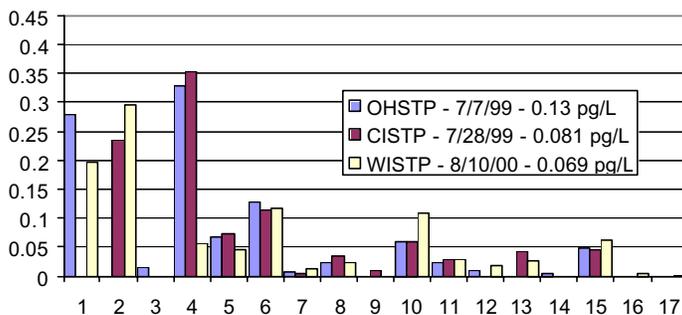


Figure 69. Samples of effluent from three facilities, Owls Head (OH), Coney Island (CI), and Wards Island (WI), have lower total TEQ concentrations. OH and WI show considerable 2,3,7,8-TCDD contributions.

Sludges and Biosolids

Figures 70 to 73 show congener patterns from wastewater treatment plant sludges and biosolids.

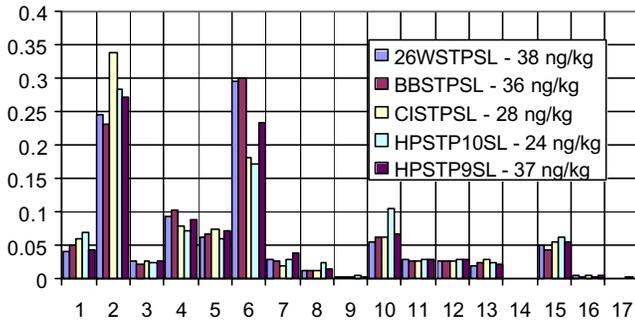


Figure 70. De-watered sludge samples (biosolids) composited daily during February, 2001 from 26th Ward (26W), Bowery Bay (BB), and two places at Hunts Point (HP), show dominance by congeners 2 and 6.

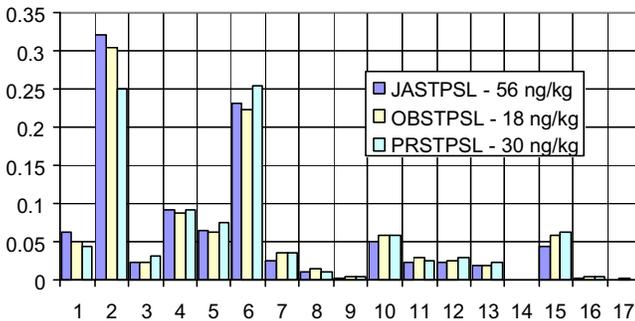


Figure 71. Biosolids samples from Jamaica (JA), Oakwood Beach (OB), and Port Richmond (PR) again show dominance by congeners 2 and 6.

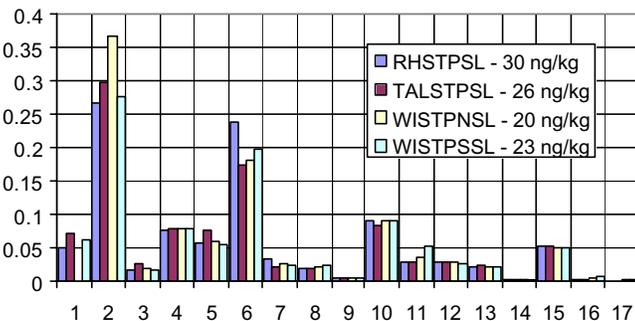


Figure 72. Biosolids samples from Red Hook (RH), Tallman Island (TAL), and two sites at Wards Island (WI), still show dominance by congeners 2 and 6.

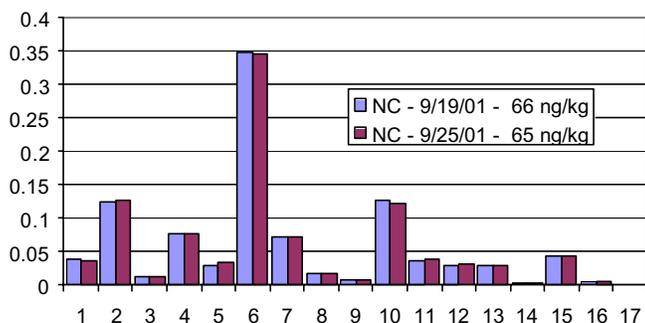


Figure 73. Sludge samples (not dewatered) were taken from the Newtown Creek (NC) facility as part of the investigation of the World Trade Center disaster. NC receives wastes from southern Manhattan. Patterns and total TEQ from the two dates were almost identical but different from the other sludges in the greater proportion of congener 6.

Dioxin congener patterns show considerable variation at various places in the system. Ambient samples from the Kills (the western harbor) all have most of the TEQ contributed by 2,3,7,8-TCDD. This effect spills into the Upper and Lower Bay and may also be seen up the Hudson and East River. However, the major sources of TEQ, the tributaries and urban waters from CSOs or treatment plants, usually have other congeners that are larger contributors.

Congeners 6 and 10 (1,2,3,4,6,7,8-HpCDD and 2,3,4,7,8-PeCDF) appear important in the Hudson at the head of tide down to the Harbor. There may be sources of congener 10 in the lower Hudson. Congeners 2 (1,2,3,7,8-PeCDD), 4 (1,2,3,6,7,8-HxCDD), and 6 may be diagnostic of urban wastewater. They are usually dominant in raw and final wastewater. De-watered sludges have fairly uniform patterns with congener 2 (29% of total TEQ on average), and congener 6 (22% of total TEQ) accounting for much of the total. Ambient waters or rivers receiving considerable treated wastewater also have high percentage contributions from these congeners. We do not know their sources. Some wastewater samples contain 2,3,7,8-TCDD but again, the sources are unknown.

Pesticides

Pesticides were analyzed with TOPS from sites throughout the harbor. Table 84 lists the analytes, the most stringent NYSDEC ambient water quality standards, number of analyses conducted by the CARP water program, and homologues under which one or more individual analytes may be grouped.

Table 84. CARP pesticides, functional groups (homologues), NYSDEC water quality standards, and number of analyses.

| Analyte | Homologues | WQS, ug/L | Number of Analyses |
|--------------------------|-----------------|-----------|--------------------|
| 2,4'-DDD | TDDT | | 601 |
| 2,4'-DDE | TDDT | | 597 |
| 2,4'-DDT | TDDT | | 594 |
| 4,4'-DDD | TDDT | 0.00008 | 604 |
| 4,4'-DDE | TDDT | 0.000007 | 604 |
| 4,4'-DDT | TDDT | 0.00001 | 602 |
| Chlordane, alpha (cis) | TChlordane | 1 | 601 |
| Chlordane, gamma (trans) | TChlordane | 1 | 601 |
| Chlordane, oxy- | TChlordane | | 596 |
| Nonachlor, cis- | TChlordane | | 597 |
| Nonachlor, trans- | TChlordane | | 599 |
| Aldrin | Aldrin/Dieldrin | 0.001 | 596 |
| Dieldrin | Aldrin/Dieldrin | 0.0000006 | 603 |
| Endrin | Endrin | 0.002 | 597 |
| Endrin aldehyde | Endrin | | 597 |
| Endrin ketone | Endrin | | 597 |
| HCH, alpha | THCH | 0.002 | 603 |
| HCH, beta | THCH | 0.007 | 596 |
| HCH, delta | THCH | 0 | 47 |
| HCH, gamma | THCH | 0.008 | 604 |
| Heptachlor | Heptachlor | 0.0002 | 596 |
| Heptachlor epoxide | Heptachlor | 0.0003 | 599 |
| Endosulfan, alpha | Endosulfan | 0.001 | 597 |
| Endosulfan, beta | Endosulfan | 0.001 | 597 |
| Endosulfan sulfate | Endosulfan | | 603 |
| Hexachlorobenzene | HCB | 0.00003 | 604 |
| Methoxychlor | Methoxychlor | 0.03 | 601 |
| Mirex | Mirex | 0.000001 | 596 |

Data Quality

Data quality is a complex evaluation. Tables 85 and 86 show for each pesticide the number of non-detections (ND); the number of occasions where the recovered mass was more than 10 times the detection level (GD, “Good DL”); occasions where the percent recovery in the associated sample delivery group (SDG) was either greater than 150% or less than 50% (GDV, “Good DL, LCS Violation”); occasions where the recovered mass was not greater than 10 times the detection level (HD, “High DL”); occasions where detection limits were relatively high and where the percent recoveries were out of bounds (HDV, “High DL, LCS Violation”); occasions where analyte was detected in the associated method blank but at levels less than one fifth of the analyte (NSB, “Non-Sig Blank”); occasions where blanks were not significant but the recoveries were out of bounds (NSBV, Non-Sig Blank, LCS Violation”); and lastly, occasions where blanks were not less than one fifth the mass of the sample (SB, “Sig. Blanks”). These tables allow evaluation of the objective of consistent detection of all analytes at all sites in both media.

Table 85. Number of analyses in each category. Glass fiber cartridges.

| PARAM | ND | GD | GDV | HD | HDV | NSB | NSBV | SB | Total |
|-------------------------|------|------|-------|------|-------|------|-------|------|-------|
| 2,4'-DDD | 17 | 144 | 2 | 34 | | | | | 197 |
| 2,4'-DDE | 51 | 65 | 1 | 79 | 1 | | | | 197 |
| 2,4'-DDT | 33 | 111 | 2 | 51 | | | | | 197 |
| 4,4'-DDD | 16 | 143 | | 16 | | 20 | 2 | | 197 |
| 4,4'-DDE | 15 | 156 | | 21 | | 3 | 2 | | 197 |
| 4,4'-DDT | 24 | 140 | 2 | 26 | | 5 | | | 197 |
| Aldrin | 71 | 37 | 1 | 72 | 1 | 17 | | | 199 |
| HCH, alpha | 64 | 21 | | 104 | 2 | 6 | | | 197 |
| HCH, beta | 80 | 16 | | 85 | 2 | 14 | | | 197 |
| HCH, delta | 13 | | | 3 | | | | | 16 |
| HCH, gamma | 40 | 54 | | 66 | | 34 | 2 | 1 | 197 |
| Chlordane,alpha (cis) | 16 | 113 | 2 | 26 | | 39 | | 1 | 197 |
| Chlordane,gamma (trans) | 12 | 125 | 2 | 26 | | 32 | | | 197 |
| Chlordane,oxy- | 83 | 37 | 2 | 70 | | 5 | | | 197 |
| Dieldrin | 14 | 143 | 2 | 29 | | 11 | | 2 | 201 |
| Endosulfan sulfate | 33 | 37 | | 49 | | 35 | | 45 | 199 |
| Endosulfan, alpha | 133 | 13 | | 50 | 1 | | | | 197 |
| Endosulfan, beta | 116 | 16 | | 65 | | | | | 197 |
| Endrin | 111 | 13 | 3 | 69 | 5 | | | | 201 |
| Endrin aldehyde | 142 | 6 | | 49 | 2 | | | 3 | 202 |
| Endrin ketone | 101 | 17 | | 77 | 4 | | | | 199 |
| Heptachlor | 52 | 73 | 1 | 52 | | 18 | | 1 | 197 |
| Heptachlor epoxide | 31 | 108 | 8 | 46 | 1 | 5 | | | 199 |
| Hexachlorobenzene | 9 | 3 | | 14 | | 115 | 2 | 64 | 207 |
| Methoxychlor | 46 | 42 | 2 | 80 | | 29 | | 1 | 200 |
| Mirex | 38 | 72 | | 64 | | 23 | | | 197 |
| Nonachlor, cis- | 35 | 88 | 2 | 60 | | 12 | | | 197 |
| Nonachlor, trans- | 18 | 114 | 2 | 32 | | 34 | | 1 | 201 |
| Grand Total | 1414 | 1907 | 34 | 1415 | 19 | 457 | 8 | 119 | 5373 |
| | 18% | 24% | 0.43% | 18% | 0.24% | 5.8% | 0.10% | 1.5% | |

Table 86. Number of analyses in each category, XAD.

| PARAM | ND | GD | GDV | HD | HDV | NSB | NSBV | SB | Totals |
|-----------------------|------|------|-------|------|-------|------|-------|------|--------|
| 2,4'-DDD | 45 | 124 | | 115 | | 3 | | | 287 |
| 2,4'-DDE | 175 | 17 | | 93 | 2 | | | | 287 |
| 2,4'-DDT | 143 | 35 | | 102 | | | | | 280 |
| 4,4'-DDD | 27 | 165 | | 88 | | 3 | | 4 | 287 |
| 4,4'-DDE | 79 | 84 | | 119 | | 3 | | 2 | 287 |
| 4,4'-DDT | 106 | 72 | | 104 | | 2 | | 2 | 286 |
| Aldrin | 160 | 30 | | 91 | | 2 | | 6 | 289 |
| HCH, alpha | 20 | 183 | | 41 | | 43 | | | 287 |
| HCH, beta | 44 | 141 | | 77 | | 25 | | | 287 |
| HCH, delta | 21 | | | | | | | | 21 |
| HCH, gamma | 9 | 170 | | 40 | | 77 | | | 296 |
| Chlordane,alpha (cis) | 23 | 156 | 5 | 68 | | 45 | | | 297 |
| Chlordane,gamma (t) | 25 | 162 | | 76 | | 34 | | | 297 |
| Chlordane,oxy- | 133 | 38 | 4 | 105 | 4 | 1 | | 1 | 286 |
| Dieldrin | 11 | 214 | | 43 | | 22 | | | 290 |
| Endosulfan sulfate | 21 | 85 | | 86 | | 98 | | 6 | 296 |
| Endosulfan, alpha | 199 | 5 | | 84 | | | | | 288 |
| Endosulfan, beta | 193 | 9 | | 86 | | | | | 288 |
| Endrin | 173 | 16 | | 97 | 2 | 1 | | 1 | 290 |
| Endrin aldehyde | 226 | | | 64 | 7 | 4 | 1 | 1 | 303 |
| Endrin ketone | 108 | 44 | | 123 | | 13 | | | 288 |
| Heptachlor | 93 | 79 | | 88 | 1 | 25 | | 1 | 287 |
| Heptachlor epoxide | 38 | 165 | | 60 | | 24 | | 3 | 290 |
| Hexachlorobenzene | 17 | 4 | | 21 | 4 | 118 | | 138 | 302 |
| Methoxychlor | 70 | 49 | | 171 | | 9 | | | 299 |
| Mirex | 201 | 10 | | 76 | | 3 | | 6 | 296 |
| Nonachlor, cis- | 96 | 61 | 4 | 109 | 1 | 16 | | | 287 |
| Nonachlor, trans- | 53 | 125 | 4 | 87 | 1 | 20 | | | 290 |
| Grand Total | 2509 | 2243 | 17 | 2314 | 22 | 591 | 1 | 171 | 7868 |
| | 32% | 29% | 0.22% | 29% | 0.28% | 7.5% | 0.01% | 2.2% | |

Table 86 underestimates the true success of XAD in collecting pesticides. In 32% of the samples, the two XAD columns exposed in series were analyzed separately. Detection of an analyte on the first but not the second column should be recorded as a success. Since each analyte in each analysis had its own detection level, it is not possible to create a table like Table 86 for the lumped XAD data. Table 87 shows the overall success (detection versus non-detection) of pesticides by lumped XAD.

Table 87. Overall success rate of XAD in capturing pesticides. Number of analyses in each category.

| Analyte | ND | Detection | Success Rate |
|-------------------------|------|-----------|--------------|
| 2,4'-DDD | 28 | 207 | 88% |
| 2,4'-DDE | 130 | 105 | 45% |
| 2,4'-DDT | 108 | 127 | 54% |
| 4,4'-DDD | 16 | 219 | 93% |
| 4,4'-DDE | 47 | 188 | 80% |
| 4,4'-DDT | 69 | 166 | 71% |
| Aldrin | 122 | 113 | 48% |
| HCH, alpha | 17 | 218 | 93% |
| HCH, beta | 41 | 194 | 83% |
| HCH, delta | 21 | 0 | 0% |
| HCH, gamma | 6 | 229 | 97% |
| Chlordane,alpha (cis) | 21 | 214 | 91% |
| Chlordane,gamma (trans) | 24 | 211 | 90% |
| Chlordane,oxy- | 97 | 138 | 59% |
| Dieldrin | 11 | 224 | 95% |
| Endosulfan sulfate | 20 | 215 | 91% |
| Endosulfan, alpha | 152 | 83 | 35% |
| Endosulfan, beta | 141 | 94 | 40% |
| Endrin | 131 | 104 | 44% |
| Endrin aldehyde | 170 | 65 | 28% |
| Endrin ketone | 84 | 151 | 64% |
| Heptachlor | 74 | 161 | 69% |
| Heptachlor epoxide | 35 | 200 | 85% |
| Hexachlorobenzene | 20 | 215 | 91% |
| Methoxychlor | 56 | 178 | 76% |
| Mirex | 158 | 77 | 33% |
| Nonachlor, cis- | 72 | 163 | 69% |
| Nonachlor, trans- | 43 | 192 | 82% |
| Grand Total | 1914 | 4451 | 70% |

Some pesticides are more significant than others. The primary pesticides in CARP are the DDTs and dieldrin. Both of these appeared in more than 90% of the samples. Chlordane (and Nonachlor) data are also useable. Others, such as delta HCH, endrin aldehyde, and mirex were never or rarely quantitated. BHCs appear to be satisfactory but may have problems. New York Bight concentrations of BHCs are scarcely different from concentrations at sites where other analytes occur at concentrations orders of magnitude greater.

Sample Findings

Table 88, using data that have adequate detection limits and are not affected by blanks, shows average concentrations of the five significant pesticide homologues (Aldrin/Dieldrin, total heptachlor, total HCH, total Chlordane, and total DDT) by sample

type. The sample types are ambient water from Long Island Sound/New York Bight (AMB-clean); ambient Hudson River samples taken below Troy (AMB-Hudson); ambient water samples from the western part of New York Harbor (AMB-Kills); ambient water samples from other parts of the harbor (AMB-Non Kills); wet weather influent to wastewater treatment plants (CSO); high and base-line flow event samples from the upper Hudson, Mohawk, and Wallkills (Major-TRIB); samples from the Bronx River, Sawmill Creek (Westchester), and the Gowanus Canal (Minor-TRIB); and final effluent samples from wastewater treatment plants (POTW). Sample types with the highest mean concentrations are highlighted.

Table 88. Mean pesticide concentration by sample type. Maxima are highlighted (good data).

| | Aldrin/Dieldrin | HCB | Heptachlor | THCH | Tchlordan | TDDT |
|-------------------|-----------------|-------|------------|------|-----------|------|
| CSO | 2.8 | 2 | 2.3 | 2.9 | 80 | 48 |
| Landfill leachate | 0.94 | 0.26 | 0.5 | 1.3 | 1.7 | 9.7 |
| Major tributaries | 3.9 | 0.13 | 0.44 | 0.26 | 2.8 | 45 |
| WPCF | 0.95 | 0.22 | 0.31 | 7.3 | 1.5 | 1.4 |
| AMB-Kills | 0.61 | 0.26 | 0.34 | 1.5 | 0.84 | 2.8 |
| Minor tributaries | 0.88 | 0.059 | 0.38 | 0.95 | 1.1 | 0.43 |
| AMB-Non_Kills | 0.2 | 0.043 | 0.066 | 1.3 | 0.19 | 0.4 |
| AMB-Hudson | 0.34 | 0.03 | 0.074 | 0.78 | 0.099 | 0.66 |
| AMB-clean | 0.047 | 0 | 0.017 | 0.97 | 0.028 | 0.09 |

Concentrations of the pesticides are, with the exception of THCH, lowest in the areas though to be cleanest. Total BHCs show the least variation between sites. This is suspicious. THCH will not be discussed further.

Table 89 shows mean concentrations of the targeted pesticides in the tributaries. The highest concentrations of all the pesticides, except heptachlor, occurred in the Wallkill.

Table 89. Mean total pesticide concentrations in major and minor tributaries, ng/L (good data).

| | Aldrin/Dieldrin | HCB | Heptachlor | Tchlordan | TDDT |
|---------------------------------|-----------------|-------|------------|-----------|-------|
| Wallkill (New Paltz) | 5.6 | 0.28 | 0.63 | 3.5 | 82 |
| Bronx River, below zoo | 1 | 0.078 | 0.67 | 1.4 | 0.68 |
| Saw Mill River (Yonkers) | 1.3 | 0.072 | 0.37 | 1.3 | 0.35 |
| Bronx River at Botanical Garden | 0.16 | 0.024 | 0.03 | 0.65 | 0.49 |
| Gowanus Canal | 0.14 | 0.038 | 0.048 | 0.12 | 0.25 |
| Mohawk R. (Cohoes) | 0.041 | 0.016 | 0.0078 | 0.039 | 0.198 |
| Hudson R. (Pleasantdale) | 0.035 | 0.01 | 0.019 | 0.014 | 0.17 |

Wallkill Trackdown

In 1997, sampling was conducted at tributaries to the Hudson River using passive samplers (PISCES). PISCES contain hexane and a polyethylene window through which hydrophobic substances pass. Passive samples are only weakly quantitative but they have the advantage of quickly and easily integrating contaminants over a span of a few weeks. Calibrations based on membrane area and water temperature were derived for PCBs and then applied to the pesticides. Figure 74 summarizes the average “concentrations” of total DDT and dieldrin from the sites.

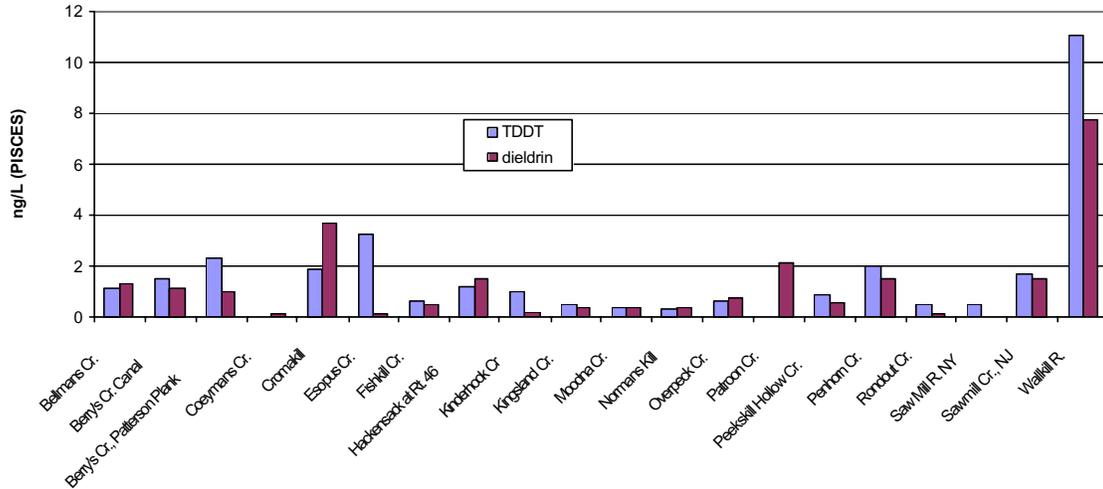


Figure 74. PISCES data, Spring of 1997.

These data pointed to the Wallkill as a pesticide source. Further investigation using PISCES, TOPS, and sediment samples, demonstrated that the pesticide source was an intensively farmed (onions, lettuce, and carrots) area around the Wallkill just north of New Jersey. This zone, called the Black Dirt, is a dried lakebed crossed by numerous drainage channels.

A sediment core taken near the mouth of the Wallkill (the Wallkill discharges to Rondout Creek and Roundout Creek empties into the Hudson River immediately below Kingston, New York) at Sturgeon Pool indicated that the highest DDT concentrations were on the surface (Figure 75).

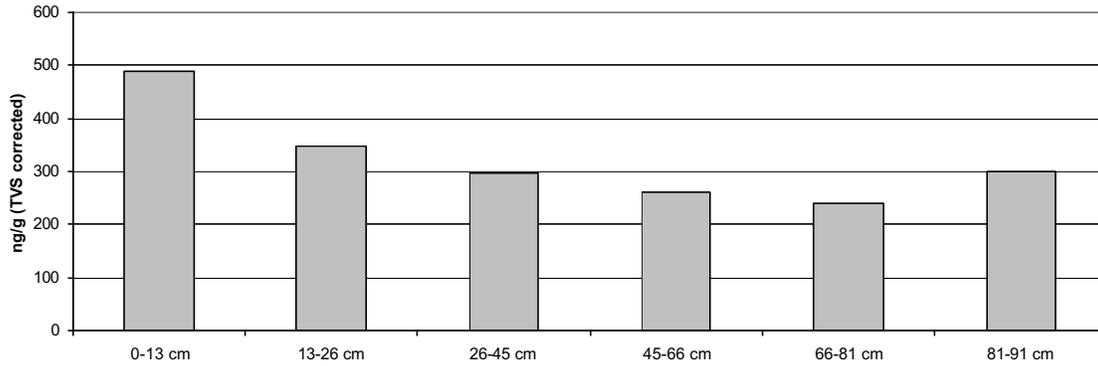


Figure 75. Total volatile solids corrected TDDT concentrations from a sediment core taken at Sturgeon Pool, Wallkill.

A soil sample taken in the Black Dirt shows that parent, unmetabolized p,p'-DDT was the most abundant species (Figure 76).

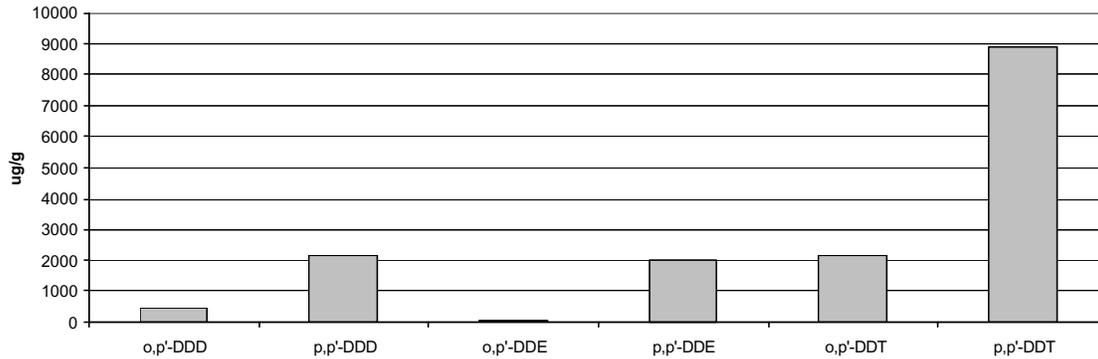


Figure 76. Individual DDTs in a Wallkill Black Dirt drainage ditch soil sample.

While it seems unlikely that DDT is currently being used, the data do not refute this notion.

Figure 77 compares the amounts of TDDT recovered from suspended solids versus XAD (dissolved) from the Wallkill samples. Almost all the TDDT, particularly during high flows, is associated with suspended sediment.

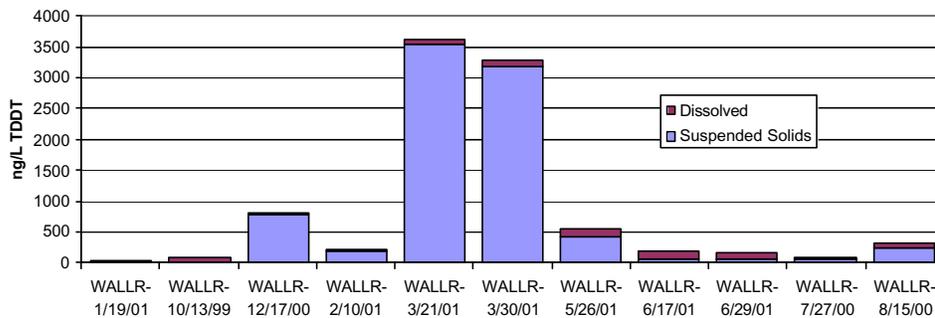


Figure 77. Stacked bars comparing TDDT in the dissolved and suspended sediment phases from Wallkill samples taken at New Paltz.

TOPS pesticide samples

Table 90 shows loadings in g/hr from events on the Hudson, Mohawk, and Wallkill for aldrin/dieldrin, heptachlor, total chlordane, and total DDT.

Table 90. Tributary event loadings (g/hr), good data.

| | Aldrin/Dieldrin | Heptachlor | Tchlordane | DDT |
|----------------------------------|-----------------|------------|------------|-------|
| Hudson-3/22/1999 - 3/23/1999 | 0.096 | 0.041 | | 0.75 |
| Hudson-3/4/1999 - 3/6/1999 | 0.081 | | | 0.49 |
| Hudson-4/1/1999 - 4/7/1999 | 0.04 | | | 0.052 |
| Hudson-4/8/1999 - 4/12/1999 | 0.049 | 0.031 | | 0.088 |
| Mohawk-3/4/1999 - 3/23/1999 | | | 0.12 | 0.41 |
| Mohawk-4/1/1999 - 4/7/1999 | 0.069 | 0.013 | 0.017 | 0.21 |
| Wallkill-1/19/2001 - 1/23/2001 | 0.05 | | | 0.054 |
| Wallkill-10/13/1999 - 10/27/1999 | 0.1 | 0.005 | 0.05 | 0.35 |
| Wallkill-12/17/2000 - 12/18/2000 | 4 | | 5.8 | 96 |
| Wallkill-2/10/2001 - 2/12/2001 | 0.28 | | 0.082 | 2.9 |
| Wallkill-3/21/2001 - 3/25/2001 | 12 | 0.76 | 4.4 | 230 |
| Wallkill-3/30/2001 - 4/2/2001 | 7.8 | 0.61 | 3.5 | 200 |
| Wallkill-5/26/2001 - 6/1/2001 | 0.98 | 0.11 | 0.75 | 7.6 |
| Wallkill-6/17/2001 - 6/19/2001 | 0.34 | 0.037 | 0.14 | 0.82 |
| Wallkill-6/29/2001 - 6/30/2001 | 0.22 | 0.025 | 0.074 | 0.45 |
| Wallkill-7/27/2000 - 7/28/2000 | 0.44 | | 0.035 | 2.3 |
| Wallkill-8/15/2000 - 8/17/2000 | 1.2 | | 0.38 | 9.9 |

Table 91 (above) indicates CSOs (sampled as wet weather influents to treatment plants) having the highest chlordane concentration. The highest chlordane concentration in CSOs came from Hunts Point in the Bronx .

Table 91. Average pesticide loads in raw wet weather influents to POTWs, ug/hr, good data.

| | Aldrin/Dieldrin | HCB | Heptachlor | Tchlordane | DDT |
|--------------------------------------|-----------------|------|------------|------------|-----|
| Hunts Point Influent | 8.2 | 16 | 29 | 1600 | 200 |
| Jamaica Influent | 23 | 4.2 | 13 | 130 | 130 |
| Bowery Bay High Side Interceptor | 9.9 | 3.4 | 6.5 | 180 | 110 |
| 26 th Ward CSO Low Side | 6.4 | 4.9 | 1.3 | 25 | 62 |
| 26 th Ward CSO, High Side | 2.5 | 2.7 | 0.72 | 15 | 65 |
| North River Influent | 1.8 | 0.57 | 0.76 | 17 | 64 |
| Manhattan Grit Chamber | 2.5 | 0.76 | 0.19 | 5.3 | 76 |
| Red Hook Influent | 2.1 | 0.85 | 0.41 | 15 | 47 |
| Bowery Bay Low Side Interceptor | 0.92 | | 0.47 | 6.6 | 8.9 |

The Hunts Point wet weather influent sample showed high concentrations for all the individual chlordane components:

Table 92. Chlordane components in the Hunts Point Influent sample.

| PARAM | Conc, ng/L |
|-------------------------|------------|
| Chlordane,alpha (cis) | 220 |
| Chlordane,gamma (trans) | 220 |
| Nonachlor, trans- | 180 |
| Nonachlor, cis- | 39 |
| Chlordane,oxy- | 0.082 |

Table 93 shows for each of the treatment plants the average pesticide loading in ug/hr from the treatment plants. The plants are listed in order of the summation of ranks of average pesticide loading.

Table 93. Average pesticide loads in final effluents, ug/hr., good data.

| | Aldrin/Dieldrin | HCB | Heptachlor | Tchlordan | DDT |
|-----------------------|-----------------|-----|------------|-----------|------|
| Newtown Creek WPCF | 20.5 | 6.1 | 10.1 | 67.3 | 129 |
| Owls Head WPCF | 22 | 4.9 | 11.5 | 60 | 30 |
| Hunts Point WPCF | 37 | 4.7 | 6.4 | 57 | 42 |
| Wards Island WPCF | 20.3 | 9.3 | 7.9 | 20.7 | 29.7 |
| Yonkers WPCF | 34.9 | 1.7 | 7.4 | 31.3 | 5 |
| Port Richmond WPCF | 8 | 7.2 | 5.4 | 26 | 10.5 |
| Bowery Bay WPCF | 14 | 2.7 | 3 | 14.9 | 27 |
| Jamaica WPCF | 11 | 1.7 | 3.7 | 17 | 12 |
| Coney Island WPCF | 15 | 1.7 | 4.1 | 13 | 12.4 |
| Tallman Island WPCF | 12.6 | 1.9 | 3.5 | 17.1 | 4.2 |
| 26th Ward WPCF | 7 | 3.3 | 1 | 8 | 20 |
| North River WPCF | 6 | 2 | 0.9 | 3 | 21 |
| Oakwood Beach WPCF | 2 | 1.2 | 3.5 | 14 | 0.5 |
| Rockaway WPCF | 2.9 | 0.4 | 1.7 | 4.1 | 2.3 |
| Rockland County WPCF | 2.2 | 1.1 | 0.7 | 3.5 | 1.5 |
| Red Hook WPCF | 1.6 | 0.4 | 0.5 | 2.2 | 4 |
| Poughkeepsie (C) WPCF | 1.8 | 0.2 | 0.1 | 0.9 | 2 |
| Rensselaer WPCF | 0.5 | 0.2 | 0.2 | 0.5 | 1.6 |

Examination of biosolids (sludge) places Hunts Point in the first rank in terms of total pesticide concentrations (Table 94). Sludges from different facilities are dewatered at Hunts Point so the concentrations there may be affected by discharges in other catchments.

There were two pesticide formulators in the Hunts Point catchment and two in the Owls Heads catchment. A fifth formulated pesticides in Manhattan and is served by North River WPCF.

Table 94. Pesticide concentrations in biosolids (ug/kg).

| Site Name | Aldrin/Dieldrin | Heptachlor | HCB | Tchlordan | DDT |
|---------------------|-----------------|------------|--------|-----------|---------|
| Hunts Point #9 | 74,000 | 5,400 | 16,000 | 230,000 | 290,000 |
| Tallman Island | 31,000 | 8,700 | 4,100 | 300,000 | 100,000 |
| Oakwood Beach | 41,000 | 12,000 | 6,100 | 290,000 | 62,000 |
| Port Richmond | 24,000 | 6,600 | 7,600 | 250,000 | 89,000 |
| Bowery Bay | 25,000 | 2,100 | 4,700 | 170,000 | 120,000 |
| Coney Island | 28,000 | 2,800 | 8,100 | 120,000 | 150,000 |
| Jamaica | 36,000 | 7,500 | 5,200 | 140,000 | 88,000 |
| Wards Island, South | 17,000 | 1,600 | 14,000 | 68,000 | 160,000 |
| Hunts Point, #10 | 13,000 | 1,000 | 6,000 | 91,000 | 150,000 |
| 26th Ward | 21,000 | 2,300 | 8,600 | 86,000 | 130,000 |
| Red Hook | 14,000 | 800 | 4,300 | 53,000 | 130,000 |
| Wards Island, North | 17,000 | 1,100 | 7,200 | 49,000 | 110,000 |

Ambient water samples show much lower concentrations than raw or treated wastewaters. The areas with the highest concentrations are the Kills (Western New York harbor) and the Hudson. Highest concentrations occurred in areas affected by former pesticide manufacturing in the Passaic/Hackensack Rivers and in the Arthur Kill.

Table 95. Average total pesticide concentrations at 13 ambient sites in the Hudson River and in NY/NJ Harbor (ng/L), good data.

| | Aldrin/Dieldrin | HCB | Heptachlor | Tchlordane | DDDT |
|--|-----------------|--------|------------|------------|-------|
| Passaic River, Mid-Tidal | 1.4 | 0.28 | 0.68 | 2 | 2.5 |
| Passaic R., Mouth, Bottom | 0.6 | 0.21 | 0.49 | 0.83 | 2.3 |
| Hackensack R., Mid-Tidal | 0.51 | 0.73 | 0.36 | 0.65 | 1 |
| Passaic R., Mouth, Surface | 0.49 | 0.26 | 0.19 | 1 | 2.5 |
| Northern Arthur Kill | 0.5 | 0.099 | 0.27 | 0.49 | 9.8 |
| Newark Bay | 0.45 | 0.12 | 0.18 | 0.5 | 1.2 |
| Hackensack R., Mouth | 0.43 | 0.14 | 0.088 | 0.27 | 0.53 |
| Raritan Bay | 0.32 | 0.068 | 0.13 | 0.26 | 0.88 |
| Lower East R., Brooklyn Br. To Hell Gate | 0.22 | 0.07 | 0.068 | 0.22 | 0.45 |
| Jamaica Bay | 0.19 | 0.1 | 0.081 | 0.12 | 0.099 |
| Upper Bay | 0.14 | 0.03 | 0.032 | 0.27 | 0.38 |
| Upper East R., Hell Gate to Throgs Neck | 0.2 | 0.0083 | 0.051 | 0.18 | 0.22 |
| Lower Bay | 0.13 | 0.01 | 0.037 | 0.067 | 0.3 |

Sediments

Pesticides concentrations appear to be usually higher in biosolids than in sediments. This may be due to the higher organic content of biosolids and that data adjusted for organic carbon would show less difference.

CARP sediment sampling also points to the Arthur Kill as an area of interest for DDT. Table 96 shows pesticide concentrations in PPB from surficial samples taken throughout the area.

Table 96. Pesticides in surficial sediment samples, ug/kg.

| | Aldrin/Dieldrin | HCB | Heptachlor | THCH | Tchlordan | DDT |
|-------------------|-----------------|-------|------------|------|-----------|------|
| Passaic R. | 40 | 36 | 6 | 5.4 | 370 | 840 |
| Arthur Kill | 18 | 62 | 0.4 | 2.7 | 88 | 3900 |
| Newtown Creek | 280 | 33 | 2.1 | 0.4 | 1600 | 1300 |
| Hackensack R. | 27 | 20.2 | 1.5 | 2 | 99 | 320 |
| Newark Bay | 10 | 8.7 | 0.14 | 1.1 | 47 | 350 |
| Raritan Bay | 3.9 | 2 | 0.15 | 0.6 | 16 | 240 |
| Harlem River | 3.4 | 1.8 | 0.095 | 0.31 | 17 | 81 |
| Hudson R. | 2.3 | 1.7 | | 0.52 | 6 | 50 |
| East River | 1.9 | 0.85 | 0.018 | 0.47 | 17 | 46 |
| Upper Harbor | 1.8 | 1.3 | 0.16 | 0.29 | 7.8 | 35 |
| Lower Harbor | 0.9 | 180 | 0.055 | 0.25 | 3.7 | 20 |
| Long Island Sound | 1.6 | 0.57 | | 0.15 | 4.9 | 22 |
| NY Bight | | 0.88 | | | | 1.5 |
| Jamaica Bay | | 0.048 | | | | 0.21 |

The highest observed TDDT concentration came from a sediment core (depths .5-1 meter) taken off Staten Island just to the northeast of Prall’s Island (Figure 78). Normalization by total volatile solids does not explain away the observation.

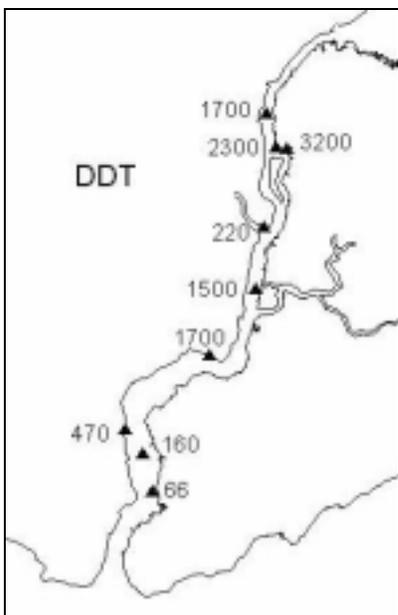
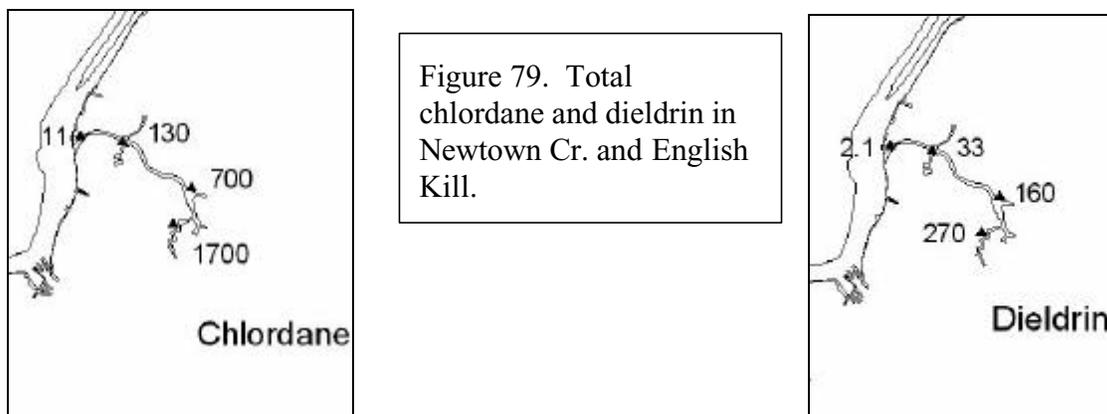


Figure 78. Average total DDT concentrations (ug/kg) in Arthur Kill core and surface grab samples.

The highest sediment concentrations of total chlordan and dieldrin occurred in Newtown Creek (Figure 79).



Next Steps

Highly contaminated particles in the Wallkill might be reduced through better stream management practices. NYSDEC is currently supporting a study on particle sources in the Black Dirt zone. In the 1930s, the Wallkill was straightened and this unnatural stream modification may be contributing to its instability. There is controversy as to the origin of the current sediment load. Some observers claim that it comes from bank erosion and the numerous drainage channels in the Black Dirt area while others profess that sediment loads come from upstream housing development. Another theory is that wind blown soils are significant.

Pesticide sources to the New York City wastewater treatment plants may be from former pesticide manufacturers. Trackdown investigations informed by the locations of these sites may be helpful. Pesticide concentrations in the Passaic River and Arthur Kill are possibly due to former manufacturing facilities. The extent to which the pesticides are still coming from terrestrial sources or are being recycled in sediments is not known.

PAHs

Polynuclear aromatic hydrocarbons (PAHs) occur as by-products of incomplete combustion. They are found in petroleum, soils, smoke, and urban wastestreams.

PAHs may be treated collectively in three ways; as summations of concentrations, molar sums, or as PAH TEQs. Molar sums are the sum of the concentrations of individual PAHs divided by their molecular weights. PAH TEQs are the sum of the products of PAH TEFs and the PAHs they relate to. TEFs for PAHs are based on values from Nisbet and LaGoy⁵. These values relate carcinogenicity of 17 PAHs to that of benzo(a)pyrene.

⁵ Nisbet, I.C.T., and P.K. LaGoy. 1992. Toxic equivalency factors (TEFs) for polycyclic aromatic hydrocarbons (PAHs). *Regulatory Toxicology and Pharmacology*. 16:290-3000.

Table 97. PAHs measured in CARP.

| PARAM | PAH_CLASS | Mol_Wt | PAH_TEF | WQS, ug/L |
|------------------------------|------------|--------|---------|-----------|
| Naphthalene | PAH-light | 128.2 | 0.001 | 16 |
| 1-Methylnaphthalene* | PAH-light | 142.2 | 0.001 | |
| 2-Methylnaphthalene* | PAH-light | 142.2 | 0.001 | |
| C1 Naphthalenes | PAH-light | 142.2 | 0.001 | |
| Acenaphthylene | PAH-light | 152.2 | 0.001 | |
| Acenaphthene | PAH-light | 154.21 | 0.001 | 6.6 |
| Biphenyl | PAH-light | 154.21 | | |
| 2,6-Dimethylnaphthalene* | PAH-light | 156.23 | | |
| C2 Naphthalenes | PAH-light | 156.23 | | |
| Fluorene | PAH-light | 166.22 | 0.001 | 2.5 |
| 2,3,5-Trimethylnaphthalene* | PAH-light | 170.26 | | |
| C3 Naphthalenes | PAH-light | 170.26 | | |
| Anthracene | PAH-medium | 178.2 | 0.01 | |
| Phenanthrene | PAH-medium | 178.24 | 0.001 | 1.5 |
| 1-Methylphenanthrene* | PAH-medium | 192.26 | | |
| C1 Phenanthrenes/Anthracenes | PAH-medium | 192.26 | | |
| Pyrene | PAH-medium | 202 | 0.001 | |
| Fluoranthene | PAH-medium | 202.26 | 0.001 | |
| 3,6-Dimethylphenanthrene* | PAH-medium | 206.28 | | |
| C2 Phenanthrenes/Anthracenes | PAH-medium | 206.28 | | |
| Benz[a]anthracene | PAH-heavy | 228.29 | 0.1 | |
| Chrysene | PAH-heavy | 228.3 | 0.01 | |
| Benzo[a]pyrene | PAH-heavy | 252.3 | 1 | 0.0006 |
| Benzo[b/j/k]fluoranthenes | PAH-heavy | 252.3 | 0.1 | |
| Benzo[b]fluoranthene | PAH-heavy | 252.3 | 0.1 | |
| Benzo[k]fluoranthene | PAH-heavy | 252.3 | 0.1 | |
| Benzo[e]pyrene | PAH-heavy | 252.32 | | |
| Perylene | PAH-heavy | 252.32 | | |
| Benzo[ghi]perylene | PAH-heavy | 276.34 | 0.01 | |
| Indeno[1,2,3-cd]pyrene | PAH-heavy | 276.34 | 0.1 | |
| Dibenz[a,h]anthracene | PAH-heavy | 278.36 | 5 | |

* Some lab reports show specific methylated PAHs (for example “1-methylnaphthalene”) and others report homologue totals, for example “C1 naphthalene”. Some AxyS data show both homologue totals and specific methylated PAHs. Data users must recognize this to avoid counting the same substances twice.

PAH Quality

TEF weighting makes some PAHs far more important than others. PAH TEQs are only calculated if the difference between assigning values of 0 or half the detection limit results in differences of less than 10%. Through the application of this screen, most of the dissolved phase PCBs become ineligible. There were 196 TOPS samples where PAHs were measured from glass fiber cartridges. Accompanying them were 135 aqueous phase PAHs samples. Because of inherent contamination with methyl naphthalenes and methyl phenanthrenes, XAD cannot be used to concentrate aqueous phase PAHs. In CARP, particle phase PAHs were captured on glass fiber filter cartridges

and aqueous phase (dissolved) PAHs were collected as part of the waste stream from TOPS. The waste stream is water that has passed through the glass fiber cartridge filter.

Evaluation of PAHs was done in three steps. The first was by analyte; the second was by sample, and the third was by sampling event. The individual analytes may have been not detected (non-detect, ND); detected at masses more than 10 times the sample specific detection level (good detection, GD); detected but at a mass less than 10 times the SPDL (high detection level, HD); found at masses more than 10 times the SPDL and more than 5 times the relevant method blank (non-significant blank, NSB); or found at more than 10 times the SPDL but less than 5 times the relevant method blank (significant blank, SB). The relevant method blank was from the same sample delivery group. Tables 98, 99, and 100 present the success (quality evaluations of GD and NSB) for each analyte in each of three media.

Table 98. PAH Data Quality, Total

| | GD | HD | ND | NSB | SB | Grand Total | Good Analyses |
|------------------------------|----|-----|-----|-----|-----|-------------|---------------|
| Naphthalene | | 12 | 7 | 26 | 17 | 62 | 42% |
| C1 Naphthalenes | | 11 | 2 | 14 | 4 | 31 | 45% |
| 1-Methylnaphthalene | 1 | 11 | 9 | 20 | 13 | 54 | 39% |
| 2-Methylnaphthalene | | 11 | 8 | 21 | 14 | 54 | 39% |
| C2 Naphthalenes | 7 | 10 | 5 | 20 | 9 | 51 | 53% |
| C3 Naphthalenes | 9 | 12 | 17 | 20 | 4 | 62 | 47% |
| Acenaphthylene | 4 | 20 | 27 | 11 | | 62 | 24% |
| Acenaphthene | 10 | 17 | 17 | 15 | 3 | 62 | 40% |
| Biphenyl | 1 | 17 | 18 | 19 | 7 | 62 | 32% |
| Fluorene | 1 | 17 | 18 | 21 | 5 | 62 | 35% |
| Anthracene | 4 | 22 | 20 | 15 | 1 | 62 | 31% |
| Phenanthrene | | 18 | 10 | 23 | 11 | 62 | 37% |
| C1 Phenanthrenes/Anthracenes | 3 | 15 | 13 | 13 | 2 | 46 | 35% |
| Pyrene | 6 | 12 | 9 | 28 | 7 | 62 | 55% |
| Fluoranthene | 5 | 14 | 8 | 29 | 6 | 62 | 55% |
| Benz[a]anthracene | 5 | 19 | 14 | 19 | 5 | 62 | 39% |
| Chrysene | 3 | 18 | 12 | 21 | 8 | 62 | 39% |
| Benzo[a]pyrene | 7 | 18 | 18 | 13 | 6 | 62 | 32% |
| Benzo[e]pyrene | 6 | 18 | 18 | 14 | 6 | 62 | 32% |
| Benzo[ghi]perylene | 4 | 16 | 19 | 15 | 8 | 62 | 31% |
| Benzo[b/j/k]fluoranthenes | 4 | 13 | 1 | 19 | 10 | 47 | 49% |
| Benzo[b]fluoranthene | | 3 | 12 | | | 15 | 0% |
| Benzo[k]fluoranthene | | 3 | 12 | | | 15 | 0% |
| Perylene | 5 | 19 | 26 | 8 | 4 | 62 | 21% |
| Indeno[1,2,3-cd]pyrene | 4 | 19 | 20 | 13 | 6 | 62 | 27% |
| Dibenz[a,h]anthracene | 2 | 14 | 35 | 5 | 6 | 62 | 11% |
| Grand Total | 91 | 379 | 375 | 422 | 162 | 1429 | 36% |

Table 99. PAH Data Quality, Dissolved Phase

| Chemical Name | GD | HD | ND | NSB | SB | Grand Total | Good Analyses |
|------------------------------|-----|-----|------|-----|-----|-------------|---------------|
| Naphthalene | | 35 | 1 | 42 | 73 | 151 | 28% |
| C1 Naphthalenes | 10 | 36 | 16 | 34 | 20 | 116 | 38% |
| 1-Methylnaphthalene | | 7 | 1 | 22 | 32 | 62 | 35% |
| 2-Methylnaphthalene | 7 | 6 | | 23 | 26 | 62 | 48% |
| C2 Naphthalenes | 43 | 18 | 47 | 15 | 15 | 138 | 42% |
| C3 Naphthalenes | 29 | 19 | 56 | 29 | 18 | 151 | 38% |
| Acenaphthylene | 6 | 38 | 99 | 3 | 5 | 151 | 6% |
| Acenaphthene | 31 | 48 | 51 | 14 | 7 | 151 | 30% |
| Biphenyl | 16 | 53 | 49 | 18 | 15 | 151 | 23% |
| Fluorene | 27 | 51 | 47 | 13 | 13 | 151 | 26% |
| Anthracene | 8 | 60 | 66 | 9 | 8 | 151 | 11% |
| Phenanthrene | 10 | 55 | 24 | 31 | 31 | 151 | 27% |
| C1 Phenanthrenes/Anthracenes | 29 | 31 | 40 | 23 | 6 | 129 | 40% |
| Pyrene | 14 | 35 | 17 | 70 | 15 | 151 | 56% |
| Fluoranthene | 7 | 53 | 25 | 51 | 15 | 151 | 38% |
| Benz[a]anthracene | 11 | 64 | 56 | 13 | 7 | 151 | 16% |
| Chrysene | 23 | 61 | 38 | 11 | 18 | 151 | 23% |
| Benzo[a]pyrene | 9 | 47 | 87 | 1 | 7 | 151 | 7% |
| Benzo[e]pyrene | 13 | 50 | 75 | 1 | 12 | 151 | 9% |
| Benzo[ghi]perylene | 4 | 83 | 53 | 1 | 10 | 151 | 3% |
| Benzo[b/j/k]fluoranthenes | 11 | 44 | 56 | 4 | 20 | 135 | 11% |
| Benzo[b]fluoranthene | | 0 | 13 | | | 13 | 0% |
| Benzo[k]fluoranthene | | 0 | 13 | | | 13 | 0% |
| Perylene | 4 | 28 | 113 | | 6 | 151 | 3% |
| Indeno[1,2,3-cd]pyrene | 5 | 60 | 76 | 2 | 8 | 151 | 5% |
| Dibenz[a,h]anthracene | 2 | 13 | 127 | 1 | 8 | 151 | 2% |
| Grand Total | 319 | 995 | 1246 | 431 | 395 | 3386 | 22% |

Table 100. PAH Data Quality, Cartridge Filters

| Chemical Name | GD | HD | ND | NSB | SB | Grand Total | Good Analyses |
|------------------------------|------|------|-----|-----|-----|-------------|---------------|
| Naphthalene | 9 | 44 | 2 | 106 | 30 | 191 | 60% |
| C1 Naphthalenes | 49 | 45 | 2 | 58 | 11 | 165 | 65% |
| 1-Methylnaphthalene* | 6 | 25 | 1 | 13 | 12 | 57 | 33% |
| 2-Methylnaphthalene* | 3 | 20 | | 24 | 10 | 57 | 47% |
| C2 Naphthalenes | 104 | 19 | 11 | 33 | 8 | 175 | 78% |
| C3 Naphthalenes | 92 | 33 | 19 | 40 | 7 | 191 | 69% |
| Acenaphthylene | 63 | 84 | 28 | 13 | 3 | 191 | 40% |
| Acenaphthene | 56 | 90 | 29 | 11 | 5 | 191 | 35% |
| Biphenyl | 52 | 88 | 11 | 25 | 15 | 191 | 40% |
| Fluorene | 77 | 62 | 13 | 29 | 10 | 191 | 55% |
| Anthracene | 65 | 68 | 15 | 39 | 4 | 191 | 54% |
| Phenanthrene | 94 | 24 | 3 | 66 | 4 | 191 | 84% |
| C1 Phenanthrenes/Anthracenes | 115 | 29 | 7 | 33 | 1 | 185 | 80% |
| Pyrene | 124 | 15 | 3 | 47 | 2 | 191 | 90% |
| Fluoranthene | 112 | 20 | 4 | 53 | 2 | 191 | 86% |
| Benz[a]anthracene | 120 | 36 | 6 | 28 | 1 | 191 | 77% |
| Chrysene | 127 | 21 | 7 | 34 | 2 | 191 | 84% |
| Benzo[a]pyrene | 111 | 44 | 11 | 22 | 2 | 190 | 70% |
| Benzo[e]pyrene | 120 | 42 | 9 | 18 | 2 | 191 | 72% |
| Benzo[ghi]perylene | 99 | 39 | 6 | 43 | 4 | 191 | 74% |
| Benzo[b/j/k]fluoranthenes | 114 | 24 | 7 | 21 | 2 | 168 | 80% |
| Benzo[b]fluoranthene | 12 | 8 | | | | 20 | 60% |
| Benzo[k]fluoranthene | 10 | 10 | | | | 20 | 50% |
| Perylene | 97 | 60 | 24 | 6 | 3 | 190 | 54% |
| Indeno[1,2,3-cd]pyrene | 109 | 44 | 6 | 27 | 2 | 188 | 72% |
| Dibenz[a,h]anthracene | 55 | 74 | 40 | 19 | 3 | 191 | 39% |
| Grand Total | 1995 | 1068 | 264 | 808 | 145 | 4280 | 65% |

The second level of evaluation addresses the suitability of an individual sample to be treated as a collective. Suitability was evaluated by assigning non-detection values of either zero or half the detection level. If the ratio of the two different summations was less than 90%, the sample was deemed unusable. Of 456 PAH samples, 402 (88%) were useable. Useable PAH samples were re-evaluated against “Good” analyses. There were 213 samples where the non-detections resulted in minor changes to the calculated nmoles/L and the quantitated analytes were present in amounts well above the detection level (10 times) and well above any method blanks (5 times). Of 465 samples, 213 are useful for evaluating nmoles/L. In the end, 186 samples are relevant. The difference between “useful” and “relevant” is that 27 otherwise useful samples were for quality control.

PAH Samples, Biosolids

On 11 occasions biosolids were analyzed for PAHs. Some the individual PAHs failed the 10 times detection level screen but they had little impact on the molar totals. The highest

PAH concentrations in sludges were seen in the industrial Hunts Point sludges and the lowest were from suburban Oakwood Beach sample.

Table 101. Collective PAHs in municipal biosolids.

| Site | Raw Sum, ppm PAH, mMoles/kg | PAH, ppm TEQ |
|----------------------------------|-----------------------------|--------------|
| Hunts Point WPCF #10 Sludge | 230 | 1.3 |
| Hunts Point WPCF #9 Sludge | 220 | 1.2 |
| 26th Ward WPCF, Sludge | 160 | 0.89 |
| Wards Island WPCF, South, Sludge | 150 | 0.86 |
| Wards Island WPCF, North, Sludge | 120 | 0.64 |
| Bowery Bay WPCF, Sludge | 110 | 0.62 |
| Coney Island WPCF, Sludge | 110 | 0.59 |
| Port Richmond WPCF, Sludge | 100 | 0.57 |
| Red Hook WPCF, Sludge | 100 | 0.56 |
| Tallman Island WPCF, Sludge | 98 | 0.54 |
| Jamaica WPCF Sludge | 81 | 0.46 |
| Oakwood Beach WPCF, Sludge | 42 | 0.22 |

The relatively large range in PAH concentrations is offset by the high consistency of PAH contributions. Figure 80 shows relative abundances of molecular weight fractions.

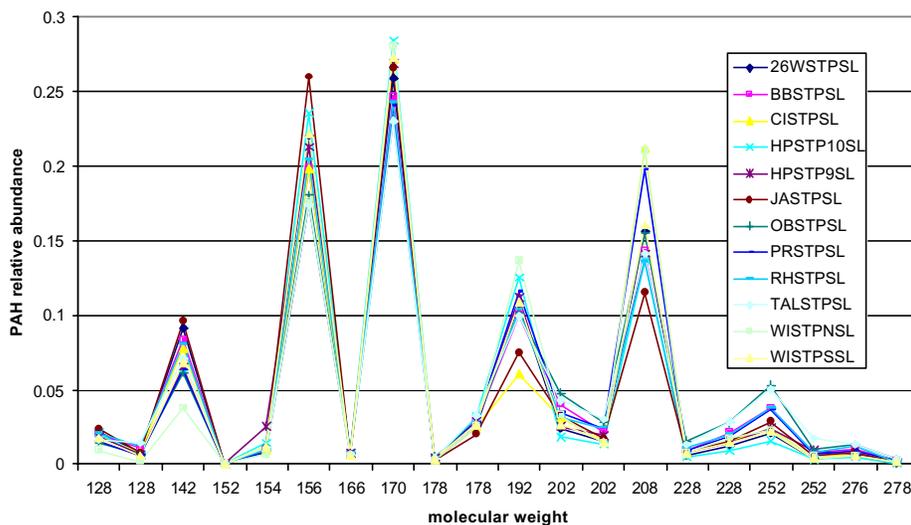


Figure 80. Relative abundances of PAHs in sludges.

About 80% of the total molar PCB mass (from sludges in the CARP analyte list) come from tri- di-, and mono-methyl naphthalenes and phenanthrenes/anthracenes.

PAHs In Water

There were two kinds of water samples, unfiltered whole water (U) and phase-separated filtered water (F). Whole water samples are easier to collect, less expensive to analyze, and possibly less susceptible to contamination. Table 102 compares samples of raw concentration (sum of all PAHs), molar concentration (nMoles/L), and raw concentration (ng/L), and B(a)P TEQs (ng/L). The labs performing the analyses are also distinguished. Samples are ranked within type by nM/L and maxima for each type are highlighted.

All samples are shown due to the difficulty of averaging. There may be large lab to lab differences and the media (filtered and unfiltered) may or may not be significant. As indicated above, the quality of these data is poor.

Only data with good detection limits were used.

Seven samples (indicated by asterisks in the Type field) were taken by the USEPA and analyzed under CARP in the investigation of the World Trade Center disaster. Two of these (CSO*, Rector St. run-off) were a slurry of dust and ash being washed off Rector St. The other 5 samples were from points just off the WTC site, at the George Washington Bridge, and off South St. in the lower East River.

Table 102. Collective PAHs in Whole Water Samples.

| Type | Site | Raw, ng/L | nM/L | TEQ, ng/kg | QC | Media | LAB | Date |
|----------|---------------------------|-----------|------|------------|----|-------|-----|----------|
| AMB | Hudson R. at Pough. | 5700 | 35 | 280 | SA | U | AAS | 4/16/99 |
| AMB | Passaic River, Mid-Tidal | 3800 | 18 | 550 | SA | F | AAS | 10/18/00 |
| AMB | Hackensack R., Mouth | 2500 | 16 | 6 | SA | F | AAS | 2/8/99 |
| AMB | Hudson R. at Pough. | 2600 | 15 | 82 | SA | U | AAS | 3/28/99 |
| AMB | Passaic River, Mid-Tidal | 1100 | 6.5 | 45 | SA | F | AAS | 8/25/99 |
| AMB | Passaic River, Mid-Tidal | 1000 | 6 | 50 | SA | F | AAS | 8/25/99 |
| AMB* | East River, South St. | 1000 | 5.7 | 37 | SA | U | AAS | 9/20/01 |
| AMB* | Hudson River, North | 900 | 4.9 | 110 | SA | U | AAS | 9/20/01 |
| AMB | Northern Arthur Kill | 840 | 4.2 | 90 | SA | F | AAS | 7/8/99 |
| AMB* | Hudson River West | 730 | 4.1 | 13 | SA | U | AAS | 9/20/01 |
| AMB* | GW Bridge | 610 | 3.1 | 110 | SA | U | AAS | 9/20/01 |
| AMB | Passaic River, Mid-Tidal | 410 | 2.7 | 2.4 | SA | F | AAS | 3/16/99 |
| AMB | Hackensack R., Mid-Tidal | 420 | 2.5 | 14 | SA | F | AAS | 10/12/99 |
| AMB | Newark Bay | 260 | 1.4 | 5.8 | SA | F | AAS | 1/27/99 |
| AMB* | Hudson River South | 200 | 1.1 | 4.4 | SA | U | AAS | 9/20/01 |
| AMB | Hudson R. S. of Harlem R. | 170 | 1 | 15 | SA | U | WSU | 6/14/00 |
| AMB | Hudson R. below Kingston | 150 | 0.97 | 2.6 | SA | F | AAS | 10/7/99 |
| AMB | Passaic R., Surface | 130 | 0.68 | 6.1 | SA | F | AAS | 6/17/99 |
| AMB | New York Bight | 100 | 0.66 | 0.1 | SA | F | AAS | 12/9/98 |
| AMB | Upper Bay | 110 | 0.61 | 0.35 | SA | F | AAS | 8/11/99 |
| AMB | Hudson R. S. of Harlem R. | 96 | 0.55 | 2.6 | SA | F | AAS | 8/12/99 |
| AMB | Upper Bay | 77 | 0.44 | 0.48 | SA | F | AAS | 12/15/98 |
| AMB | Upper Bay | 58 | 0.38 | 0.071 | DU | F | AAS | 12/15/98 |
| AMB | Passaic R., Surface | 50 | 0.26 | 0.3 | SA | F | AAS | 11/13/98 |
| CSO | Newtown Creek Influent | 800000 | 4600 | 17000 | SA | U | AAS | 1/30/01 |
| CSO | SWO-Jamaica, Ind. | 430000 | 2500 | 3000 | SA | U | AAS | 10/16/00 |
| CSO* | Rector St. run-off | 290000 | 1500 | 21000 | SA | U | AAS | 9/14/01 |
| CSO | Manhattan Pump Station | 120000 | 720 | 2700 | SA | U | AAS | 2/5/01 |
| CSO* | Rector St. run-off | 120000 | 650 | 5200 | SA | U | AAS | 9/20/01 |
| CSO | Manhattan Grit Chamber | 78000 | 410 | 21000 | SA | F | AAS | 9/24/01 |
| CSO | Red Hook Influent | 56000 | 380 | 61 | SA | F | AAS | 8/27/01 |
| CSO | Owls Head Influent | 45000 | 270 | 370 | SA | U | AAS | 11/9/00 |
| CSO | Jamaica Influent | 20000 | 130 | 16 | SA | F | AAS | 9/20/01 |
| CSO | Hunts Point Influent | 7800 | 50 | 9.7 | SA | U | AAS | 7/8/01 |
| CSO | SWO-Jamaica, Comm. | 5000 | 26 | 470 | SA | U | WSU | 6/22/00 |
| Ind. Eff | Clean Waters of NY | 15000 | 94 | 4.7 | SA | F | AAS | 9/20/99 |
| Ind. Eff | Clean Waters of NY | 3700 | 23 | 0.82 | SA | F | AAS | 4/29/99 |
| Ind. Eff | FK Plant Effluent | 130 | 0.78 | 2.1 | SA | F | AAS | 10/25/00 |
| Leachate | 1E-HMDC | 450000 | 3100 | 4700 | SA | F | AAS | 9/14/01 |
| Leachate | FK LF, 1/9 "B" | 140000 | 970 | 780 | SA | U | WSU | 5/11/00 |
| Leachate | FK LF, 1/9 "F" | 81000 | 560 | 460 | SA | U | WSU | 5/11/00 |
| Leachate | Pelham Bay | 75000 | 450 | 370 | SA | F | AAS | 11/6/98 |
| Leachate | FK LF, 1/9 Comp. | 64000 | 430 | 440 | SA | U | WSU | 5/11/00 |

Table 102 continued.

| Type | Site | Raw, ng/L | nM/L | TEQ, ng/kg | QC | Media | LAB | Date |
|-----------|--------------------------|-----------|------|------------|----|-------|-----|----------|
| Leachate | 1A-HMDC | 17000 | 120 | 31 | SA | U | WSU | 6/22/00 |
| Leachate | 1E-HMDC | 20000 | 120 | 200 | SA | U | WSU | 6/22/00 |
| Leachate | FK LF, 6/7 Comp. | 14000 | 86 | 130 | SA | U | WSU | 5/11/00 |
| Leachate | FK LF, 1/9 Comp. | 11000 | 68 | 58 | SA | F | AAS | 10/25/00 |
| Leachate | FK LF, 6/7 Comp. | 6200 | 42 | 6.3 | SA | F | AAS | 7/25/01 |
| Leachate | FK LF, 6/7 Comp. | 5700 | 38 | 7.3 | SA | F | AAS | 8/9/01 |
| Leachate | FK LF 3/4 | 5500 | 37 | 35 | SA | U | WSU | 5/11/00 |
| Leachate | 1D-HMDC | 7100 | 37 | 740 | SA | F | AAS | 9/14/01 |
| Leachate | FK LF, 6/7 Comp. | 1700 | 11 | 2.8 | SA | F | AAS | 10/25/00 |
| Leachate | 1D-HMDC | 230 | 1.1 | 32 | SA | U | WSU | 6/22/00 |
| Leachate | Pelham Bay | 48 | 0.24 | 1.6 | SA | F | AAS | 1/29/01 |
| Tributary | Saw Mill River (Yonkers) | 4000 | 24 | 5.8 | SA | F | AAS | 5/5/99 |
| Tributary | Gowanus Canal | 1100 | 6.4 | 7.5 | SA | F | AAS | 8/24/99 |
| Tributary | Saw Mill River (Yonkers) | 780 | 4.5 | 8.1 | SA | F | AAS | 11/10/98 |
| Tributary | Mohawk R. (Cohoes) | 490 | 3.3 | 1.4 | SA | U | AAS | 4/1/99 |
| Tributary | Gowanus Canal | 500 | 2.8 | 19 | SA | U | AAS | 9/28/00 |
| Tributary | Bronx River | 430 | 2.6 | 1.2 | SA | F | AAS | 7/27/99 |
| Tributary | Wallkill (New Paltz) | 350 | 2.3 | 3.1 | SA | U | AAS | 10/13/99 |
| Tributary | Bronx River | 360 | 2.3 | 3.9 | SA | F | AAS | 10/26/99 |
| Tributary | Hudson R. (Pleasantdale) | 290 | 2.1 | 0.66 | SA | U | AAS | 4/8/99 |
| Tributary | Hudson R. (Pleasantdale) | 280 | 1.8 | 2.6 | SA | U | AAS | 4/1/99 |
| Tributary | Hudson R. (Pleasantdale) | 240 | 1.5 | 2.5 | SA | U | AAS | 3/22/99 |
| Tributary | Wallkill (New Paltz) | 78 | 0.5 | 0.14 | SA | U | AAS | 5/26/01 |
| WPCF | Newtown Creek | 64000 | 400 | 46 | DU | F | AAS | 9/28/99 |
| WPCF | Newtown Creek | 63000 | 400 | 38 | SA | F | AAS | 9/28/99 |
| WPCF | Newtown Creek | 51000 | 320 | 25 | SA | F | AAS | 6/22/99 |
| WPCF | 26th Ward | 34000 | 210 | 9.1 | SA | F | AAS | 5/5/99 |
| WPCF | Port Richmond | 26000 | 170 | 22 | SA | F | AAS | 10/20/99 |
| WPCF | Hunts Point | 23000 | 130 | 240 | SA | U | AAS | 2/1/01 |
| WPCF | Oakwood Beach | 16000 | 100 | 5.6 | SA | F | AAS | 8/18/99 |
| WPCF | Poughkeepsie City | 16000 | 93 | 200 | SA | F | AAS | 8/19/99 |
| WPCF | Rockland County | 14000 | 84 | 290 | SA | F | AAS | 8/19/99 |
| WPCF | Tallman Island | 14000 | 81 | 29 | SA | F | AAS | 9/6/00 |
| WPCF | Tallman Island | 12000 | 66 | 23 | DU | F | AAS | 9/6/00 |
| WPCF | Owls Head | 7700 | 43 | 30 | SA | F | AAS | 8/23/00 |
| WPCF | Rensselaer | 6500 | 41 | 8.9 | SA | F | AAS | 1/12/99 |
| WPCF | Rockaway | 4700 | 30 | 9 | SA | F | AAS | 8/11/99 |
| WPCF | Poughkeepsie City | 3300 | 23 | 8.1 | SA | F | AAS | 4/1/99 |
| WPCF | Wards Island | 3700 | 23 | 2.6 | SA | F | AAS | 1/20/99 |
| WPCF | North River | 3300 | 20 | 6.9 | SA | F | AAS | 9/1/99 |
| WPCF | Poughkeepsie City | 3400 | 19 | 24 | SA | U | AAS | 12/5/00 |
| WPCF | Bowery Bay | 2900 | 17 | 12 | SA | F | AAS | 11/5/98 |
| WPCF | Port Richmond | 2400 | 15 | 2.2 | SA | F | AAS | 8/25/99 |
| WPCF | Red Hook | 2100 | 13 | 3 | SA | F | AAS | 2/3/99 |

Table 102 continued.

| Type | Site | Raw, ng/L | nM/L | TEQ, ng/kg | QC | Media | LAB | Date |
|------|-----------------|-----------|------|------------|----|-------|-----|---------|
| WPCF | Owls Head | 2000 | 12 | 2.3 | SA | F | AAS | 9/15/98 |
| WPCF | Jamaica | 2000 | 12 | 33 | SA | U | AAS | 2/15/01 |
| WPCF | Wards Island | 1500 | 9.2 | 2.7 | DU | F | AAS | 8/10/00 |
| WPCF | Yonkers | 930 | 5.6 | 8.5 | SA | F | AAS | 8/18/99 |
| WPCF | Rockland County | 890 | 5.4 | 5.4 | SA | F | AAS | 4/20/99 |
| WPCF | Owls Head | 620 | 3.7 | 3.4 | SA | F | AAS | 7/7/99 |
| WPCF | Tallman Island | 570 | 3.4 | 0.35 | SA | F | AAS | 2/12/99 |
| WPCF | Tallman Island | 370 | 2.2 | 0.68 | SA | F | AAS | 7/20/99 |
| WPCF | Wards Island | 320 | 1.9 | 0.57 | SA | F | AAS | 8/10/00 |
| WPCF | Yonkers | 210 | 1.3 | 0.086 | SA | F | AAS | 4/22/99 |
| WPCF | Rensselaer | 170 | 0.99 | 0.86 | SA | F | AAS | 8/11/99 |
| WPCF | 26th Ward | 180 | 0.99 | 1.4 | SA | F | AAS | 1/27/99 |
| WPCF | Oakwood Beach | 120 | 0.78 | 0.1 | SA | F | AAS | 2/11/99 |
| WPCF | Jamaica | 79 | 0.49 | 0.13 | SA | F | AAS | 2/5/99 |

Total PAHs

Attempts to perform phase separation on PAH samples were disappointing. Tables 103, 104, and 105 shows site averages of B(a)P TEQ, total nMoles/L, and raw total summations from filtered grab samples, unfiltered grab samples, and TOPS glass fiber cartridge samples. Data are screened to show only samples where assigning non-detections the value of the sample specific detection limit or zero results in a difference of less than 10%. Only data passing the screen for good detections are used. Ultimately, very few dissolved phase samples survive the quality screens. This has a profound impact in that where it is possible to compare phases on the same samples, much or most of the PAH is on the dissolved phase.

Table 103. B(a)P TEQ in aqueous and suspended particulate samples, site averages, ng/L. Censored data.

| Sample | Filtered | Particulate | Unfiltered | Total |
|---|----------|-------------|------------|-------|
| Amb: Passaic River, Mid-Tidal | 550 | 2400 | | 2950 |
| Amb: Hudson R. at Poughkeepsie | | 800 | 280 | 1080 |
| Amb: Passaic R., Mouth, Bottom | | 1000 | | 1000 |
| Amb: Passaic R., Mouth, Surface | | 960 | | 960 |
| Amb: Hackensack R., Mid-Tidal | | 700 | | 700 |
| Amb: Northern Arthur Kill | 90 | 270 | | 360 |
| Amb: Hudson R. South of Harlem R. | | 340 | 15 | 355 |
| Amb: Hudson R. below Tappen Zee | | 290 | | 290 |
| Amb: Hackensack R., Mouth | | 230 | | 230 |
| Amb: Lower East R. | | 230 | | 230 |
| Amb: Newark Bay | | 230 | | 230 |
| Amb: Upper Bay | | 170 | | 170 |
| Amb: Haverstraw Bay | | 150 | | 150 |
| Amb: Upper East R. | | 140 | | 140 |
| Amb: WTC George Washington Bridge | | | 110 | 110 |
| Amb: WTC Hudson River, North | | | 110 | 110 |
| Amb: Hudson R. below Kingston | | 42 | | 42 |
| Amb: Raritan Bay | | 34 | | 34 |
| Amb: Jamaica Bay | | 29 | | 29 |
| Amb: Lower Bay | | 28 | | 28 |
| Amb: Long Island Sound | | 16 | | 16 |
| CSO: Manhattan Grit Chamber | 21000 | 2100 | | 23100 |
| CSO: Newtown Creek Influent | | | 17000 | 17000 |
| CSO: Red Hook Influent | | 7200 | | 7200 |
| CSO: Jamaica Influent | | 3000 | | 3000 |
| CSO: SWO-Jamaica, Industrial | | | 3000 | 3000 |
| CSO: Manhattan Pump Station | | | 2700 | 2700 |
| CSO: Hunts Point Influent | | 2400 | | 2400 |
| CSO: SWO-Jamaica, Commercial | | | 470 | 470 |
| CSO: Owls Head Influent | | | 370 | 370 |
| Industrial effluent: Clean Waters of New York | | 5.8 | | 5.8 |
| Industrial effluent: FK Plant Effluent | 2.1 | | | 2.1 |
| Landfill leachate: 1E-HMDC | 4700 | | 200 | 4900 |
| Landfill leachate: FK LF, 1/9 "B" | | | 780 | 780 |
| Landfill leachate: 1D-HMDC | 740 | | 32 | 772 |
| Landfill leachate: FK LF, 1/9 Comp. | 58 | | 440 | 498 |
| Landfill leachate: FK LF, 1/9 "F" | | | 460 | 460 |
| Landfill leachate: FK LF, 6/7 Comp. | | | 130 | 130 |
| Landfill leachate: FK LF 3/4 | | | 35 | 35 |
| Landfill leachate: 1A-HMDC | | | 31 | 31 |
| Tributary: Bronx River | | 580 | | 580 |
| Tributary: Wallkill (New Paltz) | | 530 | | 530 |
| Tributary: Hudson R. (Pleasantdale) | | 290 | | 290 |
| Tributary: Mohawk R. (Cohoes) | | 270 | | 270 |

Table 103, continued.

| Sample | Filtered | Particulate | Unfiltered | Total |
|-------------------------------------|----------|-------------|------------|-------|
| Tributary: Gowanus Canal | | 240 | 19 | 259 |
| Tributary: Saw Mill River (Yonkers) | | 220 | | 220 |
| WPCF: Hunts Point | | 340 | 240 | 580 |
| WPCF: Rockland County | 290 | 40 | | 330 |
| WPCF: Rensselaer | | 180 | | 180 |
| WPCF: Poughkeepsie City | | 160 | | 160 |
| WPCF: Jamaica | | 80 | 33 | 113 |
| WPCF: 26th Ward | | 98 | | 98 |
| WPCF: Red Hook | | 97 | | 97 |
| WPCF: Bowery Bay | | 96 | | 96 |
| WPCF: Newtown Creek | | 90 | | 90 |
| WPCF: Tallman Island | | 63 | | 63 |
| WPCF: Rockaway | | 48 | | 48 |
| WPCF: Port Richmond | | 45 | | 45 |
| WPCF: Yonkers | | 37 | | 37 |
| WPCF: Wards Island | | 33 | | 33 |
| WPCF: Oakwood Beach | | 17 | | 17 |
| WPCF: North River | | 8.1 | | 8.1 |
| WPCF: Owls Head | | 6.3 | | 6.3 |

Table 104. Total molar PAHs (ng/L) by phase. Good detection limits and the difference between assigning non-detections values of zero and the detection limit is less than 10%.

| Sample | Filtered | Particulate | Unfiltered | Total |
|-------------------------------------|----------|-------------|------------|--------|
| Amb: Passaic River, Mid-Tidal | 10 | 66 | | 76 |
| Amb: Hudson R. at Poughkeepsie | | 31 | 25 | 56 |
| Amb: Passaic R., Mouth, Bottom | | 35 | | 35 |
| Amb: Passaic R., Mouth, Surface | 0.68 | 29 | | 29.68 |
| Amb: Hackensack R., Mid-Tidal | 2.5 | 22 | | 24.5 |
| Amb: Hackensack R., Mouth | 16 | 7 | | 23 |
| Amb: Hudson R. South of Harlem R. | 0.55 | 11 | 1 | 12.55 |
| Amb: Northern Arthur Kill | 4.2 | 8.3 | | 12.5 |
| Amb: Hudson R. below Tappen Zee | | 11 | | 11 |
| Amb: Lower East R. | | 7.8 | | 7.8 |
| Amb: Newark Bay | | 7.3 | | 7.3 |
| Amb: Upper Bay | 0.53 | 5.6 | | 6.13 |
| Amb: WTC East River, South St. | | | 5.7 | 5.7 |
| Amb: Haverstraw Bay | | 5.6 | | 5.6 |
| Amb: WTC Hudson River, North | | | 4.9 | 4.9 |
| Amb: Upper East R. | | 4.2 | | 4.2 |
| Amb: WTC Hudson River West | | | 4.1 | 4.1 |
| Amb: WTC George Washington Bridge | | | 3.1 | 3.1 |
| Amb: Hudson R. below Kingston | 0.97 | 1.7 | | 2.67 |
| Amb: Raritan Bay | | 1.1 | | 1.1 |
| Amb: Jamaica Bay | | 0.76 | | 0.76 |
| Amb: Lower Bay | | 0.74 | | 0.74 |
| Amb: New York Bight | 0.66 | 0.0078 | | 0.6678 |
| Amb: Long Island Sound | | 0.45 | | 0.45 |
| CSO: Newtown Creek Influent | | | 4600 | 4600 |
| CSO: SWO-Jamaica, Industrial | | | 2500 | 2500 |
| CSO: WTC Rector St. run-off | | | 1100 | 1100 |
| CSO: Red Hook Influent | 380 | 380 | | 760 |
| CSO: Manhattan Pump Station | | | 720 | 720 |
| CSO: Manhattan Grit Chamber | 410 | 130 | | 540 |
| CSO: Jamaica Influent | 130 | 200 | | 330 |
| CSO: Owls Head Influent | | | 270 | 270 |
| CSO: Hunts Point Influent | | 150 | 50 | 200 |
| CSO: SWO-Jamaica, Commercial | | | 26 | 26 |
| Ind. Eff.: Clean Waters of New York | 59 | 1.1 | | 60.1 |
| Ind. Eff.: FK Plant Effluent | 0.78 | 15 | | 15.78 |
| Leachate: 1E-HMDC | 3100 | | 120 | 3220 |
| Leachate: FK LF, 1/9 "B" | | | 970 | 970 |
| Leachate: FK LF, 1/9 "F" | | | 560 | 560 |
| Leachate: FK LF, 1/9 Comp. | 68 | | 430 | 498 |
| Leachate: Pelham Bay | 230 | | | 230 |
| Leachate: 1A-HMDC | | | 120 | 120 |
| Leachate: FK LF, 6/7 Comp. | 30 | | 86 | 116 |

Table 104 continued.

| Sample | Filtered | Particulate | Unfiltered | Total |
|-------------------------------------|----------|-------------|------------|-------|
| Leachate: 1D-HMDC | 37 | | 1.1 | 38.1 |
| Leachate: FK LF 3/4 | | | 37 | 37 |
| Tributary: Saw Mill River (Yonkers) | 14 | 7 | | 21 |
| Tributary: Walkill (New Paltz) | | 17 | 1.4 | 18.4 |
| Tributary: Gowanus Canal | 6.4 | 8.8 | 2.8 | 18 |
| Tributary: Bronx River | 2.4 | 14 | | 16.4 |
| Tributary: Mohawk R. (Cohoes) | | 8.1 | 3.3 | 11.4 |
| Tributary: Hudson R. (Pleasantdale) | | 7.9 | 1.9 | 9.8 |
| WPCF: Newtown Creek | 370 | 29 | | 399 |
| WPCF: 26th Ward | 210 | 15 | | 225 |
| WPCF: Hunts Point | | 22 | 130 | 152 |
| WPCF: Oakwood Beach | 100 | 1.7 | | 101.7 |
| WPCF: Port Richmond | 92 | 3.7 | | 95.7 |
| WPCF: Poughkeepsie City | 58 | 11 | 19 | 88 |
| WPCF: Tallman Island | 50 | 5 | | 55 |
| WPCF: Rockland County | 45 | 3.4 | | 48.4 |
| WPCF: Rockaway | 30 | 3.8 | | 33.8 |
| WPCF: Rensselaer | 21 | 7.6 | | 28.6 |
| WPCF: North River | 20 | 1.9 | | 21.9 |
| WPCF: Bowery Bay | 17 | 4 | | 21 |
| WPCF: Owls Head | 20 | 0.88 | | 20.88 |
| WPCF: Jamaica | | 3.4 | 12 | 15.4 |
| WPCF: Red Hook | 13 | 2.2 | | 15.2 |
| WPCF: Wards Island | 11 | 1.4 | | 12.4 |
| WPCF: Yonkers | 5.6 | 4.5 | | 10.1 |
| WPCF: Coney Island | | 0.53 | | 0.53 |

Table 105. Total PAH concentrations. ng/L. Data censored for high detection levels.

| Sample | Filtered | Particulate Unfiltered | Total |
|---------------------------------------|----------|------------------------|---------------|
| Ambient: Passaic River, Mid-Tidal | 1600 | 14000 | 15600 |
| Ambient: Hudson R. at Poughkeepsie | | 6500 | 4200 10700 |
| Ambient: Passaic R., Mouth, Bottom | | 7400 | 7400 |
| Ambient: Passaic R., Mouth, Surface | 89 | 6300 | 6389 |
| Ambient: Hackensack R., Mid-Tidal | 420 | 4700 | 5120 |
| Ambient: Hackensack R., Mouth | 2500 | 1500 | 4000 |
| Ambient: Hudson R. South of Harlem R. | 96 | 2400 | 170 2666 |
| Ambient: Northern Arthur Kill | 840 | 1800 | 2640 |
| Ambient: Hudson R. below Tappen Zee | | 2300 | 2300 |
| Ambient: Newark Bay | 260 | 1500 | 1760 |
| Ambient: Lower East R. | | 1600 | 1600 |
| Ambient: Upper Bay | 81 | 1200 | 1281 |
| Ambient: Haverstraw Bay | | 1200 | 1200 |
| Ambient: WTC East River, South St. | | | 1000 1000 |
| Ambient: WTC Hudson River, North | | | 900 900 |
| Ambient: Upper East R. | | 880 | 880 |
| Ambient: WTC Hudson River West | | | 730 730 |
| Ambient: WTC George Washington Bridge | | | 610 610 |
| Ambient: Hudson R. below Kingston | 150 | 350 | 500 |
| Ambient: Raritan Bay | | 230 | 230 |
| Ambient: WTC Hudson River South | | | 200 200 |
| Ambient: Jamaica Bay | | 160 | 160 |
| Ambient: Lower Bay | | 160 | 160 |
| Ambient: New York Bight | 100 | 1.5 | 101.5 |
| Ambient: Long Island Sound | | 97 | 97 |
| CSO: Newtown Creek Influent | | | 800000 800000 |
| CSO: SWO-Jamaica, Industrial | | | 430000 430000 |
| CSO: WTC Rector St. run-off | | | 210000 210000 |
| CSO: Red Hook Influent | 56000 | 75000 | 131000 |
| CSO: Manhattan Pump Station | | | 120000 120000 |
| CSO: Manhattan Grit Chamber | 78000 | 26000 | 104000 |
| CSO: Jamaica Influent | 20000 | 39000 | 59000 |
| CSO: Owls Head Influent | | | 45000 45000 |
| CSO: Hunts Point Influent | | 30000 | 7800 37800 |
| CSO: SWO-Jamaica, Commercial | | | 5000 5000 |
| Ind. Eff: Clean Waters of New York | 9300 | 210 | 9510 |
| Ind. Eff: FK Plant Effluent | 130 | 3000 | 3130 |
| Leachate: 1E-HMDC | 450000 | | 20000 470000 |
| Leachate: FK LF, 1/9 "B" | | | 140000 140000 |
| Leachate: FK LF, 1/9 "F" | | | 81000 81000 |
| Leachate: FK LF, 1/9 Comp. | 11000 | | 64000 75000 |
| Leachate: Pelham Bay | 38000 | | 38000 |
| Leachate: FK LF, 6/7 Comp. | 4600 | | 14000 18600 |
| Leachate: 1A-HMDC | | | 17000 17000 |
| Leachate: 1D-HMDC | 7100 | | 230 7330 |

Table 105 continued.

| Sample | Filtered | Particulate | Unfiltered | Total |
|---------------------------------------|----------|-------------|------------|-------|
| Leachate: FK LF 3/4 | | | 5500 | 5500 |
| Tributaries: Wallkill (New Paltz) | | 3700 | 220 | 3920 |
| Tributaries: Saw Mill River (Yonkers) | 2400 | 1500 | | 3900 |
| Tributaries: Bronx River | 390 | 3200 | | 3590 |
| Tributaries: Gowanus Canal | 1100 | 1800 | 500 | 3400 |
| Tributaries: Mohawk R. (Cohoes) | | 1700 | 490 | 2190 |
| Tributaries: Hudson R. (Pleasantdale) | | 1700 | 270 | 1970 |
| WPCF: Newtown Creek | 59000 | 5400 | | 64400 |
| WPCF: Hunts Point | | 4600 | 23000 | 27600 |
| WPCF: 26th Ward | 17000 | 2900 | | 19900 |
| WPCF: Poughkeepsie City | 9500 | 2300 | 3400 | 15200 |
| WPCF: Port Richmond | 14000 | 730 | | 14730 |
| WPCF: Oakwood Beach | 8200 | 340 | | 8540 |
| WPCF: Rockland County | 7500 | 670 | | 8170 |
| WPCF: Tallman Island | 6800 | 700 | | 7500 |
| WPCF: Rockaway | 4700 | 750 | | 5450 |
| WPCF: Rensselaer | 3400 | 1600 | | 5000 |
| WPCF: Bowery Bay | 2900 | 820 | | 3720 |
| WPCF: Owls Head | 3500 | 170 | | 3670 |
| WPCF: North River | 3300 | 220 | | 3520 |
| WPCF: Jamaica | 79 | 700 | 2000 | 2779 |
| WPCF: Red Hook | 2100 | 470 | | 2570 |
| WPCF: Wards Island | 1800 | 290 | | 2090 |
| WPCF: Yonkers | 570 | 810 | | 1380 |
| WPCF: Coney Island | | 110 | | 110 |

The quality of the PAH data are clearly problematic, particularly for the more critical dissolved phase. The source of the problem is in large measure due to inadequate field concentration. There may be problem with field contamination in some cases. Data are significantly better when the results are expressed in molar units than in B(a)P equivalents. Much of the B(a)P-like material was poorly captured in the dissolved phase samples resulting in numerous non-detections. The molar summations preferentially weights toward lighter PAHs more likely to be found in the dissolved phase but also more likely to be the result of sample contamination. Fewer problems are encountered in total PAH data but this statistic is perhaps less meaningful.

TOPS does not assist in the in-situ concentration for the dissolved phase PAHs because of the background of methylated naphthalenes and phenanthrenes on XAD. These substances are very important for total molar concentration but lack B(a)P TEFs. Proper sampling of PAHs will require experimentation.

Metals

Metal Results, Quality

Metals were analyzed from grab samples in the CARP. Table 106 shows the metals (and the phases), the number of non-detections, and the total number of samples taken.

Table 106. Metals analyzed, number of non-detections and total number of samples.

| PARAM | non-detect | total |
|-----------------------|------------|-------|
| Ag, dissolved | 7 | 46 |
| Ag, total | 3 | 53 |
| Cd, dissolved | 29 | 237 |
| Cd, total | 6 | 235 |
| Hg, dissolved | 2 | 256 |
| Hg, Methyl, dissolved | 45 | 194 |
| Hg, Methyl, total | 1 | 4 |
| Hg, total | 2 | 256 |
| Pb, dissolved | 7 | 116 |
| Pb, total | 1 | 65 |

Mercury and cadmium were the original CARP metals of concern. Lead and silver were added later in the program. The numbers of analyses reflect this.

The quality of metals data are evaluated in two parts, for mercury, and for the others. Good data are defined in Table 107.

Table 107. Criteria used to evaluate metals data.

| | Hg | Ag, Cd, or Pb | notes |
|----------------|---------------------|---------------------|---------------------------|
| blanks | > 5 x lab SDG blank | > 5 x lab SDG blank | ND is assigned value of 0 |
| spike recovery | < +/- 20% | < +/- 20% | abs diff/mean |
| duplicates | < +/- 25% | < +/- 20% | abs diff/mean |

The sampling and analytical protocol did not provide a full set of quality control information for each analysis. Metals samples were processed quickly and many sample delivery groups had very few samples. The average number of samples in a SDG was 2.7. Hence, duplicate measurements were only made from a subset of SDGs. Also, percent recoveries were usually determined from total metals, not dissolved metals. Therefore, there are fewer data on recovery efficiency from dissolved phase samples.

Table 108 shows the average relative percent recovery from duplicates. The only samples exceeding the criteria were three dissolved methyl mercury sets where the total amount of analyte was very small.

Table 108. Average Relative Percent Difference (RPD) from duplicates.

| PARAM | Average RPD |
|---------------------|-------------|
| Cd_dissolved | 1% |
| Cd_total | 3% |
| Hg_dissolved | 2% |
| Hg_Methyl_Dissolved | 14% |
| Hg_Methyl_total | 13% |
| Hg_total | 2% |
| Pb_dissolved | 1% |

Table 109 summarizes the quality of metals data for blanks and spike recovery. Standards are described in Table 107. “Rec_Ind” refers to samples where there were no appropriate recovery statistics.

Table 109. Metals quality statistics.

| | Good Blank | Good Blank | Good Blank | High Blank | High Blank | Grand Total |
|---------------------|------------|------------|------------|------------|------------|-------------|
| PARAM | Bad Rec | Good Rec | Rec_Ind | Good Rec | Rec_Ind | |
| Ag_dissolved | | | 29 | | 2 | 31 |
| Ag_total | 2 | 24 | | 2 | | 28 |
| Cd_dissolved | | | 177 | | 26 | 203 |
| Cd_total | 8 | 193 | | 10 | | 211 |
| Hg_dissolved | | 18 | 207 | | 9 | 234 |
| Hg_Methyl_Dissolved | 22 | 145 | | 16 | | 183 |
| Hg_total | 3 | 259 | | 3 | | 265 |
| Pb_dissolved | | | 89 | | 6 | 95 |
| Pb_total | 4 | 34 | | | | 38 |
| Grand Total | 39 | 673 | 502 | 31 | 43 | 1288 |

Of the metals for which we have good statistics, total cadmium and total mercury, quality data appear very good. The weakest data are for methyl mercury. These were most often non-detect, showed the worst reproducibility, and most frequently had blank contamination.

Samples

Table 110 summarizes the metals concentrations by stations. The analytes are abbreviated; D= dissolved, T=total, and M=methyl. Missing values occur when the analyte was not measured.

Table 105. Summary of station averages, ng/L

| Sample | Ag, D | Ag, T | Cd, D | Cd, T | Hg, D | Hg, DM | Hg, T | Pb, D | Pb, T |
|---------------------------------|-------|--------|-------|-------|-------|--------|-------|--------|---------|
| Amb-clean: Long Island Sound | | | 41 | 45 | 0.54 | 0.012 | 1.5 | 12 | |
| Amb-clean: New York Bight | | | 23 | 21 | 0.39 | 0.045 | 0.63 | | |
| Amb-Hudson: Haverstraw Bay | | | 27 | 46 | 2.3 | 0.054 | 6.7 | 99 | |
| Amb-Hudson: Poughkeepsie | | | 11 | 150 | 1.4 | 0.11 | 26 | 120 | |
| Amb-Hudson: below Kingston | | | 12 | 23 | 1.4 | 0.081 | 11 | 220 | |
| Amb-Hudson: below Tappan Zee | | | 68 | 100 | 2.4 | 0.034 | 29 | 57 | |
| Amb-Hudson: S. of Harlem R. | 13 | 49 | 83 | 83 | 1.6 | 0.017 | 11 | 100 | 1200 |
| Amb-Kills: Hackensack R., MT | | | 36 | 100 | 2.7 | 0.089 | 160 | 690 | |
| Amb-Kills: Hackensack R., M. | | | 56 | 70 | 1 | 0.043 | 22 | 160 | |
| Amb-Kills: Newark Bay | | | 67 | 88 | 1.6 | 0.02 | 33 | 130 | |
| Amb-Kills: Northern Arthur Kill | | | 77 | 83 | 0.85 | 0.025 | 33 | 130 | |
| Amb-Kills: Passaic R., M., Bot. | | | 61 | 150 | 1.7 | 0.027 | 87 | 380 | |
| Amb-Kills: Passaic R., M., Sur | | | 94 | 110 | 1.7 | 0.029 | 31 | 130 | |
| Amb-Kills: Passaic River, MT | 0 | 130 | 35 | 160 | 1.5 | 0.05 | 82 | 410 | 11000 |
| Amb-Non_Kills: Jamaica Bay | | | 38 | 43 | 1.1 | 0.019 | 3.4 | 93 | |
| Amb-Non_Kills: Lower Bay | | | 51 | 49 | 2.5 | 0.045 | 3.6 | 740 | |
| Amb-Non_Kills: Lower East R. | | | 56 | 65 | 0.76 | 0.017 | 14 | 130 | |
| Amb-Non_Kills: Raritan Bay | | | 58 | 61 | 1.5 | 0.016 | 7.7 | 100 | |
| Amb-Non_Kills: Upper Bay | | | 43 | 87 | 0.98 | 0.024 | 11 | 88 | |
| Amb-Non_Kills: Upper East R. | | | 78 | 70 | 2.7 | 0.0052 | 7.4 | 510 | |
| CSO: Bowery Bay High Side | 1,400 | 29,000 | 88 | 5,700 | 8.8 | | 2,900 | 2,400 | 340,000 |
| CSO: Bowery Bay Low Side | 100 | 240 | 95 | 1,200 | 11 | 0.11 | 680 | 2,600 | 80,000 |
| CSO: Coney Island Influent | | 210 | | 540 | | | 400 | | 39,000 |
| CSO: Hunts Point Influent | 87 | 3,000 | 21 | 1,100 | 8.9 | | 720 | 1,200 | 100,000 |
| CSO: Jamaica Influent | 780 | 3,100 | 52 | 500 | 20 | 1.5 | 410 | 2,000 | 32,000 |
| CSO: Manhattan Grit Chamber | 55 | 860 | 4 | 350 | 3.5 | 0.63 | 180 | 580 | 42,000 |
| CSO: Manhattan Pump Station | 790 | 2,500 | 140 | 640 | 12 | 0.33 | 690 | 1,900 | 3,200 |
| CSO: Newtown Creek Influent | 110 | 1,400 | 150 | 2,200 | 10 | 0.16 | 620 | 4,000 | 150,000 |
| CSO: North River Influent | 1,600 | 20,000 | 70 | 1,400 | 9.7 | | 1,500 | 2,600 | 310,000 |
| CSO: Owls Head Influent | | 340 | | 660 | | | 930 | | 57,000 |
| CSO: Port Richmond Influent | 60 | 230 | 43 | 270 | 12 | | 150 | 2,700 | 26,000 |
| CSO: Red Hook Influent | 110 | 3,700 | 11 | 1,100 | 16 | | 750 | 850 | 88,000 |
| CSO: SWO-Jamaica, Comm. | | | 72 | 900 | 11 | | 120 | 6,200 | |
| CSO: SWO-Jamaica, Ind. | 0 | 36 | 240 | 3,700 | 5.6 | | 78 | 4,000 | 93,000 |
| Ind eff: Clean Waters of NY | | | 46 | 63 | 0.4 | 0.041 | 0.54 | | |
| Ind eff: FK Plant Effluent | 110 | 120 | 140 | 170 | 29 | 0.24 | 37 | 640 | 1500 |
| LF leachate: 1A-HMDC | | | 24 | | 23 | | | 1,300 | |
| LF leachate: 1D-HMDC | 380 | | 710 | | 290 | 1.7 | | 66,000 | |
| LF leachate: 1E-HMDC | 50 | | 45 | | 50 | 0.88 | | 11,000 | |
| LF leachate: FK LF 3/4 | | | | 160 | | | 8.8 | 13,000 | |
| LF leachate: FK LF, 1/9 "B" | | | | 780 | | | 46 | 41,000 | |
| LF leachate: FK LF, 1/9 "F" | | | | 230 | | | 11 | 2,300 | |
| LF leachate: FK LF, 1/9 Comp. | 150 | 290 | 660 | 750 | 1900 | 0.29 | 31 | 4,300 | 6800 |
| LF leachate: FK LF, 6/7 Comp. | 91 | 130 | 20 | 94 | 1.6 | 0.17 | 3 | 610 | 150 |
| LF leachate: Pelham Bay | | | 11 | | 2.6 | | | 420 | |

Table 110 continued.

| Sample | Ag, D | Ag, T | Cd, D | Cd, T | Hg, D | Hg, DM | Hg, T | Pb, D | Pb, T |
|---------------------------------|-------|-------|-------|-------|-------|--------|-------|-------|-------|
| Major tributary: Hudson R. | | | 8.3 | 140 | 1.6 | 0.045 | 14 | 110 | |
| Major tributary: Mohawk R. | | | 14 | 180 | 2 | 0.035 | 26 | 160 | 7,900 |
| Major tributary: Wallkill | 14 | 160 | 9.5 | 290 | 4.3 | 0.059 | 42 | 150 | 6,300 |
| Minor tributary: Bronx River | | | 25 | 43 | 1.8 | 0.023 | 5.6 | 160 | |
| Minor tributary: Gowanus Canal | | | 37 | 64 | 1.1 | 0.15 | 11 | 170 | |
| Minor tributary: Saw Mill River | | | 26 | 37 | 3.5 | 0.055 | 3.5 | | |
| WPCF: 26th Ward | 71 | 430 | 43 | 49 | 5.3 | 0.036 | 19 | 920 | 3,800 |
| WPCF: Bowery Bay | | | 48 | 61 | 2.7 | 0.081 | 12 | | |
| WPCF: Coney Island | | | 19 | 23 | 3.5 | 0.08 | 9.3 | 570 | |
| WPCF: Hunts Point | 86 | 260 | 49 | 66 | 2.6 | 0.035 | 9.3 | 350 | 790 |
| WPCF: Jamaica | 150 | 690 | 57 | 170 | 2.5 | 0.11 | 38 | 450 | 1,700 |
| WPCF: Newtown Creek | 750 | 2,900 | 300 | 430 | 6.9 | 0.5 | 41 | 1,100 | 2,300 |
| WPCF: North River | | | 140 | 140 | 3.9 | 0.092 | 15 | 1,300 | 1,600 |
| WPCF: Oakwood Beach | | | 40 | 48 | 2.5 | 0.042 | 2.7 | 250 | |
| WPCF: Owls Head | | | 30 | 42 | 10 | 0.063 | 18 | 690 | |
| WPCF: Port Richmond | 110 | 390 | 94 | 95 | 3.4 | 0.05 | 11 | 680 | 1,400 |
| WPCF: Poughkeepsie City | 180 | 0 | 100 | 160 | 4.6 | 0.27 | 41 | 1,100 | 1,700 |
| WPCF: Red Hook | | | 74 | 75 | 2.3 | 0.13 | 8.6 | | |
| WPCF: Rensselaer | | | 40 | 61 | 5.3 | 0.086 | 16 | | |
| WPCF: Rockaway | | | 29 | 48 | 3.4 | 0.057 | 14 | 330 | |
| WPCF: Rockland County | | | 30 | 39 | 15 | 0.52 | 64 | 310 | |
| WPCF: Tallman Island | | | 75 | 88 | 1.8 | 0 | 9.6 | 270 | |
| WPCF: Wards Island | | | 37 | 38 | 2.5 | 0.023 | 7.9 | 420 | |
| WPCF: Yonkers | | | 51 | 75 | 4.7 | 0.097 | 61 | 1,100 | |

Average station instantaneous loads (g/hr) are shown in Table 111. These are the average of the loads measured on a sample per sample basis. This suggests that the Mohawk River is the greatest metals source. Of the treated effluents, Newtown Creek is the dominant source. Three CSOs, Bowery Bay, Jamaica, and Newtown Creek, may also be important local sources of total mercury and total lead.

Table 106. Instantaneous metal loads in g/hr.

| Site | Ag, D | Ag, T | Cd, D | Cd, T | Hg, D | Hg, DM | Hg, T | Pb D | Pb, T |
|------------------------------------|-------|-------|--------|-------|--------|----------|--------|-------|--------|
| CSO; Bowery Bay High Side | 2.9 | 60 | 0.18 | 12 | 0.018 | | 6.0 | 4.9 | 700 |
| CSO; Bowery Bay Low Side | 0.21 | 0.50 | 0.19 | 2.4 | 0.023 | 0.00023 | 1.4 | 5.3 | 160 |
| CSO; Coney Island Influent | | 0.33 | | 0.85 | | | 0.63 | | 62 |
| CSO; Hunts Point Influent | 0.21 | 7.2 | 0.050 | 2.5 | 0.021 | | 1.7 | 3.0 | 240 |
| CSO; Jamaica Influent | 3.8 | 15 | 0.25 | 2.5 | 0.099 | 0.0072 | 2.0 | 9.6 | 150 |
| CSO; Manhattan Grit Chamber | 0.092 | 1.4 | 0.0062 | 0.59 | 0.0058 | 0.0011 | 0.29 | 0.97 | 69 |
| CSO; Manhattan Pump Station | 1.7 | 5.5 | 0.29 | 1.4 | 0.025 | 0.00071 | 1.5 | 4.1 | 7.0 |
| CSO; Newtown Creek Influent | 0.24 | 3.1 | 0.32 | 4.8 | 0.022 | 0.00036 | 1.3 | 8.6 | 320 |
| CSO; North River Influent | 1.3 | 16 | 0.055 | 1.1 | 0.0077 | | 1.2 | 2.0 | 240 |
| CSO; Owls Head Influent | | 0.50 | | 0.96 | | | 1.4 | | 84 |
| CSO; Port Richmond Influent | 0.010 | 0.037 | 0.0071 | 0.044 | 0.0019 | | 0.024 | 0.45 | 4.3 |
| CSO; Red Hook Influent | 0.062 | 2.2 | 0.0064 | 0.64 | 0.0090 | | 0.43 | 0.50 | 51 |
| Industrial effluent; FK Plant Eff. | 0.012 | 0.013 | 0.014 | 0.016 | 0.0021 | 0.000015 | 0.0029 | 0.058 | 0.15 |
| Major tributary; Hudson R. | | | 24 | 300 | 3.3 | 0.11 | 38 | 270 | |
| Major tributary; Mohawk R. | | | 53 | 930 | 7.4 | 0.12 | 112 | 660 | 17,000 |
| Major tributary; Wallkill | 8.8 | 101 | 5.5 | 140 | 4.8 | 0.035 | 44 | 76 | 4,000 |
| Minor tributary; Bronx River | | | 0.42 | 0.43 | 0.032 | 0.000018 | 0.065 | 0.12 | |
| Minor tributary; Saw Mill River | | | 0.069 | 0.082 | 0.014 | 0.00010 | 0.006 | | |
| WPCF; 26th Ward | 0.74 | 4.3 | 0.47 | 0.54 | 0.061 | 0.00042 | 0.22 | 11 | 39 |
| WPCF; Bowery Bay | | | 0.99 | 1.3 | 0.058 | 0.0017 | 0.27 | | |
| WPCF; Coney Island | | | 0.28 | 0.34 | 0.056 | 0.0013 | 0.15 | 7.8 | |
| WPCF; Hunts Point | 1.9 | 5.7 | 1.1 | 1.4 | 0.055 | 0.00075 | 0.20 | 8.2 | 18 |
| WPCF; Jamaica | 2.0 | 9.6 | 0.80 | 2.6 | 0.036 | 0.0016 | 0.57 | 6.2 | 24 |
| WPCF; Newtown Creek | 41 | 150 | 13 | 19 | 0.31 | 0.021 | 1.7 | 58 | 120 |
| WPCF; North River | | | 3.5 | 3.5 | 0.095 | 0.0023 | 0.36 | 31 | 38 |
| WPCF; Oakwood Beach | | | 0.17 | 0.22 | 0.011 | 0.00018 | 0.013 | 1.5 | |
| WPCF; Owls Head | | | 0.54 | 0.76 | 0.19 | 0.0014 | 0.37 | 12 | |
| WPCF; Port Richmond | 0.73 | 2.5 | 0.66 | 0.69 | 0.025 | 0.00050 | 0.083 | 6.2 | 9.7 |
| WPCF; Poughkeepsie City | 0.12 | | 0.079 | 0.12 | 0.004 | 0.00021 | 0.033 | 0.77 | 1.2 |
| WPCF; Red Hook | | | 0.52 | 0.52 | 0.014 | 0.00089 | 0.053 | | |
| WPCF; Rensselaer | | | 0.11 | 0.17 | 0.014 | 0.00022 | 0.043 | | |
| WPCF; Rockaway | | | 0.095 | 0.17 | 0.011 | 0.00018 | 0.048 | 0.97 | |
| WPCF; Rockland County | | | 0.082 | 0.11 | 0.046 | 0.0015 | 0.19 | 1.1 | |
| WPCF; Tallman Island | | | 0.33 | 0.35 | 0.014 | | 0.064 | 1.7 | |
| WPCF; Wards Island | | | 1.2 | 1.2 | 0.076 | 0.0007 | 0.25 | 15 | |
| WPCF; Yonkers | | | 0.72 | 1.1 | 0.067 | 0.0013 | 0.90 | 16 | |

Table 112 shows results of biosolids that were composited over the month of February, 2001. Not all of the 14 NYCDEP treatment plants treat sludges and those which do, may treat material from different facilities. Material from a particular facility often includes sludge from other plants

Table 107. Dewatered sludges, ng/g.

| | Ag, T | Cd, T | Hg, T | Pb, T |
|-----------------------------|---------|--------|-------|---------|
| SLUDGE: 26th Ward | 34,000 | 4,900 | 2,600 | 250,000 |
| SLUDGE: Bowery Bay | 120,000 | 11,000 | 2,200 | 310,000 |
| SLUDGE: Coney Island | 53,000 | 4,200 | 2,800 | 240,000 |
| SLUDGE: Hunts Point #10 | 92,000 | 10,000 | 4,100 | 110,000 |
| SLUDGE: Hunts Point #9 | 57,000 | 7,600 | 2,000 | 340,000 |
| SLUDGE: Jamaica | 35,000 | 4,800 | 2,300 | 190,000 |
| SLUDGE: Oakwood Beach | 62,000 | 1,900 | 1,600 | 110,000 |
| SLUDGE: Port Richmond | 50,000 | 3,500 | 1,500 | 230,000 |
| SLUDGE: Red Hook | 89,000 | 6,500 | 2,300 | 350,000 |
| SLUDGE: Tallman Island | 88,000 | 6,400 | 2,600 | 200,000 |
| SLUDGE: Wards Island, North | 61,000 | 3,600 | 1,700 | 220,000 |
| SLUDGE: Wards Island, South | 180,000 | 5,700 | 2,600 | 340,000 |

Trackdown

Trackdown investigations attempted to discover the mercury source at the Rockland County WPCF. Samples were taken on March 8, 2000 and July 10, 2002 from sewers at sites selected by county personnel as capturing the major areas of the catchment (Figure 81). These failed to find clear evidence of a single source. A similar effort in the Newtown Creek area (February 15, 2000 and January 18, 2001) may have found pipes worth exploring but follow-up sampling has not been done.

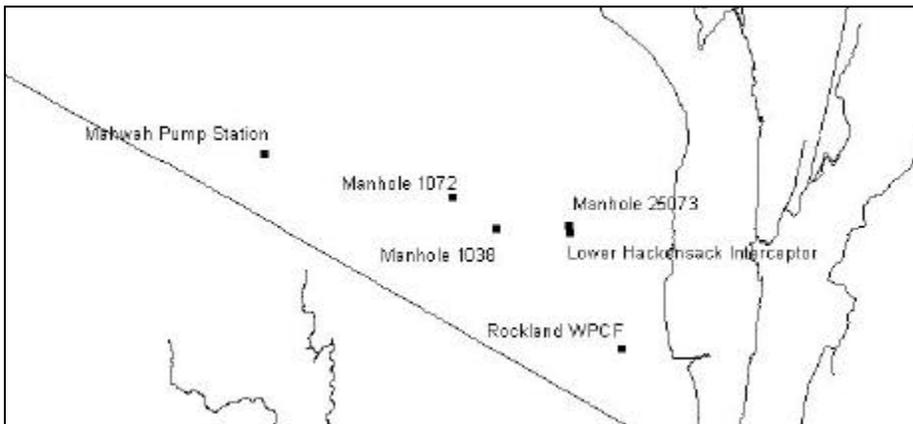


Figure 81. Locations of mercury trackdown sites in Rockland County.

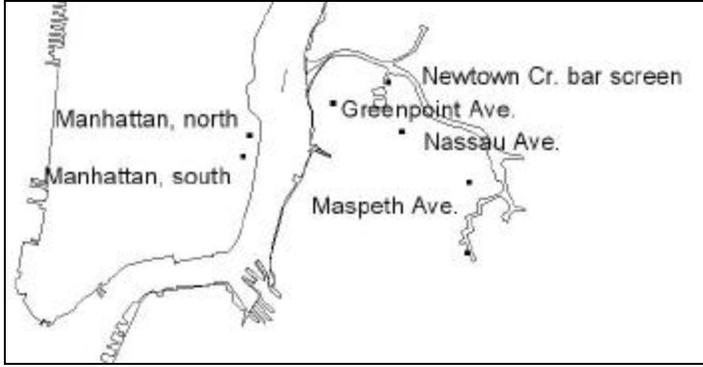


Figure 82. Mercury trackdown sites in the Newtown Creek area.

Table 108. Mercury trackdown at Newtown Creek and Rockland County.

| Newtown Creek WPCF | Hg, dissolved Hg, total | |
|---|-------------------------|------|
| Greenpoint Ave.: 1/18/2001 | 16.1 | |
| Greenpoint Ave.: 2/15/2000 | | 69.2 |
| Johnson Ave.: 1/18/2001 | 25.5 | |
| Johnson Ave.: 2/15/2000 | | 112 |
| Manhattan, north: 1/18/2001 | 14.7 | |
| Manhattan, north: 2/15/2000 | | 192 |
| Manhattan, south: 1/18/2001 | 31.3 | |
| Manhattan, south: 2/15/2000 | | 66.7 |
| Maspeth Ave. & Verick St.: 1/18/2001 | 11.2 | |
| Maspeth Ave. & Verick St.: 2/15/2000 | | 236 |
| Nassau Ave.: 1/18/2001 | 37.8 | |
| Newtown Cr. WPCF, bar screen: 1/18/2001 | 31 | |
| Rockland County | | |
| Lower Hackensack Interceptor: 7/10/2002 | | 456 |
| Mahwah Pump Station: 3/8/2000 | 2.34 | |
| Mahwah Pump Station: 7/10/2002 | | 395 |
| Manhole 1038: 3/8/2000 | 3.46 | |
| Manhole 1038: 7/10/2002 | | 245 |
| Manhole 1072: 3/8/2000 | 4.29 | |
| Manhole 1072: 7/10/2002 | | 223 |
| Manhole 25073: 3/8/2000 | 1.24 | |
| Manhole 25073: 7/10/2002 | | 493 |
| Rockland WPCF, after screen: 7/10/2002 | | 308 |
| Rockland WPCF, bar screen: 3/8/2000 | 3.58 | |

Accessory Parameters

Particulate Organic Carbon (POC)

Hydrophobic contaminants are preferentially transported on particles, particularly the organic fraction. Particulate organic carbon (POC) is a measure of the organic content of filterable particles.

POC information is important in modeling transport of hydrophobic contaminants. To help in evaluating the data, duplicate samples (more than one analysis from a single sample) and replicates (more than one sample from a cruise or visit to a point source) were analyzed. The results were evaluated as relative percent differences (RPDs) where the range (maximum minus minimum) was divided by the average. Table 114 shows averages, counts, and standard deviations of RPDs groups by time (duplicates), events (replicates) and sites. Counts are the number of RPDs, not the number of samples involved.

Table 109. RPDs (as percents) comparing duplicates, replicates, and multiple samples at a site

| | Average | Count | StDev |
|-----------------------------|---------|-------|-------|
| Time (duplicates) | 39 | 36 | 37 |
| Sampling event (replicates) | 57 | 31 | 61 |
| Sites | 140 | 45 | 83 |

This analysis shows that multiple analyses taken from the same sample are more similar than samples taken from a location at different times which are, in turn, more similar than samples taken from the same site on different days. This table also illustrates the difficulties in measuring POC.

Table 115 shows the general trends in POC concentrations of the various sample types investigated by CARP.

Table 110. POC concentrations by sample type.

| Sample type | Average | Count | StDev |
|------------------|---------|-------|-------|
| CSO/SWO | 24 | 4 | 31 |
| WPCF | 3.4 | 51 | 5.8 |
| Major Tribs | 3.3 | 37 | 3.7 |
| AMB-Kills | 0.95 | 28 | 0.76 |
| AMB-Non_Kills | 0.82 | 23 | 0.5 |
| Minor Tribs | 0.72 | 13 | 0.6 |
| AMB-Hudson | 0.56 | 16 | 0.27 |
| Treated Leachate | 0.3 | 4 | 0.11 |
| AMB-Clean | 0.13 | 8 | 0.049 |

The averages range over two orders of magnitude between the Long Island Sound/New York Bight samples and CSOs/SWOs. Generally, the range of POC concentrations across sites is much smaller than those for chemicals. Some of this may be due to very different types of organic carbon. For example, the organic carbon in the Bight may be largely due to plankton, where as that in the rivers may be largely due to suspended sediments. This relationship will be explored when we compare POC with suspended sediment.

Some of the sample types show relatively high variabilities. The highest variability is from CSOs and SWOs. Major tributaries, ambient samples, and POTW final effluents also have standard deviations greater than the mean. Tables 116, 117, and 118 show, respectively, more detail for these particular sample types.

Table 111. POC in Major Tributaries

| | Average | Count | StDev |
|--------------------------|---------|-------|-------|
| Wallkill (New Paltz) | 4.6 | 12 | 5.3 |
| Mohawk R. (Cohoes) | 3 | 12 | 2.9 |
| Hudson R. (Pleasantdale) | 2.1 | 11 | 1.6 |

Table 112. POC in Ambient Waters

| | Average POC | Count | StDev |
|------------------------------|-------------|-------|-------|
| Hudson R. at Poughkeepsie | 2.5 | 8 | 0.75 |
| Passaic River, Mid-Tidal | 1.5 | 4 | 1.1 |
| Hackensack R., Mid-Tidal | 1.5 | 4 | 1 |
| Lower Bay | 1.1 | 4 | 0.74 |
| Jamaica Bay | 1 | 4 | 0.31 |
| Passaic R., Mouth, Bottom | 1 | 4 | 0.59 |
| Raritan Bay | 0.94 | 4 | 0.62 |
| Hackensack R., Mouth | 0.89 | 4 | 0.8 |
| Lower East R. | 0.86 | 4 | 0.41 |
| Northern Arthur Kill | 0.74 | 4 | 0.17 |
| Hudson R. S. of Tappen Zee | 0.74 | 4 | 0.14 |
| Passaic R., Mouth, Surface | 0.58 | 3 | 0.81 |
| Hudson R. South of Harlem R. | 0.5 | 5 | 0.27 |
| Hudson R. below Kingston | 0.49 | 3 | 0.22 |
| Haverstraw Bay | 0.49 | 4 | 0.4 |
| Upper East R. | 0.49 | 4 | 0.36 |
| Newark Bay | 0.47 | 5 | 0.18 |
| Upper Bay | 0.41 | 3 | 0.28 |
| Long Island Sound | 0.17 | 3 | 0.051 |
| New York Bight | 0.099 | 5 | 0.021 |

Table 113. POC in POTW final effluents.

| | Average | Count | StDev |
|------------------|---------|-------|-------|
| Poughkeepsie (C) | 16 | 3 | 20 |
| Newtown Creek | 8.7 | 3 | 5.4 |
| Rockland County | 4.4 | 3 | 4 |
| Yonkers | 4.3 | 3 | 5 |
| Port Richmond | 4 | 3 | 2.6 |
| Jamaica | 3 | 3 | 3.2 |
| Owls Head | 2.7 | 3 | 1.9 |
| Rensselaer | 2.3 | 3 | 1.8 |
| Red Hook | 2.2 | 2 | 0.36 |
| Bowery Bay | 1.9 | 3 | 2.8 |
| 26th Ward | 1.8 | 3 | 1 |
| Coney Island | 1.6 | 3 | 1.2 |
| Hunts Point | 1.5 | 3 | 1.7 |
| Tallman Island | 1.5 | 3 | 0.75 |
| Oakwood Beach | 1.1 | 2 | 0.16 |
| North River | 1.1 | 3 | 0.65 |
| Wards Island | 0.85 | 3 | 0.74 |
| Rockaway | 0.39 | 2 | 0.08 |

Dissolved Organic Carbon (DOC)

DOC is filtered and acidified in the field. Relative to POC, there is much less handling and hence, there is less opportunity for contamination. Furthermore, the sample, filtered water, is more homogeneous than subsamples of a filter. This is reflected in the small relative percent differences from replicates (multiple samples taken during a cruise or visit).

Table 114. DOC concentrations by sample type.

| Sample type | Average | Count | StDev |
|---------------|---------|-------|-------|
| LANDF | 450 | 10 | 490 |
| CSO | 300 | 2 | 57 |
| INDEF | 63 | 3 | 98 |
| WPCF | 32 | 44 | 120 |
| Minor_TRIB | 7.4 | 9 | 3.3 |
| AMB-Kills | 7 | 27 | 3.2 |
| AMB-Hudson | 4.8 | 14 | 2 |
| Major_TRIB | 4.7 | 35 | 1.6 |
| AMB-clean | 4.5 | 9 | 5.5 |
| AMB-Non_Kills | 4.4 | 23 | 3.2 |

Table 115. DOC concentrations in tributaries

| | Average | Count | StDev |
|--------------------------|---------|-------|-------|
| Wallkill (New Paltz) | 6.2 | 12 | 1.9 |
| Hudson R. (Pleasantdale) | 3.9 | 11 | 0.59 |
| Mohawk R. (Cohoes) | 3.9 | 12 | 0.42 |

Table 116. DOC concentrations in ambient sites.

| | Average | Count | StDev |
|----------------------------|---------|-------|-------|
| Northern Arthur Kill | 10 | 3 | 7.4 |
| Hackensack R., Mid-Tidal | 9.7 | 4 | 2.9 |
| Upper East R. | 7.3 | 4 | 6.8 |
| Passaic R., Mouth, Bottom | 6.9 | 4 | 1.9 |
| Hackensack R., Mouth | 6.9 | 4 | 1.7 |
| Hudson R. S. of Tappen Zee | 6.8 | 5 | 3.5 |
| Passaic R., Mouth, Surface | 6.3 | 4 | 1.3 |
| Passaic River, Mid-Tidal | 5.9 | 3 | 1.1 |
| New York Bight | 5.6 | 5 | 7.6 |
| Upper Bay | 5.2 | 4 | 2.7 |
| Hudson R. below Kingston | 4.8 | 3 | 1.3 |
| Hudson R. S. of Harlem R. | 4.4 | 4 | 1.6 |
| Newark Bay | 4.3 | 5 | 1.1 |
| Haverstraw Bay | 4.3 | 3 | 0.62 |
| Jamaica Bay | 3.9 | 4 | 1.1 |
| Hudson R. at Poughkeepsie | 3.7 | 8 | 0.61 |
| Raritan Bay | 3.3 | 4 | 0.71 |
| Lower East R. | 3.3 | 3 | 0.53 |
| Long Island Sound | 3.2 | 4 | 0.94 |
| Lower Bay | 3.2 | 4 | 0.81 |

Table 117. DOC concentrations in final WPCF effluents.

| | Average | Count | StDev |
|------------------|---------|-------|-------|
| Red Hook | 420 | 2 | 580 |
| Newtown Creek | 24 | 3 | 3.9 |
| Rensselaer | 22 | 2 | 4.3 |
| Rockland County | 21 | 3 | 6.9 |
| Poughkeepsie (C) | 20 | 2 | 13 |
| Port Richmond | 17 | 3 | 3.2 |
| Yonkers | 13 | 3 | 5 |
| North River | 11 | 2 | 1.2 |
| Jamaica | 11 | 1 | |
| Oakwood Beach | 9.9 | 3 | 0.97 |
| Bowery Bay | 9.4 | 3 | 1.7 |
| Hunts Point | 9.4 | 2 | 0.43 |
| Owls Head | 9 | 1 | |
| 26th Ward | 8.8 | 3 | 1.7 |
| Tallman Island | 8.5 | 2 | 0.26 |
| Rockaway | 8.2 | 3 | 0.88 |
| Coney Island | 8.1 | 3 | 0.59 |
| Wards Island | 7 | 3 | 1.1 |

The unusually high concentration at Red Hook is due to a single sample taken on February 3, 1999 and having a reported value of 837 mg/L. One other DOC sample from Red Hook (April 14, 1999) had a reported concentration of 12 mg/L.

Dissolved organic carbon concentrations were greatest in landfill leachate, CSOs, and one POTW final effluent sample. The mean of the DOC field blanks was 1.6 (StDev=1.4).

Suspended Sediments (SS)

Suspended sediments were processed by USGS (major tributaries and Hudson River at Poughkeepsie) by weighing the entire contents of the sample bottles. The alternative approach, usually called “TSS” (total suspended solids) takes a well-stirred aliquot from the sample container. TSS is an appropriate parameter for samples that have no dense particles that might settle out between mixing and pouring. TSS is commonly used to evaluate WPCF effluent when sand grains, for example, are unlikely to be present. Suspended sediment is the appropriate parameter for surface waters. During CARP, filtration was performed in the field using either continuous pumping through a filter or grab samples that were poured through a vacuum filter. Distilled water was subsequently pumped through the filter to remove salts. In essence, all samples are “suspended sediment”.

The reproducibility of suspended sediments within a survey was assessed by replicates (several samples taken in a survey). The mean and median relative percent differences (RPDs) were 44% and 23% respectively.

Table 123. Area averages, suspended sediments.

| | Average | Count | StDev |
|---------------|---------|-------|-------|
| CSO | 220 | 6 | 110 |
| Major_TRIB | 95 | 35 | 130 |
| AMB-Hudson | 34 | 24 | 62 |
| AMB-Kills | 23 | 34 | 19 |
| INDEF | 17 | 4 | 13 |
| AMB-Non_Kills | 14 | 27 | 12 |
| WPCF | 12 | 61 | 14 |
| Minor_TRIB | 7.5 | 14 | 8 |
| AMB-clean | 4.8 | 10 | 2.5 |

Table 118. Suspended sediment averages from major tributaries.

| | Average | Count | StDev |
|--------------------------|---------|-------|-------|
| Mohawk R. (Cohoes) | 110 | 12 | 130 |
| Walkill (New Paltz) | 100 | 12 | 160 |
| Hudson R. (Pleasantdale) | 72 | 11 | 81 |

Table 119. Suspended sediments from ambient sites.

| | Average | Count | Std Dev |
|----------------------------|---------|-------|---------|
| Haverstraw Bay | 120 | 3 | 170 |
| Hudson R. at Poughkeepsie | 90 | 8 | 28 |
| Hackensack R., Mid-Tidal | 43 | 4 | 9.6 |
| Hudson R. S. of Tappan Zee | 34 | 3 | 17 |
| Passaic R., Mouth, Bottom | 33 | 3 | 18 |
| Lower East R. | 25 | 4 | 23 |
| Passaic River, Mid-Tidal | 24 | 4 | 17 |
| Hudson R. S. of Harlem R. | 18 | 6 | 15 |
| Jamaica Bay | 17 | 4 | 13 |
| Hudson R. below Kingston | 17 | 3 | 10 |
| Upper East R. | 14 | 3 | 9.8 |
| Lower Bay | 13 | 3 | 9.8 |
| Passaic R., Mouth, Surface | 13 | 4 | 9 |
| Northern Arthur Kill | 12 | 3 | 8.7 |
| Hackensack R., Mouth | 11 | 4 | 9.4 |
| Newark Bay | 11 | 4 | 9 |
| Upper Bay | 9.8 | 4 | 5.6 |
| Raritan Bay | 8.2 | 4 | 4.7 |
| Long Island Sound | 6.3 | 3 | 2.2 |
| New York Bight | 5.1 | 4 | 3.2 |

Table 120. Suspended sediments from WPCFs.

| | Average Count | StDev | |
|-----------------|---------------|-------|-----|
| Poughkeepsie | 37 | 3 | 42 |
| Newtown Creek | 29 | 4 | 13 |
| Hunts Point | 18 | 4 | 20 |
| Jamaica | 13 | 4 | 5 |
| Yonkers | 13 | 4 | 10 |
| Owls Head | 12 | 4 | 10 |
| Rockland County | 11 | 3 | 7 |
| Port Richmond | 9.5 | 3 | 6.6 |
| Rensselaer | 9.4 | 3 | 5.4 |
| Rockaway | 9.4 | 3 | 7.8 |
| Bowery Bay | 8.1 | 3 | 7.7 |
| 26th Ward | 7.8 | 4 | 3.7 |
| Red Hook | 7.4 | 2 | 0.5 |
| North River | 5.6 | 4 | 1.8 |
| Coney Island | 5.1 | 4 | 3.5 |
| Oakwood Beach | 4.2 | 3 | 2 |
| Tallman Island | 3.7 | 3 | 0.4 |
| Wards Island | 3.3 | 3 | 1.3 |

Loads

Average instantaneous loads of the three accessory parameters are given for the tributaries and the WPCFs, all in metric tons per hour (mT/hr). Of the major tributaries, the Mohawk appears to be the greatest source. Newtown Creek is the largest source among the WPCFs for POC and SS but Red Hook seems the largest WPCF source of DOC. The value should be considered in light of the very small sample size (two observations) and the disparity between the two observations.

Table 121. Loads (in metric tons/hour) of accessory parameters.

| Site | POC mT/hr | DOC mT/hr | SS mT/hr |
|---------------------------------|-----------|-----------|----------|
| Major tributary: Hudson R. | 6.3 | 7.9 | 228 |
| Major tributary: Mohawk R. | 8.8 | 8.0 | 339 |
| Major tributary: Wallkill | 4.7 | 2.1 | 124 |
| Minor tributary: Bronx River | 0.0005 | 0.055 | 0.040 |
| Minor tributary: Saw Mill River | 0.0038 | 0.081 | 0.079 |
| WPCF: 26th Ward | 0.012 | 0.090 | 0.073 |
| WPCF: Bowery Bay | 0.041 | 0.17 | 0.17 |
| WPCF: Coney Island | 0.017 | 0.12 | 0.062 |
| WPCF: Hunts Point | 0.029 | 0.21 | 0.75 |
| WPCF: Jamaica | 0.088 | 0.14 | 0.19 |
| WPCF: Newtown Creek | 0.39 | 1.0 | 1.1 |
| WPCF: North River | 0.035 | 0.27 | 0.11 |
| WPCF: Oakwood Beach | 0.0044 | 0.044 | 0.018 |
| WPCF: Owls Head | 0.034 | 0.16 | 0.12 |

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| | | | |
|-------------------------|--------|-------|-------|
| WPCF: Port Richmond | 0.030 | 0.14 | 0.07 |
| WPCF: Poughkeepsie City | 0.022 | 0.018 | 0.049 |
| WPCF: Red Hook | 0.012 | 2.2 | 0.040 |
| WPCF: Rensselaer | 0.0071 | 0.068 | 0.032 |
| WPCF: Rockaway | 0.0013 | 0.027 | 0.036 |
| WPCF: Rockland County | 0.010 | 0.060 | 0.027 |
| WPCF: Tallman Island | 0.0079 | 0.061 | 0.027 |
| WPCF: Wards Island | 0.028 | 0.21 | 0.10 |
| WPCF: Yonkers | 0.047 | 0.17 | 0.17 |

CONCLUSIONS AND RECOMMENDATIONS

The compilation shown here only begins to display the CARP data. Loading calculations and mass transport still require a great deal of further effort. Furthermore, more work will go into integrating results from the biota and sediment programs with the water section. Much of the focus of the present discussion has been on sources. The major tributaries seem to be the dominant loading sources for all of the analytes but some congener evidence suggests, for PCBs and dioxins, significant unidentified sources. The efficiency of transport of toxic substances through 150 miles of estuary still needs research. This has begun.

For the most part, areas of the harbor thought to be cleanest are freest of the target substances and those areas thought to be most contaminated, are. Perhaps the weakest element in sampling was the CSOs. These should have been sampled more intensively but the logistics of CSO sampling, even wet-weather influents, are formidable. Tributary sampling can also be improved. The Wallkill New Paltz site may have been too far from the river's mouth. The timing of the upper Hudson samples may have been somewhat inappropriate in that CARP concentrated on high flow periods to the neglect of times when bioperturbation may have resulted in significant loads. Too much effort may have gone into sampling wastewater treatment plants. They do not appear to be particularly significant loading sources.

One of the most unexpected discoveries was the prevalence of a hitherto unknown PCB congener, 3,3'-dichlorobiphenyl. Trackdown work has both confirmed the source and opened questions (the source of and disappearance of IUPAC 77) that have yet to be resolved. While very important in a few sewage treatment plant outfalls, 3,3'-DiCB is probably not a significant component of sediments or of biota. CARP has confirmed significant loadings from the upper Hudson River. However, downstream PCB loads are apparent in the changing congener distributions between upstream and downstream areas. A clear PCB source is seen in the 26th Ward WPCF in Brooklyn. However, the impact of this source on the receiving body, Jamaica Bay, is not apparent. There is also not a clear route to the remediation of the 26th Ward Aroclor 1260. Initial trackdown work at the Newtown Creek WPCF suggests the desirability of tracking PCBs back on two sewer mains. This has not yet been done.

Dioxin/furan "fingerprints" have turned out to be stable and perhaps diagnostic. Reduction in harbor dioxin concentrations will require gaining a much better understanding of the sources of some of the non-2,3,7,8-TCDD congeners. There may be a significant source of 2,3,4,7,8-PeCDF in the lower Hudson. Some other dioxins might turn out to be characteristic of urban wastewater. Related investigations from the World Trade Center disaster suggest that other dioxin-like substances should be addressed. These other substances include the co-planar PCBs and the brominated dioxins/furans. Better analytical methods and a great deal of toxicological work are required before routine monitoring of some of these dioxin-like materials becomes practical.

An early discovery of CARP was the significance of the Wallkill as a source of DDTs and dieldrin. Control of this source will require attention of sediment transport and sediment loading. Mechanisms for doing this have yet to be revealed. The actual impact of the Wallkill pesticides on the Hudson River and on the Harbor may be confused by the location of the sampling point, upstream of a deposition area (Sturgeon Pool).

Among the weakest parts of CARP was PAH sampling. The problems of addressing at least two very different models of PAH impact were not sufficiently appreciated at the beginning and were not corrected during CARP. There also needs to be a better toxicological understanding of the impact of PAHs on benthic test organisms. Further sampling might be postponed until the toxicology is better understood and the sampling problems are corrected. We lack a good way to field-concentrate dissolved PAHs. CARP has shown that some of the methylated PAHs are very abundant.

Metal loading appears to be largely driven by the major tributaries, but there are exceptions. Mercury trackdown at Rockland WPCF was unsuccessful, but that work should be completed. Metals sources to the Newtown Creek WPCF appear numerous but success in identifying and reducing sources will be challenging. Some of the CSOs may also be significant metals sources but the limited sampling calls the ultimate value of the loadings into question. There should be more work done on metals from CSOs and SWOs. Portable low-level mercury analyzers may play a useful role in describing mercury sources and in ultimate remediation.

Perhaps the most fruitful area for follow up will be the accessory parameters, particularly suspended sediment and POC. These very inexpensive parameters may yield important information on the behavior of the overall estuarine/harbor system. Sampling for the accessory parameters was weak as the numerous holes in the database attest. Effort is underway to begin developing suspended sediment data from the Hudson (ny.usgs.gov/projects/poused/index.html), Schoharie Creek, and the Wallkill. This should lead to better information about loadings of sediment to the basin which, in turn, may help in reducing the sediments.

CARP has been very successful in the development of field and analytical methods, particularly for PCBs, chlorinated dioxins/furans, and some of the chlorinated pesticides. Substances often reported as non-detected or simply ignored because of detection limits were routinely quantitated. The measurement of trace organic chemicals in the open ocean is perhaps unique. CARP has led to the development of powerful data management systems without which the volume of data would have utterly swamped the investigators.

This report is the conclusion of the first phase of CARP. Subsequent work could include addressing areas of weakness, follow-up on chemical source identification, and the design and implementation of a cost-effective long term monitoring program that would document effective clean-up and timely identification of emerging pollutants.

ABBREVIATIONS

Table 122. Abbreviations and acronyms.

| abbr. | full name |
|--------|---|
| AAS | Axys Analytical Services, a contract laboratory |
| Ag | silver |
| AMB | abbreviation for "ambient" |
| BAF | bioaccumulation factor |
| CARP | Contaminant Assessment Reduction Project |
| CCMP | Comprehensive Conservation Management Plan |
| Cd | cadmium |
| CFS | cubic feet per second |
| CSO | combined sewer overflow |
| DOC | dissolved organic carbon |
| DU | duplicate |
| EB | equipment blank |
| Eff. | effluent |
| EPA | US Environmental Protection Agency |
| FB | field blank |
| FGS | Frontier Geoscience, a contract lab |
| FK | Fresh Kills |
| HCB | hexachlorobenzene |
| HCH | hexachlorocyclohexane |
| HEP | Harbor Estuary Program |
| Hg | mercury |
| HMDC | Hackensack Meadowlands Development Commission |
| HpCDD | heptachlorodibenzo dioxin |
| HpCDF | heptachlorodibenzo furan |
| HRGC | high resolution gas chromatography |
| HRMS | high resolution mass spectrometry |
| HxCDD | hexachlorodibenzo dioxin |
| HxCDF | hexachlorodibenzo furan |
| Inf | influent |
| IUPAC | International Union of Applied and Physical Chemistry |
| mg | milligram |
| mg/L | milligram per liter, part per million |
| MGD | million gallons per day |
| ng | nanogram |
| ng/L | nanogram/liter, or part per trillion |
| NYCDEP | New York City Department of Environmental Protection |
| NYCDOS | New York City Department of Sanitation |
| NYSDEC | New York State Department of Environmental Conservation |
| OCDD | octachlorodibenzo dioxin |
| OCDF | octachlorodibenzo furan |
| PAH | polynuclear aromatic hydrocarbon |
| Pb | lead |

Table 129 continued.

| abbr. | full name |
|--------|---|
| PCB | polychlorinated biphenyl |
| PeCDD | pentachlorodibenzo dioxin |
| PeCDF | pentachlorodibenzo furan |
| pg | picogram |
| pg/L | picogram per liter (part per quadrillion) |
| PISCES | Passive In-Situ Chemical Extraction Sampler |
| POC | particulate organic carbon |
| PVSC | Passaic Valley Sewerage Commissioners |
| QTS | Quanterra, now called Severn Trent Laboratories, a contract lab |
| SA | sample |
| SDG | sample delivery group |
| SS | suspended sediment |
| SWO | storm water outfall |
| TCDD | tetrachlorodibenzo dioxin |
| TCDF | tetrachlorodibenzo furan |
| TDDT | total DDT |
| TEF | toxic equivalency factor |
| TEQ | toxic equivalence |
| TSS | total suspended solid |
| USACOE | US Army Corps of Engineers |
| USGS | US Geological Survey |
| WHO | World Health Organization |
| WPCF | water pollution control facility |
| WQS | water quality standard |
| WSU | Wright State University, acting as a contract lab |
| XAD | XAD, not an abbreviation. |